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# LONG-TERM ENVIRONMENTAL MONITORING PROGRAM

RESULTS AND INTERPRETATIONS FROM SAMPLING, 2008-2012



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Cover image - view from the mussel community at Disk Island, PWS, July 2013; logistics provided by Northwind Aviation, Homer. Photo by Bill Driskell.

**ABBREVIATIONS**

**Stations:**

AMT	Alyeska Marine Terminal, Port Valdez
AIB	Aialik Bay, west of Seward
COH	Constantine Harbor, Hinchinbrook Entrance, PWS
DII	Disk Island, Knight Island Group, western PWS
GOC	Gold Creek, Port Valdez
KNH	Knowles Head, eastern PWS
SHB	Sheep Bay, eastern PWS
SHH	Shuyak Harbor, Kodiak
SLB	Sleepy Bay, Latouche Island, western PWS
WIB	Windy Bay, Outer Kenai Peninsula
ZAB	Zaikof Bay, Montague Island, central PWS
ABL	NOAA/NMFS Auke Bay Laboratory, Juneau AK
AHC	aliphatic hydrocarbons (same as saturated hydrocarbons – SHC)
ANS	Alaskan North Slope
BWTF	Alyeska Terminal’s Ballast Water Treatment Facility
DSI	Dissolved Signal Index
DW	Dry Weight
EVOS	<i>Exxon Valdez</i> oil spill
EVTHD	<i>Exxon Valdez</i> Trustees Hydrocarbon Database
EMAP	US EPA Environmental Monitoring and Assessment Program
GC/FID	gas chromatography/flame ionization detector
GC/MS	gas chromatography/mass spectrometry
GERG	Geochemical and Environmental Research Group, Texas A&M University
KLI	Kinnetic Laboratories, Inc., Anchorage AK
MDL	analytic method detection limit
NIST	National Institute of Standards and Technology
NMFS	National Marine Fisheries Service
NOAA	National Oceanographic and Atmospheric Administration
PAH	polycyclic (or polynuclear) aromatic hydrocarbons
PECI	Payne Environmental Consultants, Inc., Encinitas, CA
PGS	particle grain size
PSI	Particulate Signal Index
PWS	Prince William Sound
RCAC	Regional Citizens’ Advisory Council
SAC	Scientific Advisory Committee for PWSRCAC
SCAT	Shoreline Cleanup Assessment Team
SHC	saturated hydrocarbons (same as AHC: n-alkanes + pristane and phytane)
SIM	selected ion monitoring
SRM	NIST standard reference material
TAHC	total AHC
TIC	total inorganic carbon
TOC	total organic carbon
TPAH	total PAH
TSHC	total saturated hydrocarbons (same as total alkanes)
UCM	unresolved complex mixture

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## 3 ABSTRACT

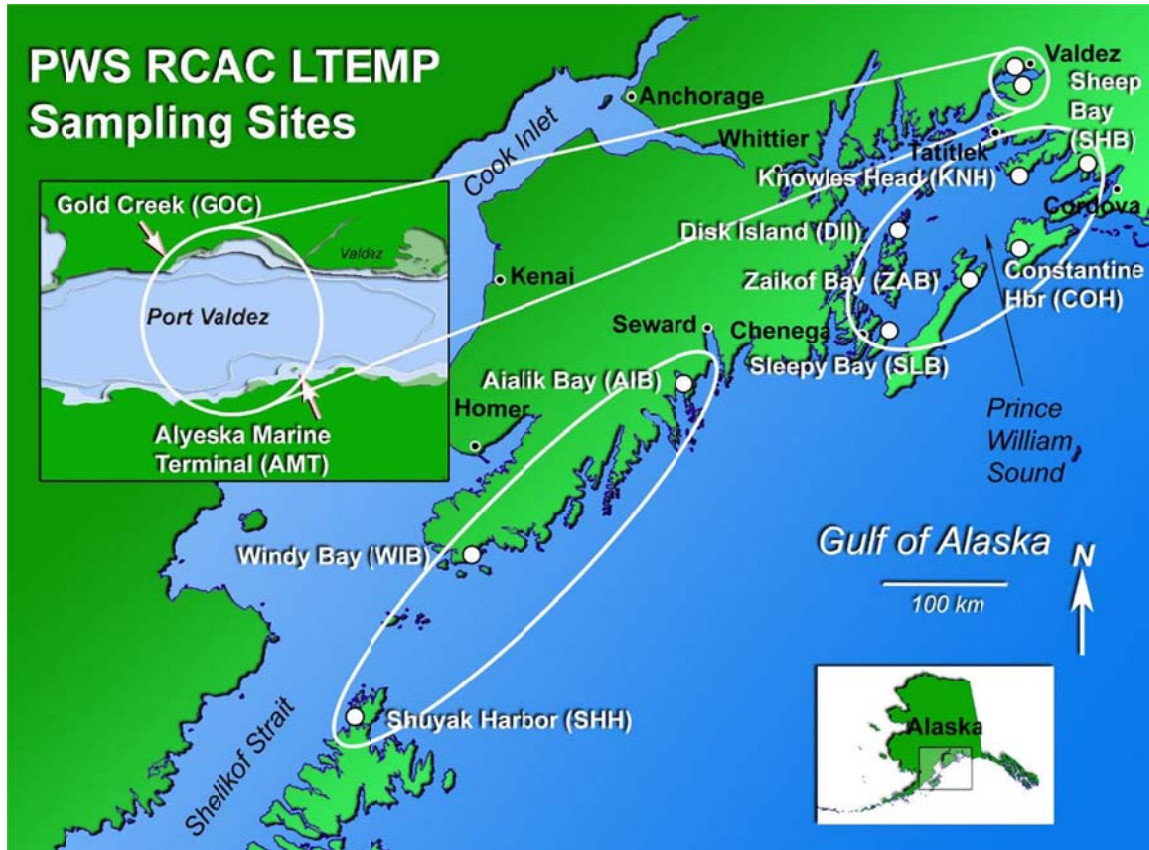
4 Petrogenic hydrocarbon inputs from the Alyeska Marine Terminal (AMT) and tanker operations are declining in  
5 mussels and sediments. The decrease likely reflects a combination of reduced ballast-water-treatment-facility  
6 (BTWF) discharge volumes from decreased North Slope oil production, the transition to double-hulled tankers with  
7 segregated ballast tanks, and improved BWTF efficiency at removing particulate/oil-phase polycyclic aromatic  
8 hydrocarbons (PAH) but also reflects an apparent regional decline evident in Knowles Head mussels. Composition  
9 is shifting away from petrogenic to pyrogenic in mussels and sediments; even the sediments at the terminal are  
10 beginning to look more like the background-reference, pyrogenic profiles at Gold Creek (GOC). Low levels of ANS  
11 oil are evident in Port sediments based on biomarkers; concentrations of these recalcitrant, source-specific  
12 indicators are seven times higher at AMT than GOC. Beyond Port Valdez, replicates at each of the seven remote  
13 sites show remarkable site fidelity in both PAH and saturated hydrocarbon (SHC) patterns, even at trace  
14 levels. Compared to 2004-05 West Coast Mussel Watch data and the more recent Alaskan Mussel Watch sites,  
15 LTEMP results for the Prince William Sound (PWS) and Gulf of Alaska (GOA) sites demonstrate that the region is  
16 exceptionally clean.

## 17 INTRODUCTION

18 In July 2008 and April 2009, all ten LTEMP stations (Figure 1) were visited in a sampling scheme designed, in part,  
19 to monitor for lingering impacts from the Exxon Valdez oil spill (EVOS). In subsequent years, the “grand round”  
20 sampling was reduced to only once every five years, and in Port Valdez and Knowles Head, the spring and fall  
21 collections were discontinued. Annual summer collections continued at the Port and nearby Knowles Head sites  
22 specifically monitoring terminal and tanker operations. This report summarizes findings for July 2008 through July  
23 2012, a period comprising the previous contract’s final year (2008-2009) for visiting all sites plus years 1-4 of the  
24 current 5-year monitoring cycle. As appropriate, the results are presented with the overall perspective/trend  
25 analysis from the inception of the program. Next year’s report will cover the recently completed July 2013 visit to  
26 all sites and summarize the program’s results and findings since 1993. Some of these data have been previously  
27 presented in a 2011 interim data report (Driskell and Payne 2011).

## 28 METHODS

29 Collection and analytical methods have been described in previous LTEMP reports (Payne et al., 2008a, 2010a).  
30 Briefly, three replicates of mussels are collected by hand at each site while triplicate sediment samples are  
31 collected from two locations within the Port using a modified Van Veen grab. Sampling protocols have remained  
32 constant, and the NMFS Auke Bay Laboratory has continued with the analytical chemistry measurements using  
33 their Standard Operating Procedures (SOPs) as detailed in our previous reports.



34

35 Figure 1 Map of the LTEMP sites. Circled regions represent sites with similar hydrocarbon signatures and events (discussed  
36 later).

37 New to this contract period is the inclusion of biomarkers for sediment samples along with the usual polycyclic  
38 aromatic hydrocarbons (PAH) and saturated hydrocarbon (SHC) data (Table 1). Petroleum biomarkers are unique  
39 for a given oil source and thus facilitate detection of AMT-derived, Alaska North Slope (ANS) crude-oil constituents  
40 in Port sediments. Since the inception of the program, low-level, combustion-derived PAH have been observed in  
41 the GOC sediments; oil-specific biomarkers would be much more useful at teasing out an ANS signal against the  
42 Port's glacial and pyrogenic inputs than the ubiquitous PAH. The biomarkers would not be as effective for the  
43 mussel tissues since unlike the accumulating sediments, mussels regularly purge and at most stations are currently  
44 only carrying trace-level, dissolved components.

45 Table 1 List of polycyclic aromatic hydrocarbon (PAH), saturated hydrocarbon (SHC), and biomarkers analytes in this study,  
46 along with analyte abbreviations used in figures throughout this report, and internal and surrogate standards.

ANALYTES	Abbreviation	Internal Standard	Surrogate Standard
<b>PAH</b>			
<b>Naphthalene</b>	N	A	1
<b>C1-Naphthalene</b>	N1	A	1
C2-Naphthalene	N2	A	2
C3-Naphthalene	N3	A	2
C4-Naphthalene	N4	A	2

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ANALYTES	Abbreviation	Internal Standard	Surrogate Standard
<b>Biphenyl</b>	BI	A	2
<b>Acenaphthylene</b>	AC	A	2
<b>Acenaphthene</b>	AE	A	2
<b>Fluorene</b>	F	A	2
C1-Fluorenes	F1	A	2
C2-Fluorenes	F2	A	2
C3-Fluorenes	F3	A	2
<b>Dibenzothiophene</b>	D	A	3
C1-Dibenzothiophene	D1	A	3
C2-Dibenzothiophene	D2	A	3
C3-Dibenzothiophene	D3	A	3
C4-Dibenzothiophene	D4	A	3
<b>Anthracene</b>	A	A	3
<b>Phenanthrene</b>	P	A	3
C1-Phenanthrene/Anthracene	PA1	A	3
C2-Phenanthrene/Anthracene	PA2	A	3
C3-Phenanthrene/Anthracene	PA3	A	3
C4-Phenanthrene/Anthracene	PA4	A	3
<b>Fluoranthene</b>	FL	A	3
<b>Pyrene</b>	PYR	A	3
C1-Fluoranthene/Pyrene	FP1	A	3
C2-Fluoranthene/Pyrene	FP2	A	3
C3-Fluoranthene/Pyrene	FP3	A	3
C4-Fluoranthene/Pyrene	FP4	A	3
<b>Benzo(a)Anthracene</b>	BA	A	4
Chrysene	C	A	4
C1-Chrysenes	C1	A	4
C2-Chrysenes	C2	A	4
C3-Chrysenes	C3	A	4
C4-Chrysenes	C4	A	4
<b>Benzo(b)fluoranthene</b>	BB	A	5
<b>Benzo(k)fluoranthene</b>	BK	A	5
<b>Benzo(e)pyrene</b>	BEP	A	5
<b>Benzo(a)pyrene</b>	BAP	A	5
<b>Perylene</b>	PER	A	6
<b>Indeno(1,2,3-cd)pyrene</b>	IP	A	5
<b>Dibenzo(a,h)anthracene</b>	DA	A	5
<b>Benzo(g,h,i)perylene</b>	BP	A	5
<b>Total PAH</b>	<b>TPAH</b>		<b>5</b>

**Saturated hydrocarbons (SHC or n-alkanes)**

n-Decane	C10	B	7
n-Undecane	C11	B	7
n-Dodecane	C12	B	7
n-Tridecane	C13	B	7



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ANALYTES	Abbreviation	Internal Standard	Surrogate Standard
n-Tetradecane	C14	B	8
n-Pentadecane	C15	B	8
n-Hexadecane	C16	B	8
n-Heptadecane	C17	B	8
Pristane	Pristane	B	8
n-Octadecane	C18	B	9
Phytane	Phytane	B	9
n-Nonadecane	C19	B	9
n-Eicosane	C20	B	9
n-Heneicosane	C21	B	9
n-Docosane	C22	B	10
n-Tricosane	C23	B	10
n-Tetracosane	C24	B	10
n-Pentacosane	C25	B	10
n-Hexacosane	C26	B	10
n-Heptacosane	C27	B	10
n-Octacosane	C28	B	10
n-Nonacosane	C29	B	11
n-Triacontane	C30	B	11
n-Hentriacontane	C31	B	11
n-Dotriacontane	C32	B	11
n-Tritriacontane	C33	B	11
n-Tetratriacontane	C34	B	11
<b>Total n-Alkanes</b>	<b>TALK</b>		
Calibrated analytes are identified by boldface. Internal standards: A = hexamethylbenzene; B = dodecylcyclohexane. Surrogate standards: 1 = naphthalene-d8, 2 = acenaphthene-d10, 3 = phenanthrene-d10, 4 = chrysene-d12, 5 = benzo[a]pyrene-d12, 6 = perylene-d12, 7 = dodecane-d26, 8 = hexadecane-d34, 9 = eicosane-d42, 10 = tetracosane-d50, and 11 = triacontane-d62.			

47 Petroleum Biomarkers

Class	Biomarker	Abbrev
Terpanes	C23 tricyclic terpane	T4
	C24 tricyclic terpane	T5
	C25 tricyclic terpane (a)	T6
	C25 tricyclic terpane (b)	T6
	C24 tetracyclic terpane	T6a
	C26 tricyclic terpane (a)	T6c
	C26 tricyclic terpane (b)	T6b
	C28 tricyclic terpane (a)	T8
	C28 tricyclic terpane (b)	T7
	C29 tricyclic terpane (a)	T10
	C29 tricyclic terpane (b)	T9

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Hopanes	18 $\alpha$ (H),21 $\beta$ (H)-22,29,30-trisnorhopane	Ts
	17 $\alpha$ (H),21 $\beta$ (H)-22,29,30-trisnorhopane	Tm
	17 $\alpha$ (H),18 $\alpha$ (H),21 $\beta$ (H)-28,30-bisnorhopane	14a
	17 $\alpha$ (H),21 $\beta$ (H)-25-norhopane	14b
	17 $\alpha$ (H),21 $\beta$ (H)-30-norhopane	T15
	18 $\alpha$ (H),21 $\beta$ (H)-30-norneohopane	T16
	17 $\beta$ (H),21 $\alpha$ (H)-30-norhopane (normoretane)	T17
	18 $\alpha$ (H) and 18 $\beta$ (H)-oleanane	T18
	17 $\alpha$ (H),21 $\beta$ (H)-hopane	T19
	17 $\alpha$ (H)-30-nor-29-homohopane	
	17 $\beta$ (H),21 $\alpha$ (H)-hopane (moretane)	T20
	22S-17 $\alpha$ (H),21 $\beta$ (H)-30-homohopane	T21
	22R-17 $\alpha$ (H),21 $\beta$ (H)-30-homohopane	T22
	Gammacerane	T22a
	22S-17 $\alpha$ (H),21 $\beta$ (H)-30,31-bishomohopane	T26
	22R-17 $\alpha$ (H),21 $\beta$ (H)-30,31-bishomohopane	T27
	22S-17 $\alpha$ (H),21 $\beta$ (H)-30,31,32-trishomohopane	T30
	22R-17 $\alpha$ (H),21 $\beta$ (H)-30,31,32-trishomohopane	T31
	22S-17 $\alpha$ (H),21 $\beta$ (H)-30,31,32,33-tetrakishomohopane	T32
	22R-17 $\alpha$ (H),21 $\beta$ (H)-30,31,32,33-tetrakishomohopane	T33
22S-17 $\alpha$ (H),21 $\beta$ (H)-30,31,32,33,34-pentakishomohopane	T34	
22R-17 $\alpha$ (H),21 $\beta$ (H)-30,31,32,33,34-pentakishomohopane	T35	
Steranes	C22 5 $\alpha$ (H),14 $\beta$ (H),17 $\alpha$ (H)-sterane	
	C27 20S-13 $\beta$ (H),17 $\alpha$ (H)-diasterane (diacholestane)	S4
	C27 20R-13 $\beta$ (H),17 $\alpha$ (H)-diasterane (diacholestane)	S5
	C27 20S-5 $\alpha$ (H),14 $\alpha$ (H),17 $\alpha$ (H)-cholestane	S12
	C27 20R-5 $\alpha$ (H),14 $\beta$ (H),17 $\beta$ (H)-cholestane	S14
	C27 20S-5 $\alpha$ (H),14 $\beta$ (H),17 $\beta$ (H)-cholestane	S15
	C27 20R-5 $\alpha$ (H),14 $\alpha$ (H),17 $\alpha$ (H)-cholestane	S17
	C28 20S-5 $\alpha$ (H),14 $\alpha$ (H),17 $\alpha$ (H)-ergostane (methylcholestane)	S20
	C28 20R-5 $\alpha$ (H),14 $\beta$ (H),17 $\beta$ (H)-ergostane (methylcholestane)	S22
	C28 20S-5 $\alpha$ (H),14 $\beta$ (H),17 $\beta$ (H)-ergostane (methylcholestane)	S23
	C28 20R-5 $\alpha$ (H),14 $\alpha$ (H),17 $\alpha$ (H)-ergostane (methylcholestane)	S24
	C29 20S-5 $\alpha$ (H),14 $\alpha$ (H),17 $\alpha$ (H)-stigmastane (ethylcholestane)	S25
	C29 20R-5 $\alpha$ (H),14 $\beta$ (H),17 $\beta$ (H)-stigmastane (ethylcholestane)	S26
	C29 20S-5 $\alpha$ (H),14 $\beta$ (H),17 $\beta$ (H)-stigmastane (ethylcholestane)	S27
	C29 20R-5 $\alpha$ (H),14 $\alpha$ (H),17 $\alpha$ (H)-stigmastane (ethylcholestane)	S28

49 **RESULTS AND DISCUSSION**

50 As in previous LTEMP reports, all analytes are reported on a ng/g dry weight (DW) basis, and corrected for  
 51 surrogate recoveries. Surrogates are novel or deuterated compounds added in known amounts to each sample in  
 52 order to assess the efficiency of extraction and analysis. In the Auke Bay Laboratory SOPs, surrogate recoveries are  
 53 considered acceptable if they are between 30% and 120%. If more than 10 recoveries from an entire string  
 54 (analytic batch) fall outside the recovery targets, the string fails QA criteria and is reprocessed. If the majority of  
 55 the failed recoveries occur in an individual sample, only that sample is reprocessed. Surrogate recovery standards  
 56 were met for 95.1% of all surrogate hydrocarbons analyzed during this reporting period. For the new biomarker  
 57 analytes, median surrogate recovery was 114.4% for 36 analyses. For different QA reasons, the 2009 sediment  
 58 batch was re-extracted and reanalyzed.

59 Laboratory method blanks for each analytic sample batch demonstrated no significant background interference  
 60 from analytical procedures, thus assuring that the analytes in the field samples represented environmental  
 61 constituents and not analytical artifacts. In addition, new method detection limits (MDLs), established in 2010,  
 62 dramatically improved the confidence in accurately quantifying low-level PAH. Dropping an order of magnitude  
 63 below previous MDLs, PAH limits are now down to a miniscule median 0.04 ng/g dry weight in sediments and 0.8  
 64 ng/g dry weight in tissues (Table 2). Likewise, MDLs for SHC dropped to a median 0.5 ng/g dry weight in sediments  
 65 and 21 ng/g dry weight in tissues.

66 Table 2 Historic and current method detection limits (MDL) for quantifying hydrocarbon analytes at Auke Bay Laboratory. All  
 67 values in ng/g dry weight (based on average 2008-2012 dry wt/wet wt measurements).

Analyte	Abbrev	Tissue			Sediment		
		1993	1996	2010	1993	1996	2010
naphthalene	N	24.4	19.1	1.9	0.98	2.48	0.06
2-methylnaphthalene		11.0	12.1	1.1	0.90	1.44	0.06
1-methylnaphthalene		14.6	31.4	1.0	1.98	1.08	0.06
2,6-dimethylnaphthalene		7.8	6.3	0.8	1.88	0.71	0.04
2,3,5-trimethylnaphthalene		4.7	7.6	0.7	0.90	0.37	0.02
biphenyl*	BI	8.3	26.6	20.7	4.13	0.90	0.19
acenaphthylene	AC	6.3	21.3	1.0	2.40	0.73	0.04
acenaphthene	AE	6.0	8.7	0.4	1.17	0.54	0.04
fluorene	F	5.6	19.7	1.3	1.75	1.31	0.04
dibenzothiophene	DBT	3.9	12.7	0.6	0.92	0.44	0.04
phenanthrene	P	9.2	12.3	0.8	1.29	0.71	0.02
1-methylphenanthrene		6.2	28.1	0.9	2.62	0.56	0.02
anthracene	A	7.2	9.9	0.9	3.19	0.50	0.02
fluoranthene	FL	13.1	23.1	0.8	0.88	0.44	0.02
pyrene	PYR	10.9	23.8	0.8	1.00	0.38	0.02
benzo(a)anthracene	BAA	4.8	8.2	0.3	0.33	0.31	0.04
chrysene	C	6.1	13.1	0.6	2.54	0.81	0.02
benzo(b)fluoranthene	BBF	6.7	37.6	0.6	1.17	0.46	0.08

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benzo(k)fluoranthene	BKF	8.7	18.6	0.9		0.77	0.35	0.08
Benzo(e)pyrene	BEP	9.1	22.1	0.7		1.15	0.60	0.06
Benzo(a)pyrene	BAP	7.2	17.8	1.1		3.10	0.44	0.02
Perylene	PER	10.4	25.0	5.1		1.12	0.98	0.02
Indeno(1,2,3-cd)pyrene	IND	4.9	11.8	0.4		0.65	0.38	0.06
dibenzo(a,h)anthracene	DBA	4.2	14.7	0.3		0.48	0.35	0.06
benzo(ghi)perylene	BGHI	6.1	41.0	0.6		0.90	0.77	0.02
						0.00	0.00	0.00
min		3.9	6.3	0.3		0.33	0.31	0.02
max		24.4	41.0	20.7		4.13	2.48	0.19
median		7.2	18.6	0.8		1.15	0.56	0.04
extracted sample mass (g wet weight)		10	8	10		20	40	40
n-nonane	C9			20				0.97
n-decane	C10	186	643	31		5.9	6.7	0.52
n-undecane	C11	109	523	39		4.3	7.9	0.90
n-dodecane	C12	71	305	39		6.4	7.7	2.10
n-tridecane	C13	143	211	91		5.5	13.3	1.69
n-tetradecane	C14	272	284	38		4.3	15.5	1.58
n-pentadecane	C15	251	214	21		19.4	16.6	1.06
n-hexadecane	C16	394	168	14		8.5	18.7	0.58
n-heptadecane	C17	526	166	16		10.9	17.5	0.27
pristane	Prist	283	159	32		20.2	15.5	0.38
n-octadecane	C18	204	99	7		23.3	27.3	0.44
phytane	Phyt	204	98	8		23.3	19.2	0.50
n-nonadecane	C19	240	76	6		5.3	6.9	0.23
n-eicosane	C20	274	56	7		10.4	8.1	0.12
n-heneicosane	C21	201	38	7		9.1	8.1	0.19
n-docosane	C22	75	60	13		6.0	8.3	0.21
n-tricosane	C23	149	52	9		21.3	8.1	0.62
n-tetracosane	C24	172	49	24		33.7	10.4	0.44
n-pentacosane	C25	101	25	15		7.3	8.6	0.27
n-hexacosane	C26	104	84	18		20.4	6.3	0.50
n-heptacosane	C27	196	78	32		16.4	5.1	0.31
n-octacosane	C28	179	137	30		13.5	16.9	0.44
n-nonacosane	C29	332	111	18		14.7	7.7	0.40
n-triacontane	C30	219	175	20		17.9	4.5	1.10
n-hentriacontane	C31	0	0	26		0.0	0.0	0.40

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n-dotriacontane	C32	153	394	21		23.3	11.6	0.42
n-tritriacontane	C33	0	0	83		0.0	0.0	0.98
n-tetratriacontane	C34	89	407	99		13.9	57.3	2.92
n-pentatriacontane	C35	0	0	127		0.0	0.0	5.15
n-hexatriacontane	C36	0	0	134		0.0	0.0	7.00
min		71	25	6		4.3	4.5	0.12
max		526	643	134		33.7	57.3	7.00
median		196	137	21		13.5	8.6	0.50
extracted sample mass (g wet weight)		8	8	10		20	40	40

68 \* Biphenyl is sometimes encountered as a laboratory artifact (as it was during the MDL study) thereby elevating the MDL for  
69 this constituent compared to the other analytes. It is not utilized to a great extent in the LTEMP data analysis.

## 70 PORT VALDEZ SEDIMENTS

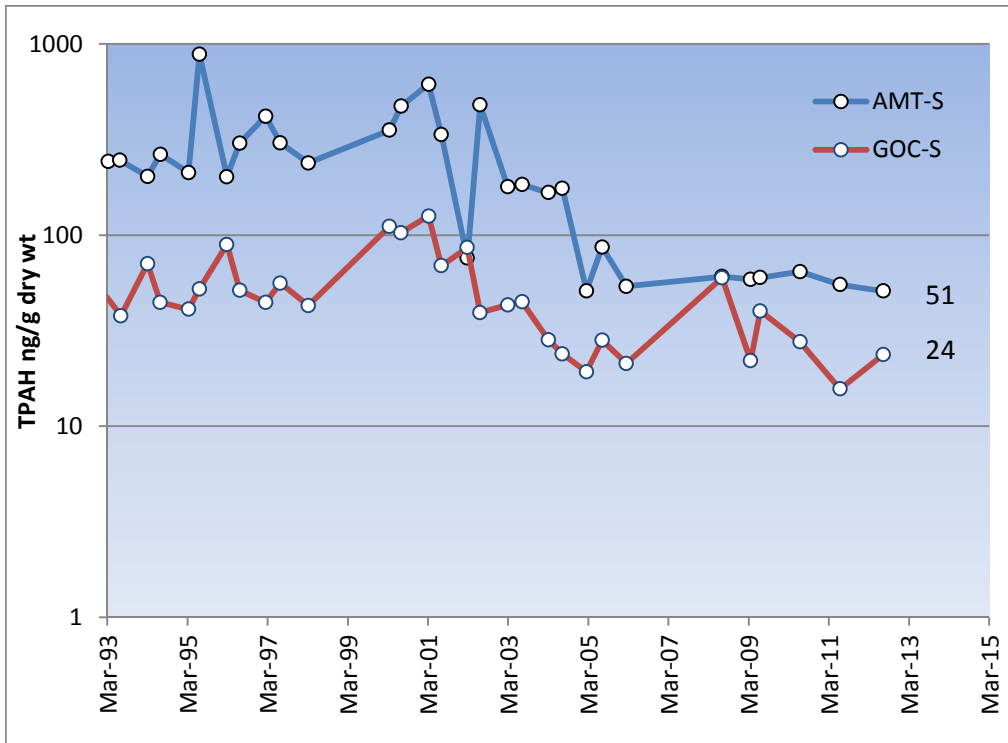
71 In sediments from the Berth 4 site at the Alyeska Marine Terminal (AMT), average TPAH levels are very low,  
72 plateaued around 50-60 ng/g DW since March 2005 (Figure 2), but there has been a shift in the patterns. PAH  
73 patterns (examples in Figure 3; all samples in Appendix 1) suggest that the higher-molecular-weight PAH became  
74 more pyrogenic (i.e. from combustion sources, less petrogenic) starting in July 2009. A pyrogenic pattern is  
75 recognized by the dominant parent PAH relative to the alkylated homologues (e.g., in the bottom of Figure 3, the  
76 Ph, Fl and C are the non-alkylated parents); in petrogenic signatures, the analyte groups form a hump pattern in  
77 which the parent PAH no longer dominates. In the SHC plots, by the next year, July 2010, petrogenic alkanes begin  
78 to diminish relative to odd-carbon-numbered biogenic n-alkanes (C<sub>25</sub>, C<sub>27</sub>, C<sub>29</sub>) derived from terrestrial plant waxes.  
79 By 2011, the intermediate-molecular-weight PAH (fluorenes and phenanthrenes/anthracenes) also became more  
80 pyrogenic in character with the SHC still dominated by biogenic constituents.

81 The Gold Creek reference site (GOC) sediments had lower TPAH concentrations than AMT throughout the duration  
82 of the program (Payne et al., 2008b) excepting two spill events (Figure 2), most recently with extremely low  
83 concentrations, 15-60 ng/g DW. GOC PAH profiles were dominated by pyrogenics (Payne et al., 2008a,b; 2010a)  
84 from April 2000 through this 2008-2012 reporting period (see Figure 4 and Appendix 1). In addition to the higher-  
85 molecular-weight pyrogenic PAH, Gold Creek sediments also contain a dominant and invariant suite of N<sub>0</sub>-N<sub>4</sub>  
86 naphthalenes that are believed to derive from glacial and riverine sediment input to the Port (Payne et al.  
87 2010a,b). SHC in the sediments are almost exclusively dominated by biogenic sources from marine phytoplankton  
88 and terrestrial plant waxes (i.e., n-C<sub>15</sub>, n-C<sub>17</sub>, pristane, and an odd-carbon-numbered suite of n-alkanes between  
89 n-C<sub>23</sub> and n-C<sub>33</sub>, respectively).

90 In comparing the AMT and GOC sediment profiles, it is interesting that as the petrogenic profiles at AMT are  
91 disappearing, by July 2012 its signatures (Figure 3 bottom) begin to resemble the Gold Creek reference sediments  
92 (Figure 4). Only traces of the higher-molecular-weight petroleum waxes (> n-C<sub>32</sub>) remain whereas they used to  
93 dominate the SHC profiles (Payne et al., 2008a,b).

94 In our 2010 LTEMP report (Payne et al., 2010a), we cautioned that at extremely low concentrations using our  
95 algorithmic PAH source-phase assignment model is problematic whereby a very minute change in a trace pattern  
96 could flip a phase assignment. For the 2008-2012 data where nearly all patterns at trace levels and near MDLs,  
97 we've abandoned that approach in favor of pattern descriptions and graphics. We also include the interpretations  
98 of Mark Carls (Appendix 3) in which he's applied a multivariate scoring procedure (Carls, 2006) to assign type and  
99 sources. His results generally corroborate what we're seeing in the histogram patterns.

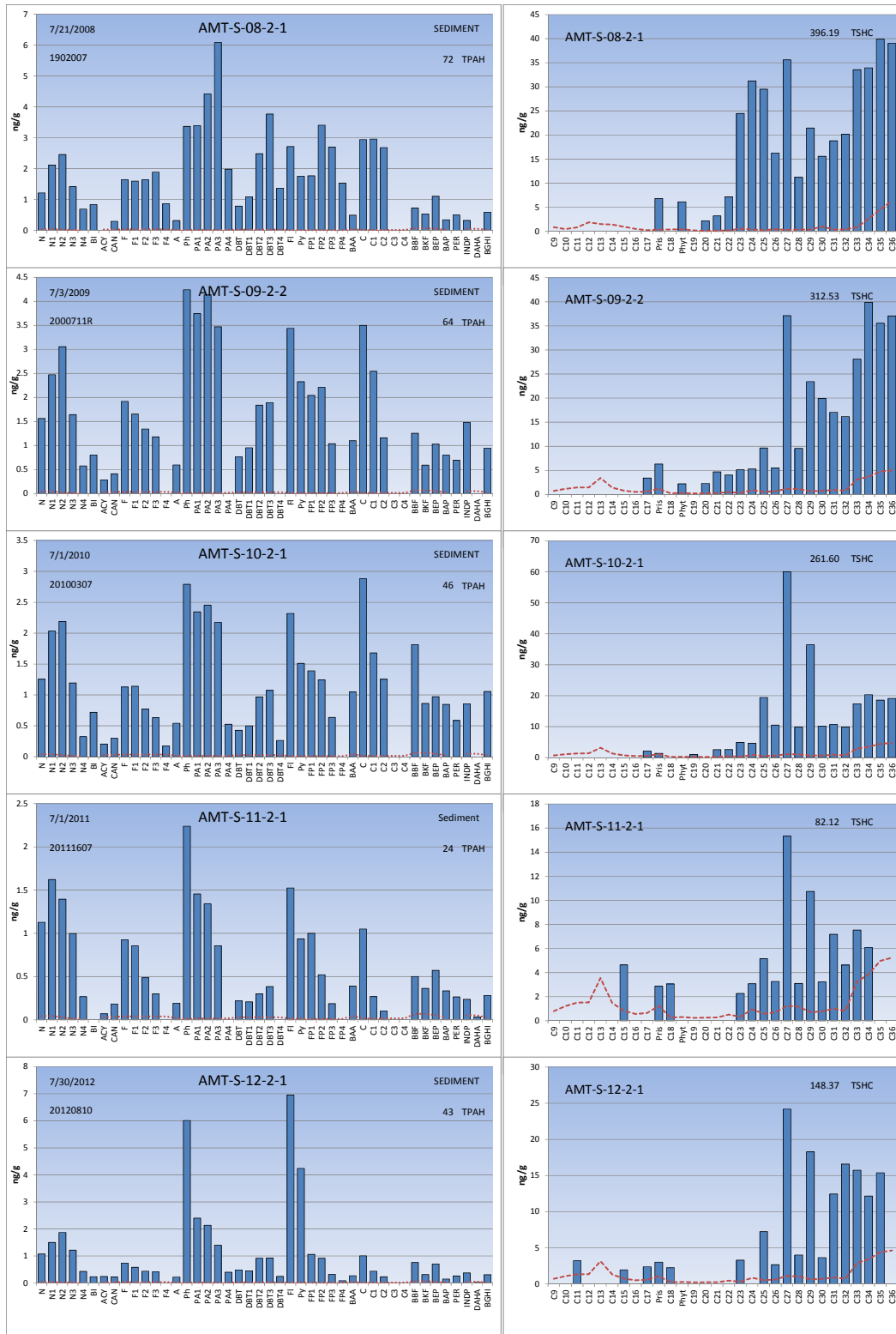
100 We also noted that without biomarker data, attributing the very minor petrogenic signal observed in Gold Creek  
101 sediments to Alyeska Marine Terminal discharges would be confounded. From their 2005 EMP biomarker data,  
102 Shaw et al. (2005) concluded that EMP Stations 40 and 50 (near Gold Creek but at greater depths in the Port)  
103 contained petrogenic components derived from Alaska North Slope crude oil. To confirm this finding that our data  
104 only hinted at, a suite of biomarkers was added to the sediment analyte list with the 2009-2014 LTEMP contract.



105

106 Figure 2 Time series of TPAH in sediments at Alyeska Terminal and Gold Creek.

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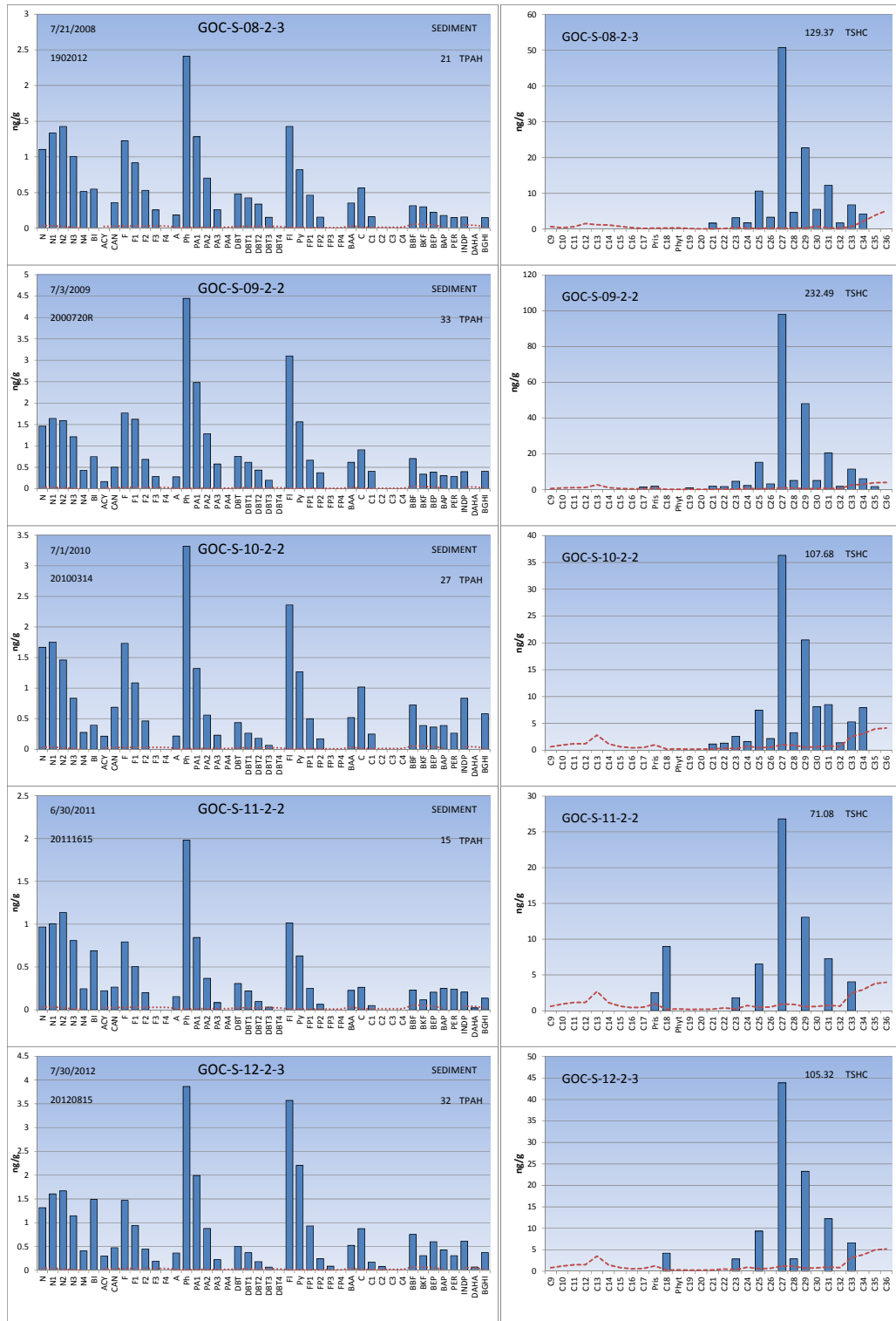


107

108 Figure 3 Example PAH and SHC signatures of sediments at Alyeska Terminal between July 2008 and July 2012 showing the  
 109 progression from a primarily petrogenic to pyrogenic PAH signature with increasing terrestrial biogenic SHC and decreasing  
 110 higher-molecular-weight n-alkane residuals. Red dashed line is sample-specific MDL. Sample ID code comprises station-matrix-  
 111 year-season-rep.



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112

113 Figure 4 Example PAH and SHC signatures of sediments at Gold Creek between July 2008 and July 2012 showing the  
 114 essentially invariant background naphthalene components and pyrogenic, parent-dominated, PAH. SHC patterns reflect  
 115 constant terrestrial and marine biogenic input. Red dashed line is sample-specific MDL.

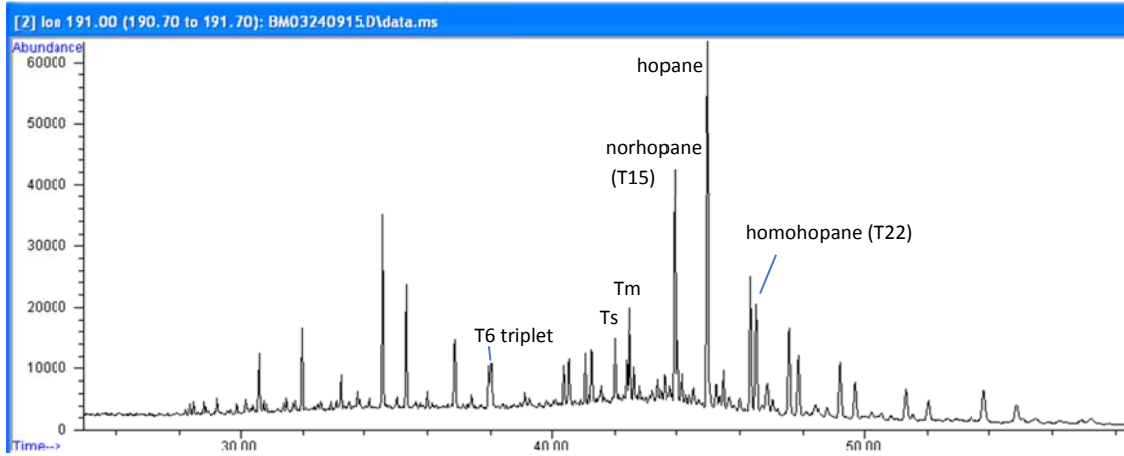
## 116 SEDIMENT BIOMARKERS

117 At our request, in 2011, the Auke Bay Laboratory (ABL) modified their analytical procedures to report petroleum  
118 biomarkers. The method is an extension to the normal GC/MS 8270 method used to acquire PAH data;  
119 conceptually, three new ions are added to the mass spectrometer's list for selective ion monitoring (SIM). As a lab  
120 calibration standard, ABL runs the original *Exxon Valdez* oil (from the ship's hold, an ANS crude oil), in addition to a  
121 NIST standard reference material (SRM). But there are no certified standard values for biomarkers in these oils; we  
122 can only compare ABL results to other lab results for validation. We've done this for both ANS and Deepwater  
123 Horizon oils (also analyzed at ABL and now a NIST standard). The results are favorable; ABL's ANS reference oil  
124 results compare quite well with published ANS results (Figure 5); however, the lab has issues separating the triplet  
125 T6 terpane compounds, T6a, T6b and T6c (identified as peaks 8, 9 & 10 (Figure 5 center). This is unfortunate since  
126 the triplets are often used as diagnostic markers. In ABL-quantified plots of the field samples (Figure 6), there also  
127 seem to be deficits in the mid-terpane analytes (left third of the plot) but with this small number of samples from  
128 only two locations (and only two analytic strings), it is difficult to know, at this stage, if these results reflect actual  
129 environmental conditions or analytic artifacts. It is also possible that the biomarker patterns for ANS crude oil  
130 have changed slightly since 1989 as oil from several new North Slope fields are blended into the pipeline mixture.  
131 More samples of ANS oil are being requested from Alyeska to assess the PAH and biomarker composition of the  
132 current blend. Presumably, this will also reflect the biomarker composition in oil residues in the treated ballast  
133 water currently being discharged into the Port.

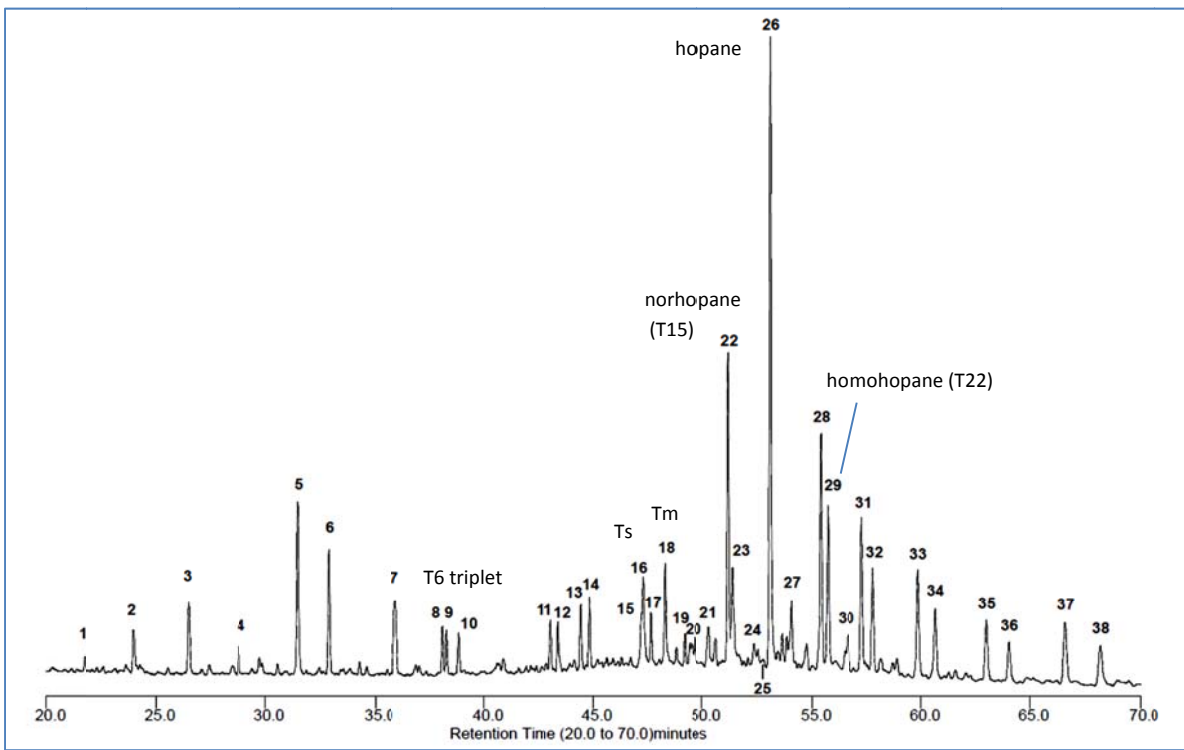
134 Multiple approaches have been suggested for interpreting biomarker data but some degree of expert-guided  
135 pattern matching must be employed. Most approaches involve various diagnostic ratios (Wang and Stout, 2007)  
136 with several ratios normalized on the highly conservative 17 $\alpha$ (H),21 $\beta$ (H)-hopane (also labeled T19 or C30 hopane).  
137 But depending on the local environs and despite the purported persistence of biomarkers, all ratios are not equally  
138 effective and must be individually evaluated for a given spill/habitat. With LTEMP data, we initially screened  
139 results graphically with an overlaid ANS reference (Figure 6). We then used the frequently reliable, Ts/Tm and  
140 norhopane (T15)/hopane plus a suggested 22R homohopane (T22)/hopane ratios to confirm the visual similarities.  
141 Normally, the T6 triplets ratio also would be added to this list but not if the lab is having analytic issues. The final  
142 suite of diagnostic ratios can then be the basis for a rigorous methodology advocated by European agencies  
143 (NordTest plots) or adapted for other multivariate approaches (various authors in Wang & Stout, 2007). Mark  
144 Carls is currently exploring a novel scoring method.

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145



146



1	C19 Tricyclic terpane	14	C29 [22R] Tricyclic terpane	27	17b21a Moretane [C30]
2	C20 Tricyclic terpane	15	C30 [22S] Tricyclic terpane	28	22S Homohopane [C31]
3	C21 Tricyclic terpane	16	18a Trisnorneohopane [C27 Ts]	29	22R Homohopane [C31]
4	C22 Tricyclic terpane	17	C30 [22R] Tricyclic terpane	30	Gammacerane [C30]
5	C23 Tricyclic terpane	18	17a Trisnorneohopane [C27 Tm]	31	22S Bishomohopane [C32]
6	C24 Tricyclic terpane	19	C31 [22S] Tricyclic terpane	32	22R Bishomohopane [C32]
7	C25 Tricyclic terpane	20	C31 [22R] Tricyclic terpane	33	22S Trishomohopane [C33]
8	C26 [22S] Tricyclic terpane	21	Bisnorhopane [C28]	34	22R Trishomohopane [C33]
9	C26 [22R] Tricyclic terpane	22	Norhopane [C29]	35	22S Tetrakishomohopane [C34]
10	C24 Tetracyclic terpane	23	18a Neonorhopane [C29]	36	22R Tetrakishomohopane [C34]
11	C28 [22S] Tricyclic terpane	24	17b21a Normoretane [C29]	37	22S Pentakishomohopane [C35]
12	C28 [22R] Tricyclic terpane	25	Oleanane [C30]	38	22R Pentakishomohopane [C35]
13	C29 [22S] Tricyclic terpane	26	Hopane [C30]		

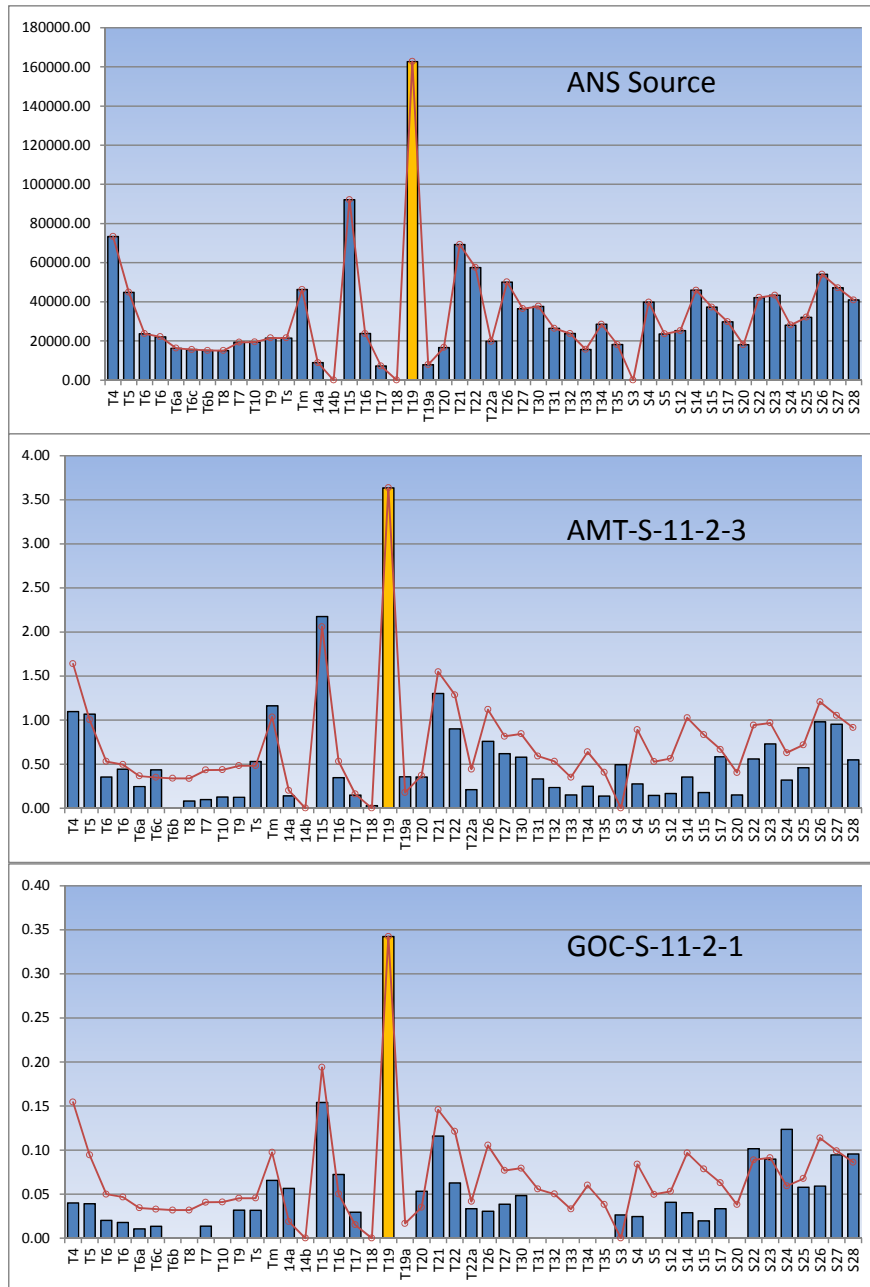
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Figure 5 GC/MS m/z 191 ion traces for ANS reference oil from Auke Bay Laboratory (upper plot) and a similar North Slope oil (lower plot) (Lilis et al., 1998) demonstrating comparability of detection.

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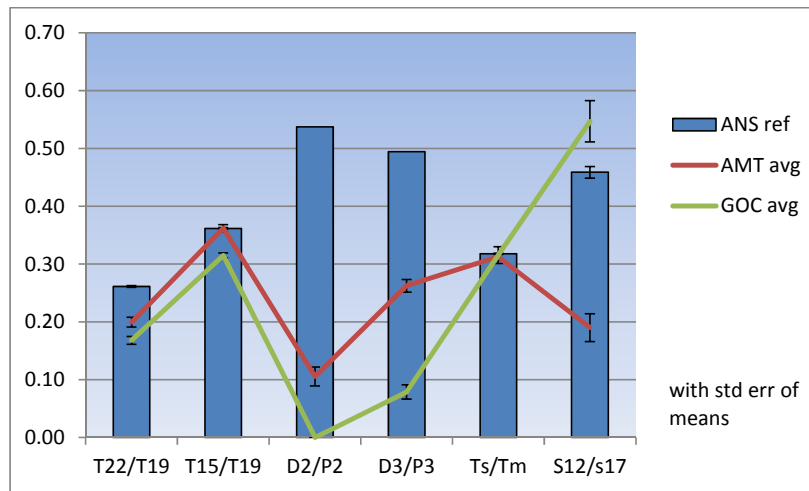


150

151 Figure 6 Representative ABL-quantified biomarker plots of EVOS (ANS) Crude Oil standard (top), AMT sediments from July  
 152 2011 (center), and GOC sediments from July 2011 (bottom). Red line denotes ANS reference normalized to the sample's  
 153 hopane (highlighted in gold).

154 From the 2011 and 2012 LTEMP collections, eighteen samples were analyzed. Visual inspection of individual plots  
 155 for sediment biomarkers and their associated diagnostic ratios (Appendix 2) show general agreement by stations  
 156 and across sampling years. The AMT sites are consistently similar to the ANS reference profile in the mid-hopane  
 157 suite of analytes (Ts through T22) (Figure 6). At GOC, the "fit" is similar but looser, and there are more non-  
 158 detected analytes because the concentrations of biomarkers are lower (total biomarkers avg: 31.0 ng/g DW at  
 159 AMT vs. 4.4 at GOC). Thus, there are some concerns with detection limits. However, even with a looser fit, the

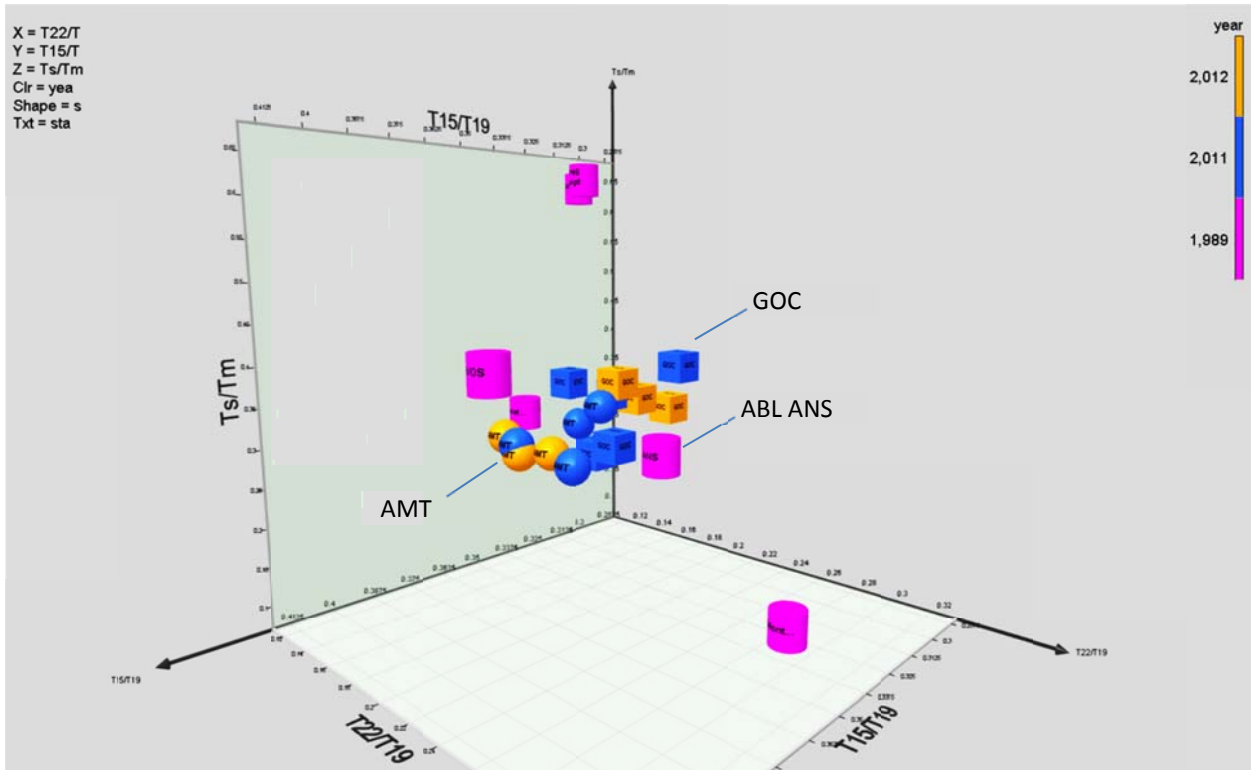
160 selected GOC ratios are in agreement with the reference (Figure 7) for the three selected diagnostic ratios (T22,  
 161 T15 and Ts/Tm), and when GOC's total biomarker concentrations peak (GOC-S-11-2-1D, 7.1 ng/g DW and GOC-S-  
 162 12-2-1, 9.5 ng/g DW), the fit is excellent (Appendix 2).



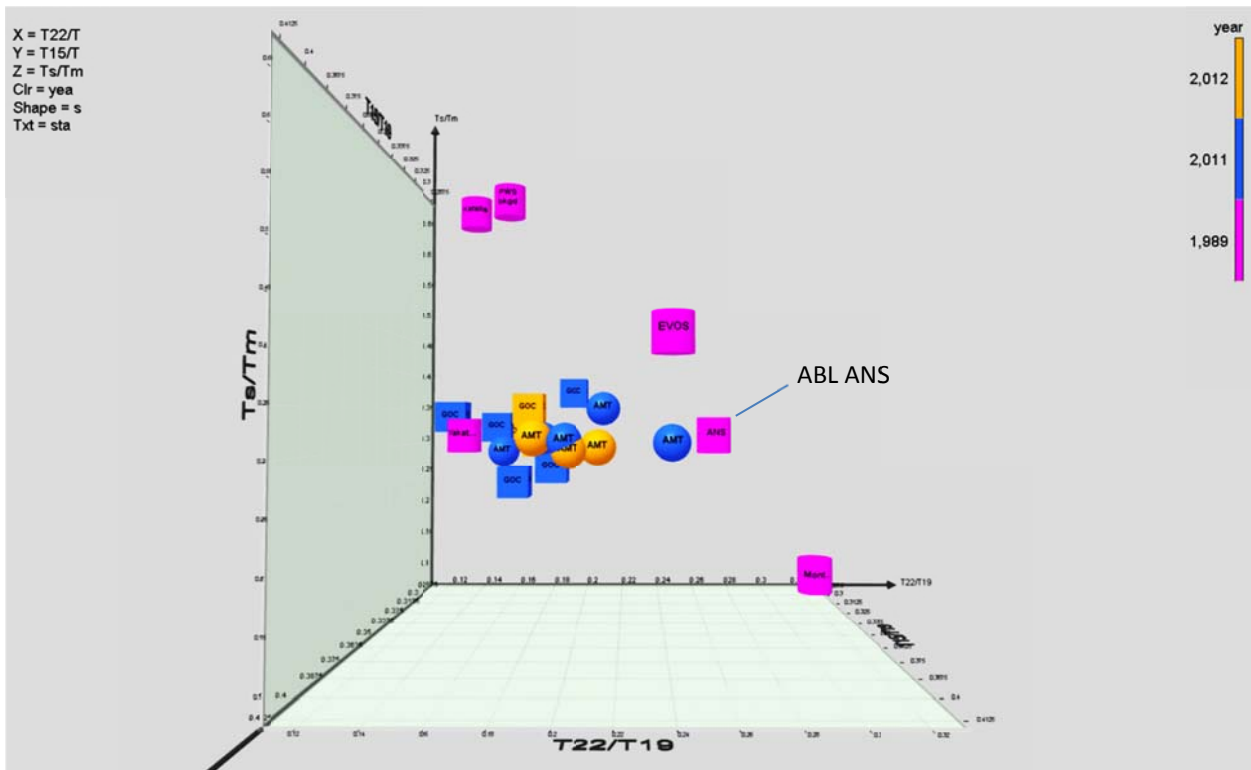
163  
 164 Figure 7 Comparison of average GOC and AMT diagnostic ratios with ANS reference. Only the three best ratios were used for  
 165 final evaluations.

166 Plotting the diagnostic ratios in 3D space (Figure 8) shows samples clustering by station. Both stations differ from  
 167 the EVOS-era ANS reference (in pink) with AMT (spheres) being more similar (closer) to the reference. Also note  
 168 that GOC tends to remain tightly clustered across years (in these three parameters), more so than at AMT, which  
 169 surprisingly suggests more a variable background influence in the AMT biomarker signature. Additional reference  
 170 samples (described below) are also shown.

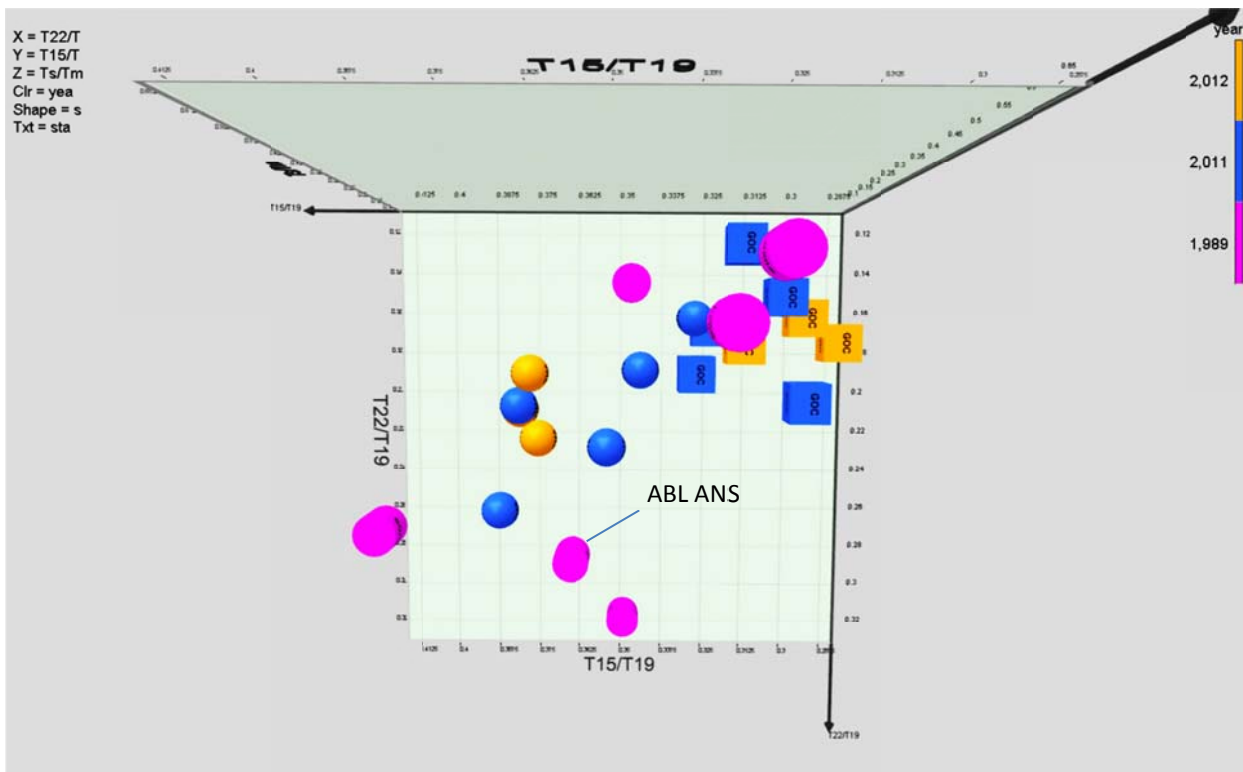
171 In these plots, the primary distinction between stations is driven by norhopane (T15) levels. But these are just the  
 172 fine-scale differences; overall, these profiles are significantly different from two other petrogenic sources found in  
 173 the Sound. As depicted by Bence et al. (1996) (Figure 9), the ANS-origin EVOS oil differs from Katalla and Yakataga  
 174 regional sources east of PWS, primarily by the absence of the oleanane biomarker (seen just before the C30  
 175 hopane) plus various subtler differences (e.g., note Ts and Tm proportions in all non-EVOS plots). In the Gulf of  
 176 Alaska, the Katalla/Yakataga materials become finely fragmented and get transported westward by Gulf currents  
 177 eventually reaching all the way to Shelikof Straits and down the Alaska Peninsula but with some deposited in the  
 178 PWS basin to become the much disputed (during-EVOS-era) PWS background. Also found in intertidal regions of  
 179 the Sound are tarballs from Monterey formation oil (Kvenvolden et al., 1995), remnants from oil spilled during the  
 180 1964 earthquake. Imported in the pre-pipeline era from California, Monterey oil is also distinct from ANS oil by its  
 181 inclusion of both oleanane and 28,30-bisnorhopane (just before norhopane Figure 9). Oleanane comes from the  
 182 flowering angiosperms which only appeared in the later Triassic formations; apparently the North Slope deposits  
 183 never saw the bloom. In one of the eighteen LTEMP sediment samples, a trace level of oleanane was detected,  
 184 again confirming that like the dispersion seen in the 3D plots, the biomarker patterns represent mixtures with  
 185 other background biomarker sources.



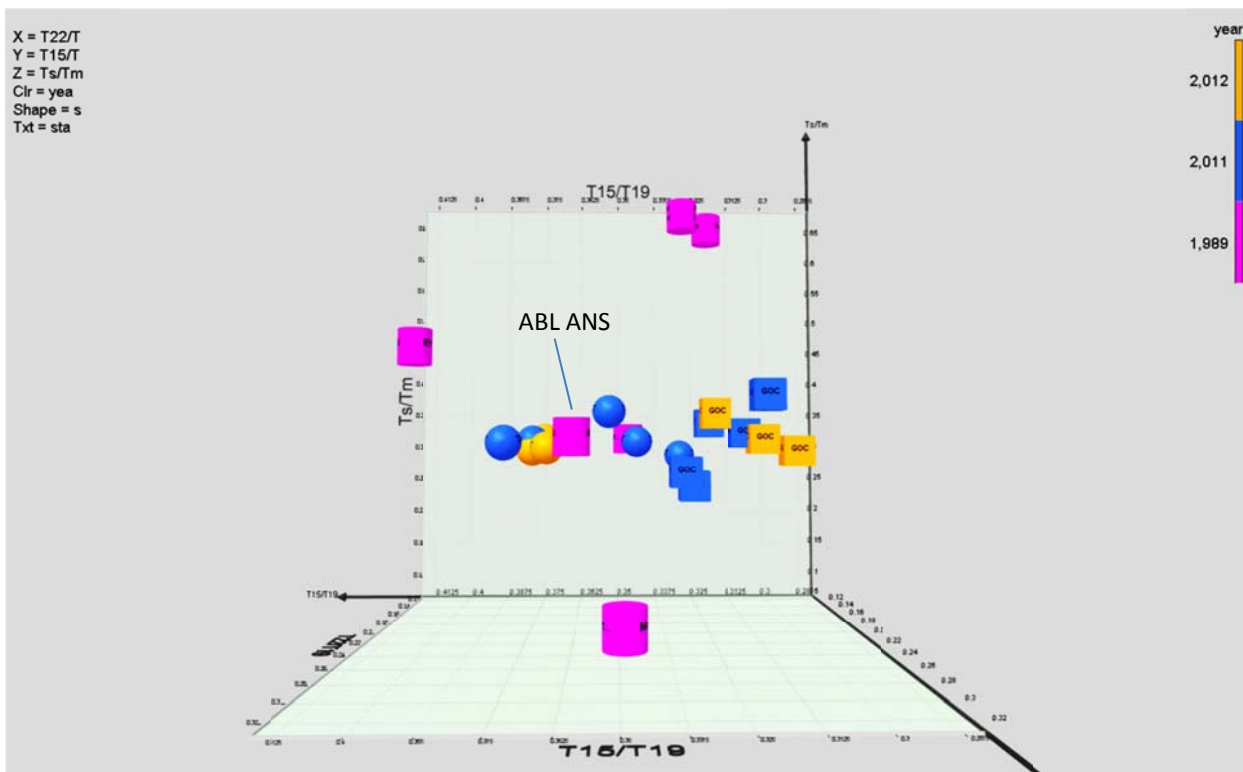
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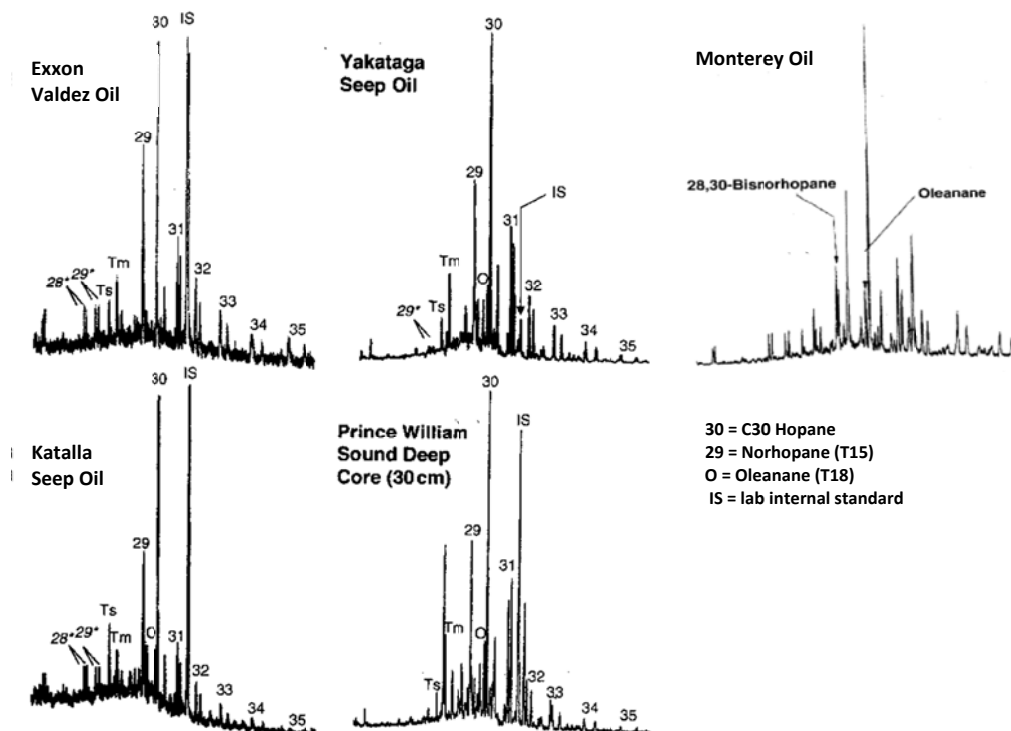
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189

190 Figure 8 Diagnostic biomarker parameters showing discrete 3D clustering of AMT (spheres) and GOC (cubes) sediment  
 191 samples; oblique, front, top and right perspectives. Reference oils as pink cylinders; other colors indicate sampling years (1989,  
 192 2011 & 2012).

193



194 Figure 9 Ion traces of hydrocarbon sources found in PWS. Adapted from Wang and Stout, 2007.

195 Our preliminary assessments, pending accumulating a larger time series of samples, additional lab experience, and  
 196 a more current ANS source sample, lead us to conclude that the mid-hopane-centric diagnostic ratios confirm the  
 197 presence of ANS-derived biomarkers in both AMT and GOC sediments. Further, the form of variation in diagnostic  
 198 biomarkers is site specific with both locations showing variants of the ANS source (i.e., weathering and/or mixtures  
 199 with background biomarkers). These findings are corroborated by Shaw et al. (2005) in Alyeska's EMP, who, using  
 200 C30 hopane and C29 norhopane, reported finding ANS biomarkers in deep Valdez basin sediments.

## 201 SEDIMENT GRAIN SIZE

202 Sediment grain size samples were recently analyzed for all 2006-2013 collections, comprising both historic and  
 203 "future" samplings. These data are presented in two formats: the standard cumulative (%) grain size curves and in  
 204 3D plots (Figure 10 & Figure 11). For this project, the grain size data only serve to demonstrate the constancy and  
 205 comparability of the sampling site environs. Both sites are dominated by glacial flour inputs, showing  
 206 approximately equal portions of clay and silt with minor sand components (Table 3) and both sites show minor  
 207 trends and outliers. In the 3D plots, note there are annual shifts to higher sand content at both locations albeit still  
 208 a minor component (<10%) and with a return to original conditions at GOC in 2013. More dramatically, at AMT,  
 209 there has been a steady increase in clay content. The locations occupy a heterogeneous fjord floor dynamically  
 210 swept by tidal currents (and prop wash at AMT) plus with sampling guided by GPS, the sites have been accruing  
 211 grab-sampler pock marks and drag scars at the same locations for 20 years. Reassuringly, when we get off-site at  
 212 GOC, we begin to see gravel in the grab. These shifts are noted with only modest confidence considering the non-  
 213 rigorous collection methods, i.e., spooning up 250 ml of sample remnants after collecting the less consolidated  
 214 surface floc for hydrocarbons.

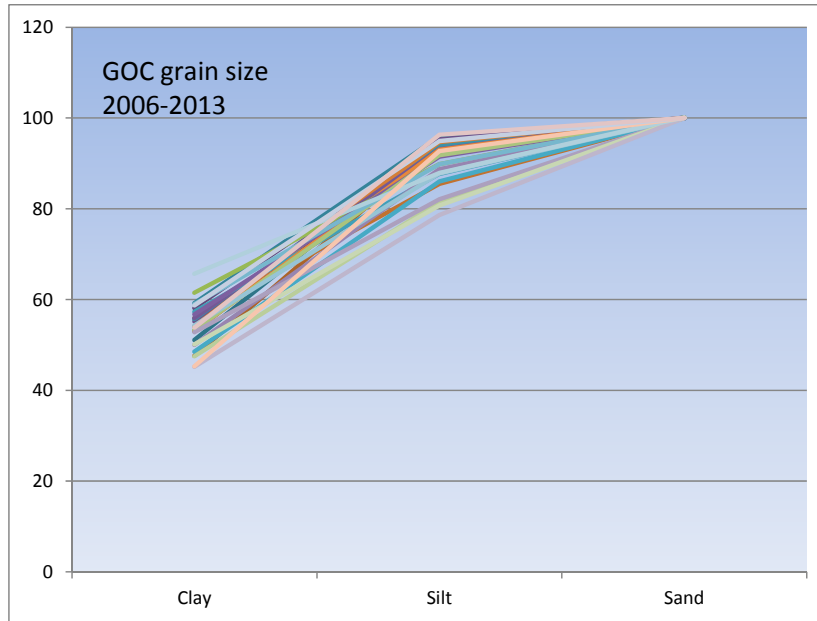


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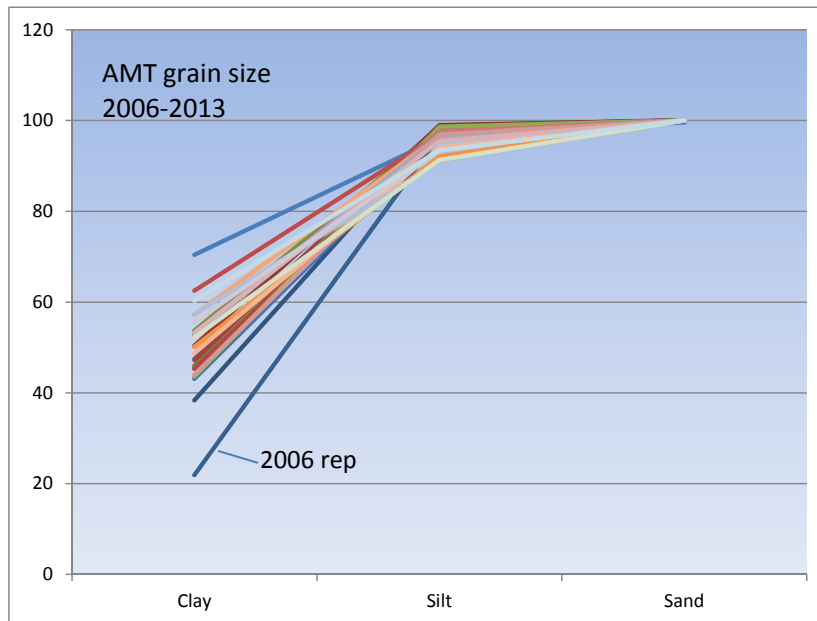
215 Table 3 Average grain size components for GOC and AMT, 2006-2013

	% Clay	% Silt	% Sand
AMT	49.9	46.4	3.6
GOC	54.0	36.2	9.8

216

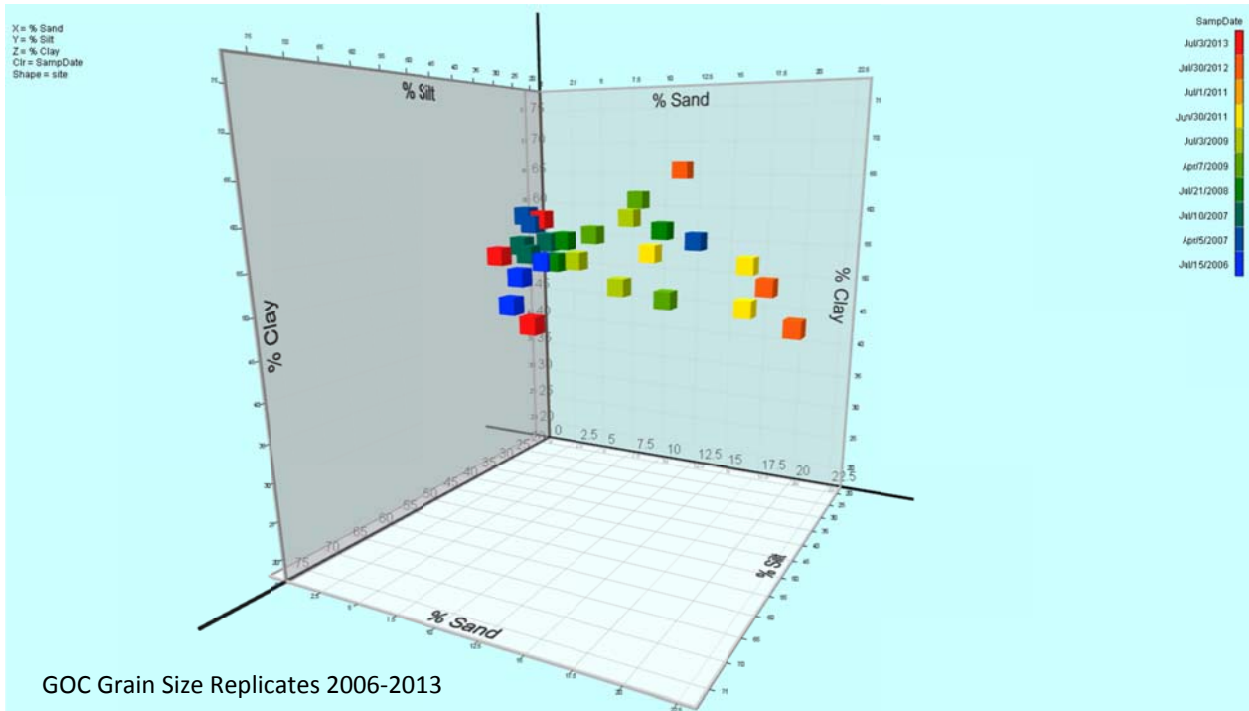


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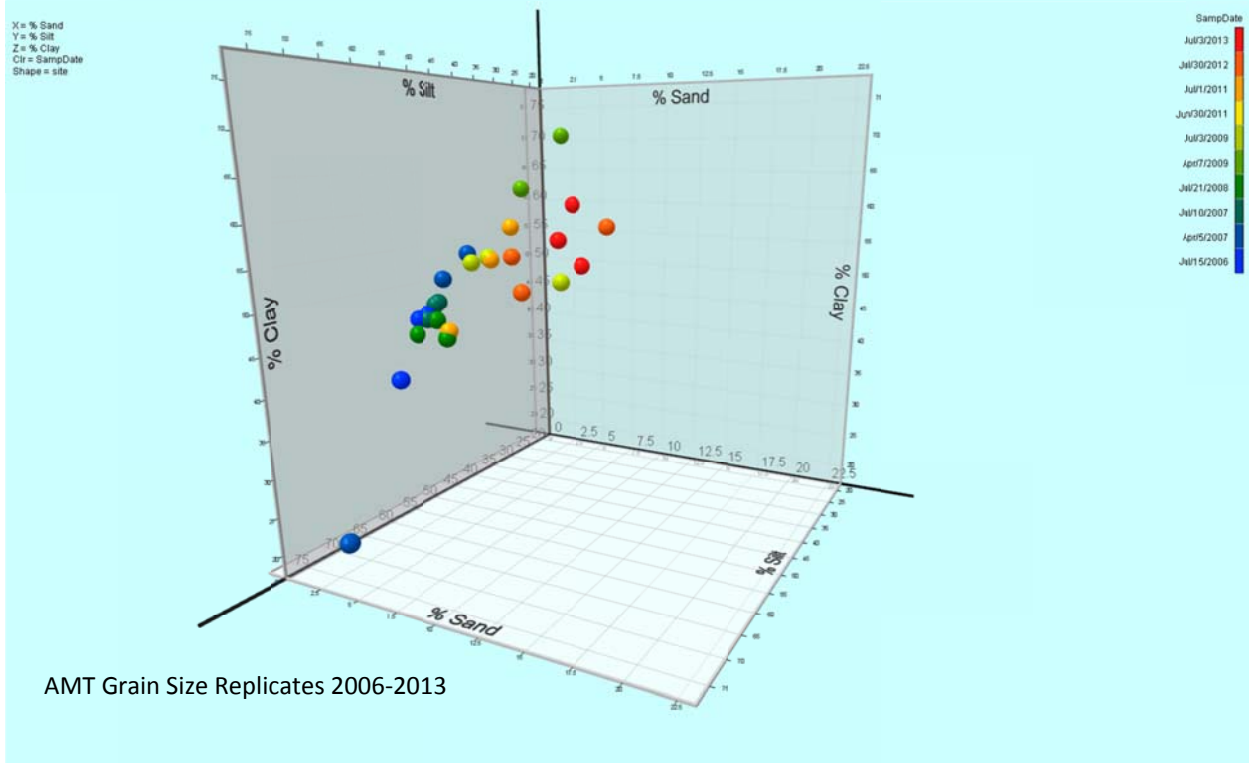


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219 Figure 10 Cumulative grain size curves (%) for GOC and AMT, 2006-2013.



220



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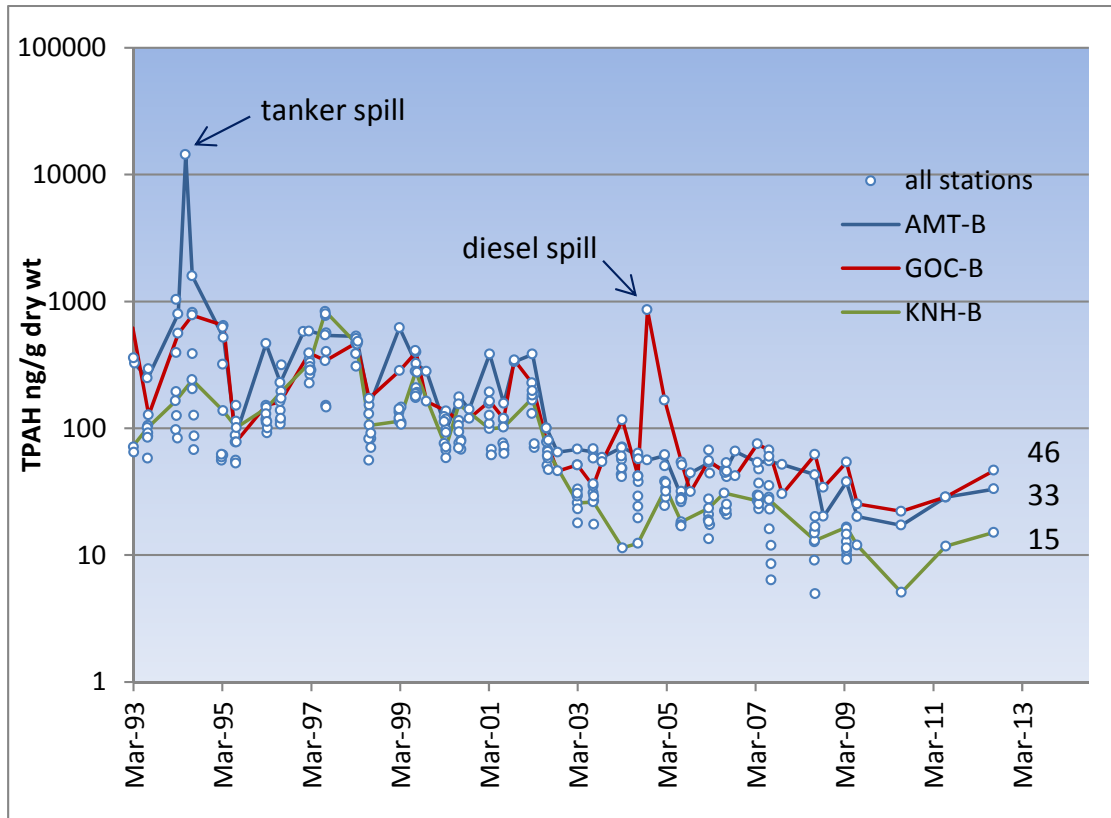
222 Figure 11 3D plots of grain size components from GOC and AMT 2006-2013. Sampling years color coded (blue 2006, red  
223 2013). Note shifts to higher clay content at AMT plus higher sand content at both and then retro-shift in 2013 GOC samples back  
224 to original 2006 cluster.

## 225 TISSUES IN PORT VALDEZ AND KNOWLES HEAD

226 In July 2010, average TPAH levels in AMT tissue samples approached all-time lows, ~17 ng/g DW (Figure 12), and  
227 although they appear to have increased slightly (~46 ng/g DW) in July 2012, the majority of the PAH are pyrogenic  
228 not petrogenic in nature (see Figure 13 and other profiles in Appendix 1). Also note the appearance of perylene in  
229 July 2011 and 2012. This is a naturally occurring 4-ring unsubstituted PAH generated by biologic processes or in  
230 the early stages of diagenesis in marine sediments (Bence et al., 2007) and potentially being neither a non-  
231 petrogenic or pyrogenic PAH, is not included in TPAH calculations. Most of the saturated hydrocarbons (SHC) in  
232 AMT mussel tissues throughout this period are dominated by below-MDL biogenic constituents (e.g., n-C<sub>15</sub>, n-C<sub>17</sub>,  
233 pristane, and n-C<sub>27</sub>). In July 2011, two of the three replicate tissue samples also showed traces of below-MDL,  
234 higher-molecular-weight SHC in the C<sub>23</sub>-C<sub>30</sub> range suggesting a possible petrogenic input, but this was not reflected  
235 in the PAH patterns.

236 Average TPAH levels in GOC mussel tissues have consistently been very low except for the Fall 2004 diesel spill  
237 (cleared by the following July – Figure 12). Beginning in the 1990's, mixed dissolved-phase, petrogenic, and  
238 pyrogenic signals were common at this site, roughly trending with similar patterns or discharge events at the  
239 Alyeska Marine Terminal (Payne et al., 2008a,b; 2010a). Between March 2003 and July 2008, with the exception of  
240 the aforementioned diesel spill, the signals became largely pyrogenic with occasional petrogenic components  
241 (DBTs) (Figure 14). From July 2005 to present, the TPAH levels have consistently been in the 28-75 ng/g DW range.  
242 SHC patterns at this site over the 2008-2012 period have been almost exclusively trace-level biogenic (e.g., n-C<sub>15</sub>,  
243 n-C<sub>17</sub>, pristane, and n-C<sub>27</sub>). Exceptions included two of three replicates in April 2009, and one of three replicates in  
244 June/July 2011 when traces of petrogenic C<sub>24</sub>-C<sub>30</sub> SHC appeared (see Appendix 1). In none of these cases, however,  
245 were petrogenic sources suggested by the PAH profiles. Interestingly, traces of perylene noted in the AMT mussels  
246 in July 2011 and 2012 were also observed at GOC during the same collections.

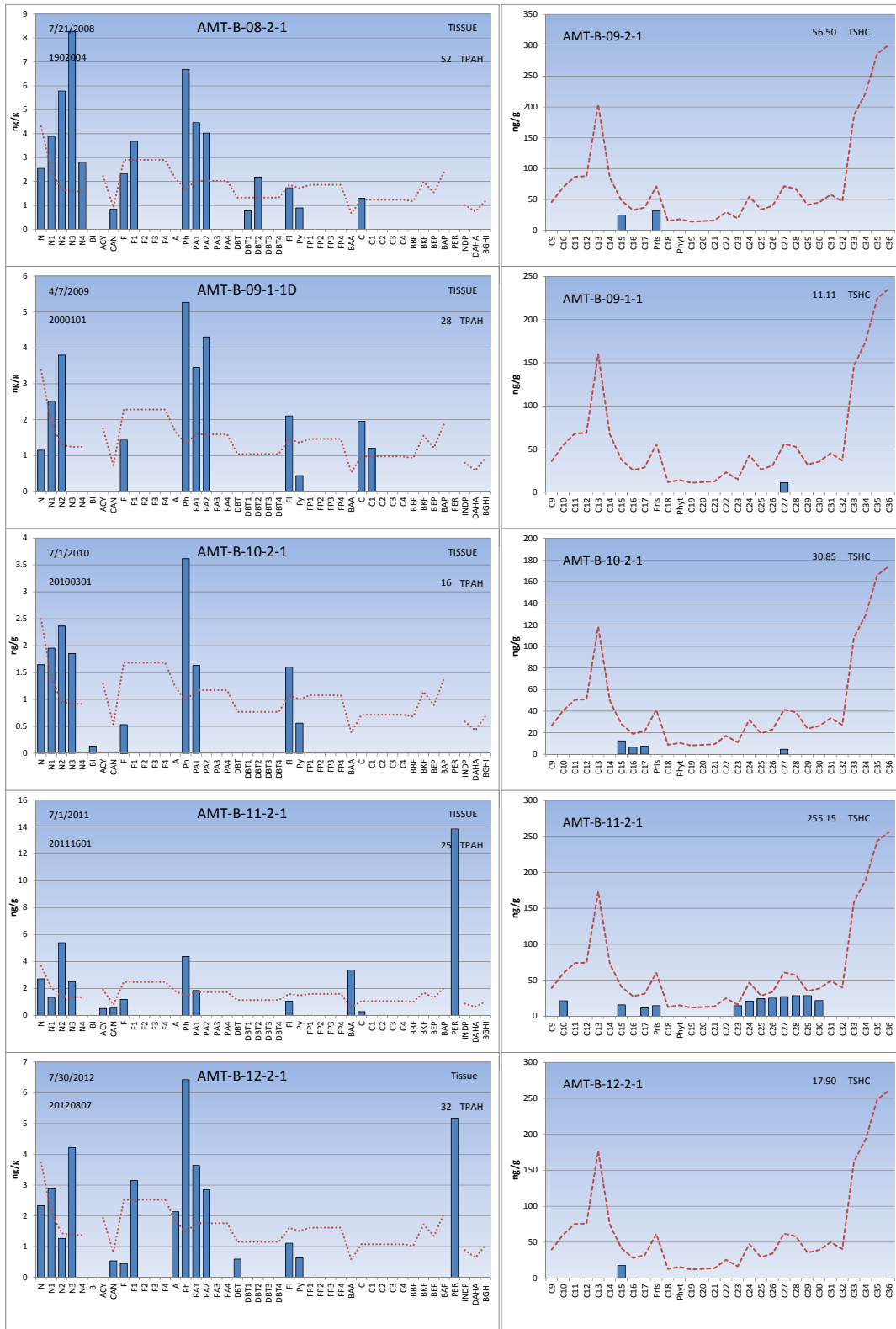
247 Average TPAH levels in mussel samples from the tanker anchorage at KNH have been consistently low, ranging  
248 from 5-24 ng/g DW since March 2004 (Figure 12). With the exception of one sample in July 2004 that showed a  
249 petrogenic PAH profile (Payne et al., 2008a,b; 2010), the samples from this site have been dominated by trace level  
250 (at or just-above detection limit) naphthalenes and combustion-derived anthracenes/phenanthrenes since July  
251 1999; this pattern has continued throughout the July 2008-July 2012 period (see Figure 15 and additional profiles in  
252 the Appendix 1). Also note the appearance of perylene that was also observed in the mussels at AMT and GOC in  
253 2011 and 2012. Presumably this reflects some region-wide marine event. The near- or just below-MDL SHC  
254 patterns at KNH between July 2008 and July 2012 are almost exclusively biogenic (e.g., n-C<sub>15</sub>, n-C<sub>17</sub>, and pristane)  
255 except for one of three replicates in July 2011 when a slightly fuller, but below-MDL C<sub>24</sub>-C<sub>30</sub> pattern suggested the  
256 possibility of a petrogenic source. The PAH profile, however, did not support this observation (Figure 15).



257

258 Figure 12 Time series of mussel tissues TPAH from Alyeska Terminal, Gold Creek and Knowles Head sites.

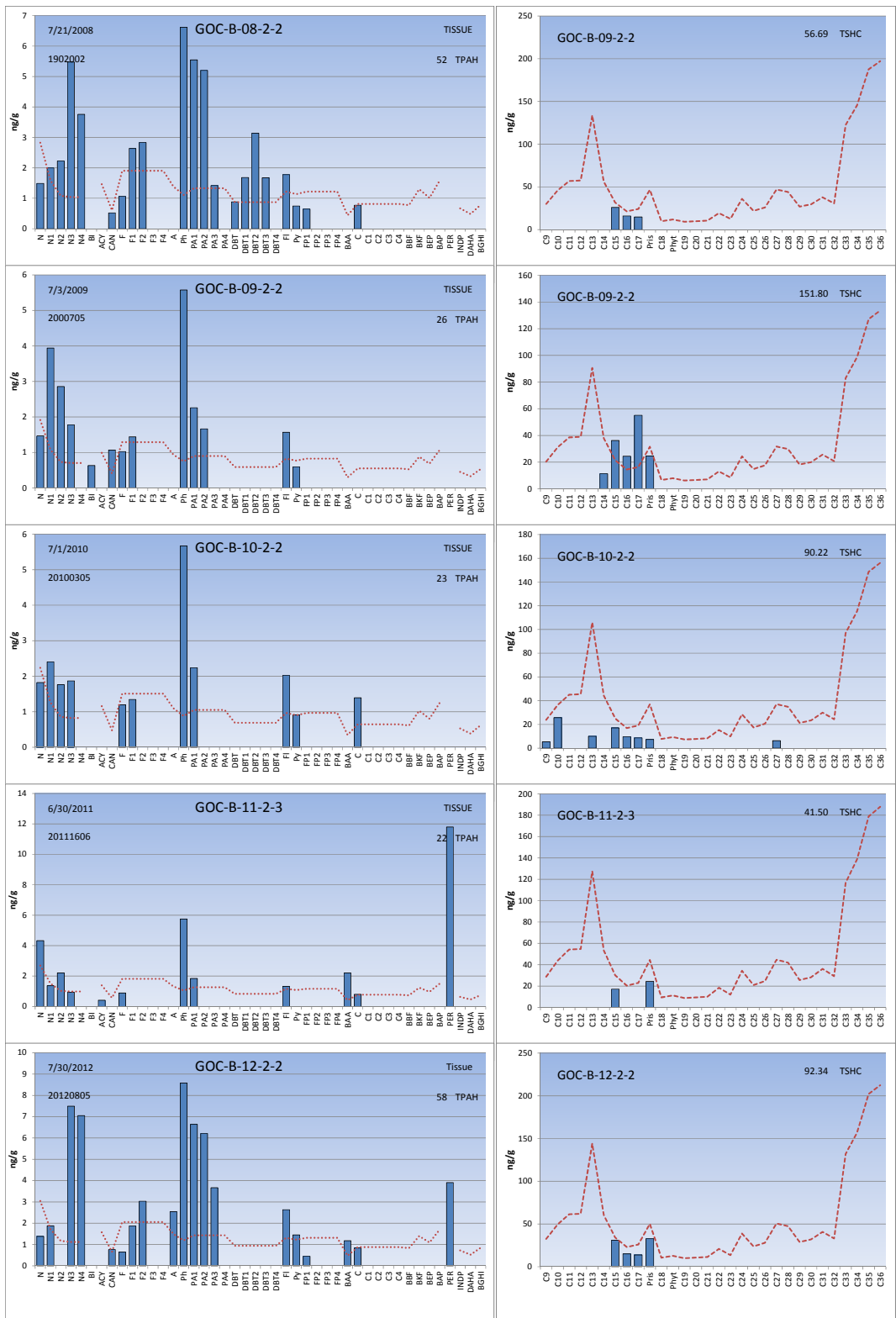
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259

260 Figure 13 Representative PAH and SHC profiles from AMT mussel tissues between July 2008 and July 2012 showing primarily  
 261 combustion products, perylene, and planktonic biogenic SHC and a trace of crude oil (in July 2011). Red dashed line is sample-  
 262 specific MDL.

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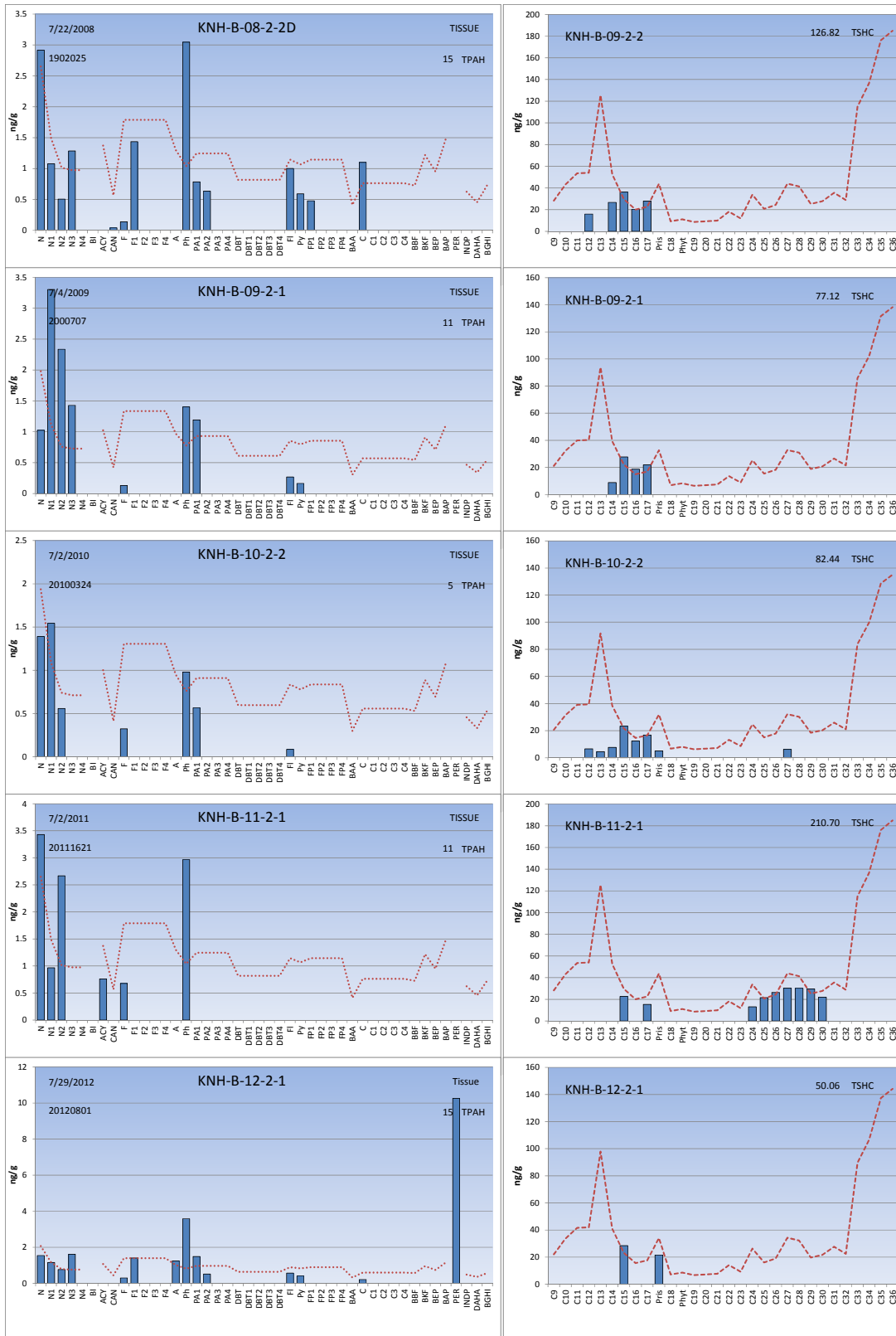


263

264 Figure 14 Representative PAH and SHC profiles from GOC mussel tissues between July 2008 and July 2012 showing primarily  
 265 combustion products, perylene, and planktonic/terrestrial biogenic SHC. Note the occurrence of perylene in 2011 and 2012 was  
 266 also observed in the AMT mussels. Red dashed line is sample-specific MDL.

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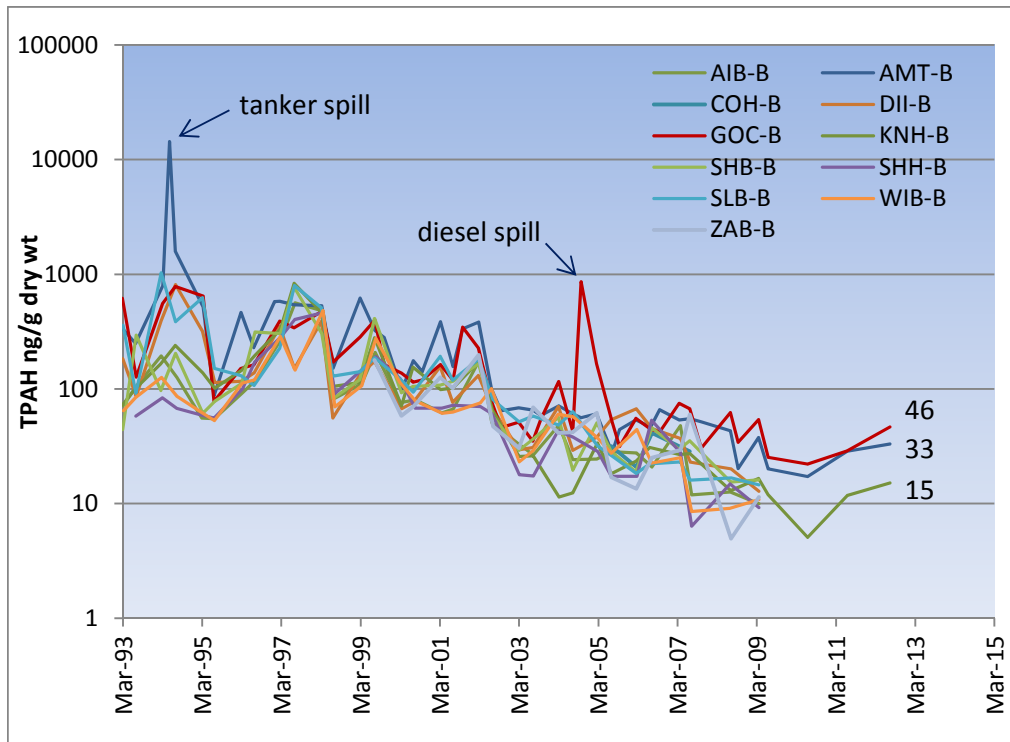
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270

Figure 15 Representative PAH and SHC profiles from KNH mussel tissues between July 2008 and July 2012 showing primarily trace-level combustion products, perylene (in 2012), and planktonic/terrestrial biogenic SHC.

271 PRINCE WILLIAM SOUND AND GOA MUSSEL TISSUES

272 Only one additional sampling (April 2009) has been added to the time series since the publication of our 2011 data  
 273 report, and as of that sampling, overall TPAH levels at these central PWS and GOA sites continued to decline  
 274 (Figure 16) although more recently, the annually-sampled Valdez sites have marginally rebounded.



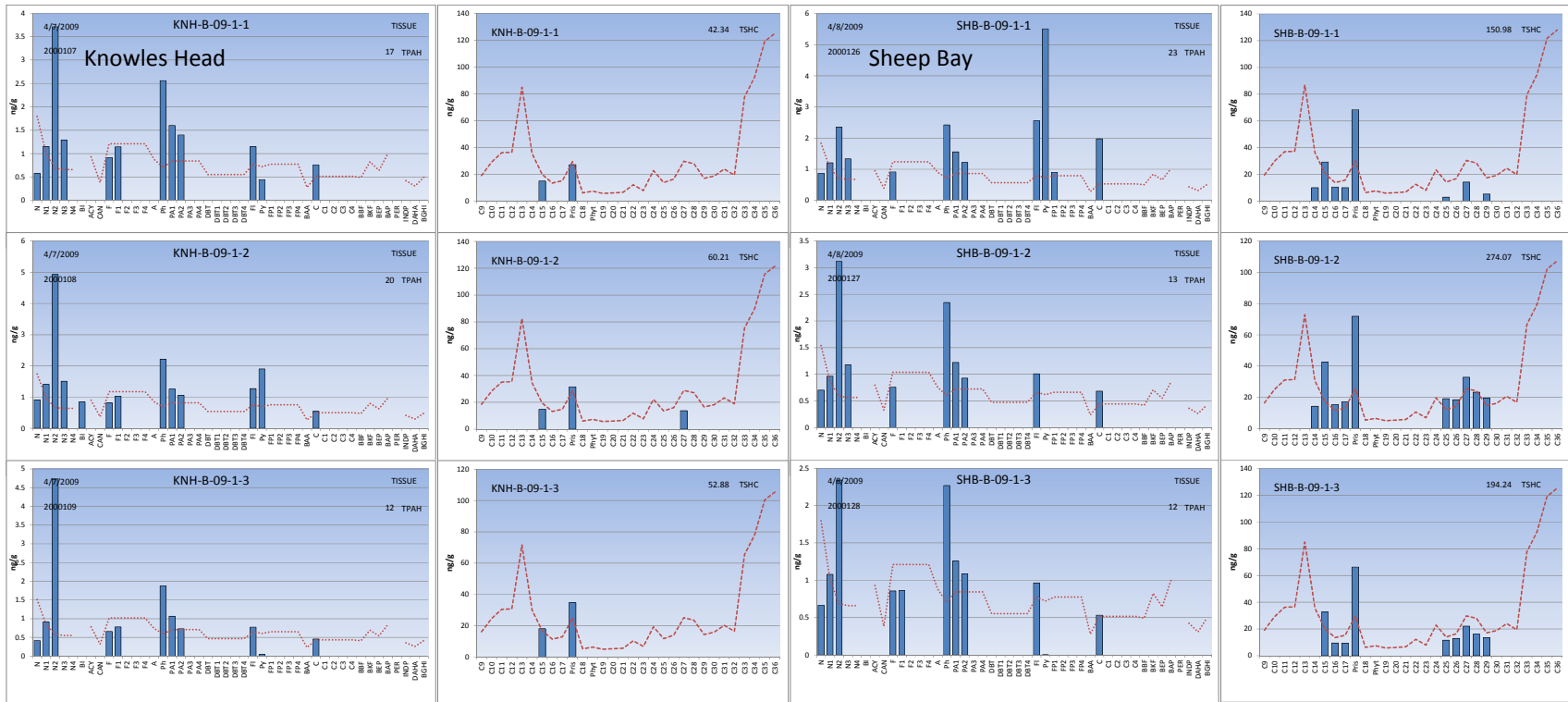
275  
 276 Figure 16 Mussel-tissue TPAH trends for all stations (1993-2012).

277 All of the PAH levels measured in the April 2009 samples are extremely low (mostly near or just below even the  
 278 new MDLs), yet remarkable site fidelity exists within the PAH and SHC patterns for each stations' replicates (Figure  
 279 17 through Figure 20). And to a degree, regional patterns are still visible although weaker than in the previously  
 280 reported 2006-08 data set (Payne et al., 2010a). We believe that despite the at-or-below MDL character of the  
 281 data, the within-site and within-region consistency would not be achievable if data quality were uncontrolled at  
 282 these trace-levels. The laboratory method blanks were also very clean (primarily only showing below-MDL traces  
 283 of the problematic biphenyl), which again supports the notion that rather than laboratory noise, the observed  
 284 patterns were truly due to low-level constituents in the mussel tissues at each site.

285



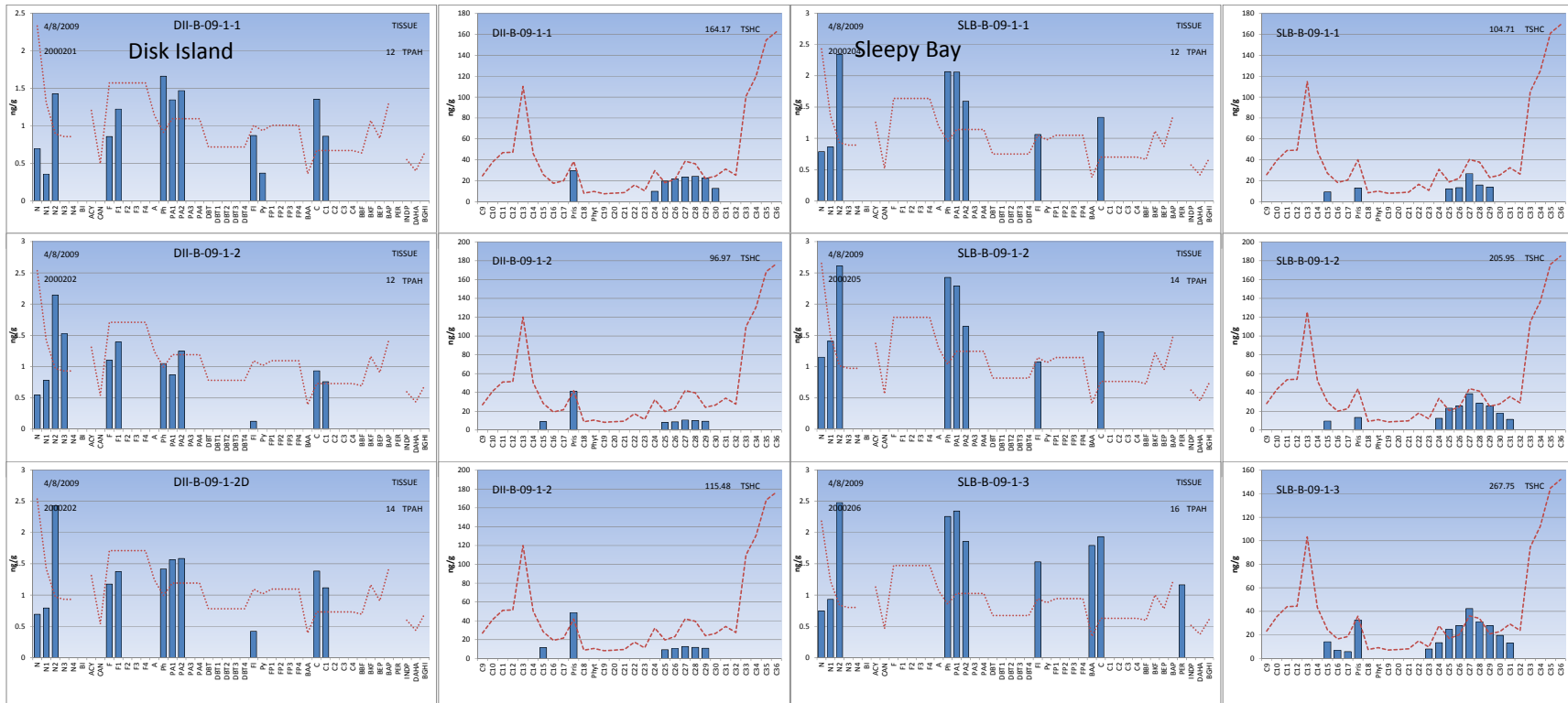
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286

287 Figure 17 Mussel-tissue PAH histogram plots showing within-site fidelity among replicates and regional similarities between eastern PWS sites.

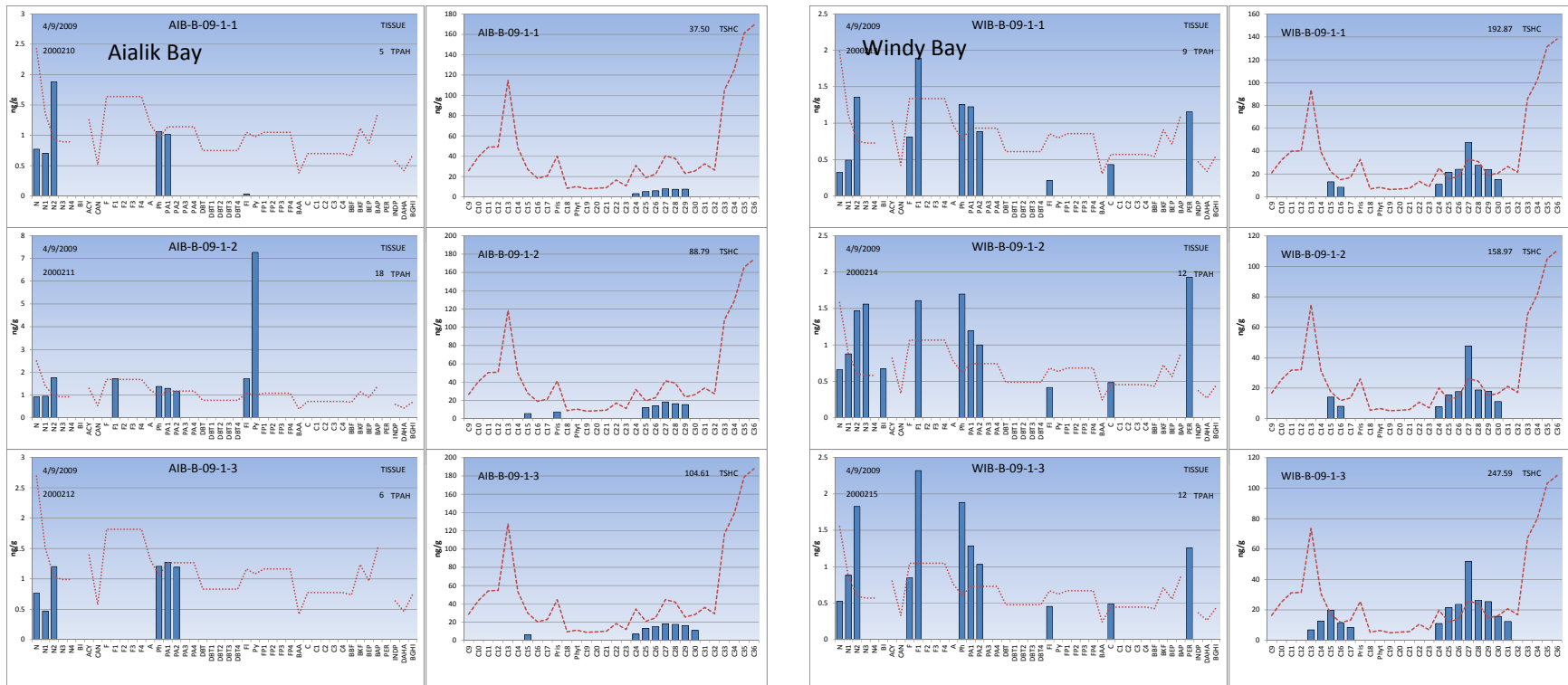
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288

289 Figure 18 Mussel-tissue PAH histogram plots showing within-site fidelity among replicates and regional similarities between central and previously EVOS-oiled PWS sites.

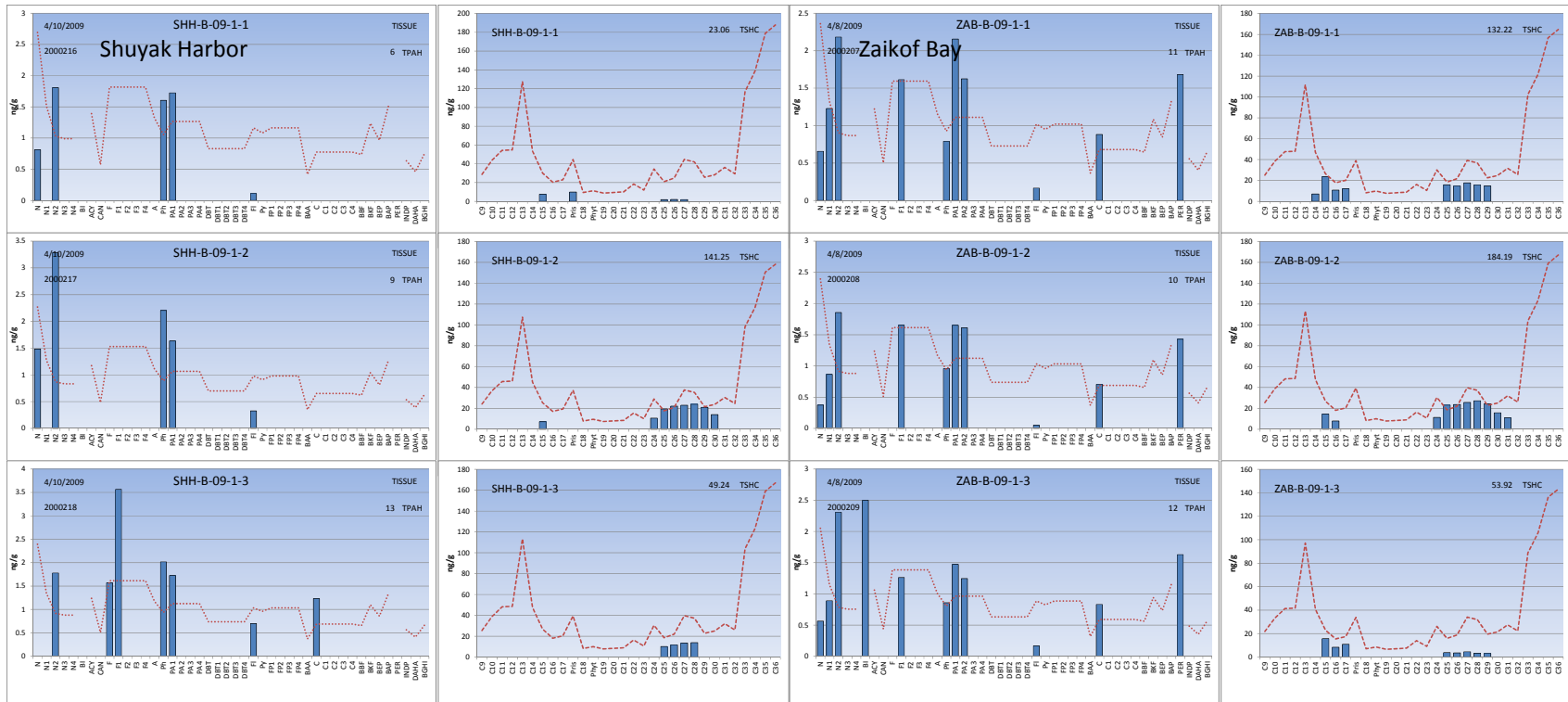
# LTEMP Report 2008-2012



290

291 Figure 19 Mussel-tissue PAH histogram plots showing within-site fidelity among replicates and regional similarities between outer coast sites.

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292

293 Figure 20 Mussel-tissue PAH histogram plots showing within-site fidelity among replicates.

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294 All of the PWS and GOA stations (Figure 17 through Figure 20) show ubiquitous low-level naphthalene signals that  
295 we think are part of the overall background pattern (Payne et al., 2010a,b). Likewise, Knowles Head (KNH), Sheep  
296 Bay (SHB), and Windy Bay (WIB) show pyrogenic phenanthrenes/anthracenes along with lower concentrations of  
297 fluorenes and isolated peaks for parent fluoranthene, pyrene, and chrysene (all combustion products). The  
298 relative phenanthrene/anthracene homologue distribution patterns at Disk Island (DII) and Sleepy Bay (SLB) (both  
299 formerly EVOS-oiled sites) show a little more complexity and within-site variability suggesting the possibility of a  
300 more petrogenic influence, e.g., oil was still observed in disturbed intertidal substrates at DII in April 2009 (Payne  
301 et al., 2010a). Zaikof Bay (ZAB) also shows evidence of a minor petrogenic phenanthrene/anthracene pattern, and  
302 its fluoranthene/chrysene/perylene profile is unique among all the stations. Finally, Shuyak Harbor (SHH) and  
303 Aialik Bay (AIB) (with the exception of one replicate) show the least complex PAH patterns of all the stations  
304 examined. The SHC patterns at all of these sites are mostly below the sample-specific MDLs, but there appears to  
305 be sporadic indications of petrogenic SHC (n-C<sub>23</sub>-n-C<sub>31</sub>), again with remarkable site fidelity at several of the  
306 stations.

307 In the Sound, with the bulk of EVOS oil either dissipated or the remnants buried and sequestered, the regional  
308 sites that are physically remote from the chronic low input of BWTF discharge are currently looking very clean. As  
309 stated in our last report, TPAH levels in both the PWS and GOA regions are dropping at nearly identical rates. This  
310 parallelism suggests the decreases are influenced by similar (oceanographic-scale?) processes, but the occasional  
311 asynchrony of peak events also suggests regional variation in the dynamics. So what broad-scale sources might  
312 appear as an ambient background signal? Possibilities include atmospheric deposits from forest fires and volcanic  
313 activity, leaching of water-soluble constituents from the pervasive source-rock (oil shale or coal that constitutes  
314 much of the PAH-bearing material being transported through the region), some upwelling/climate-driven events, a  
315 combination of these or some novel mechanism relating to decadal oscillations and/or global warming. Based on a  
316 limited-scale partitioning experiment with clean seawater and intertidal sediment from Constantine Harbor (Payne  
317 et al. 2010b), we now believe that sedimentary material previously believed to only contain sequestered and non-  
318 bioavailable PAH, is a component of the ubiquitous dissolved-phase (naphthalene) signals observed throughout  
319 Prince William Sound.

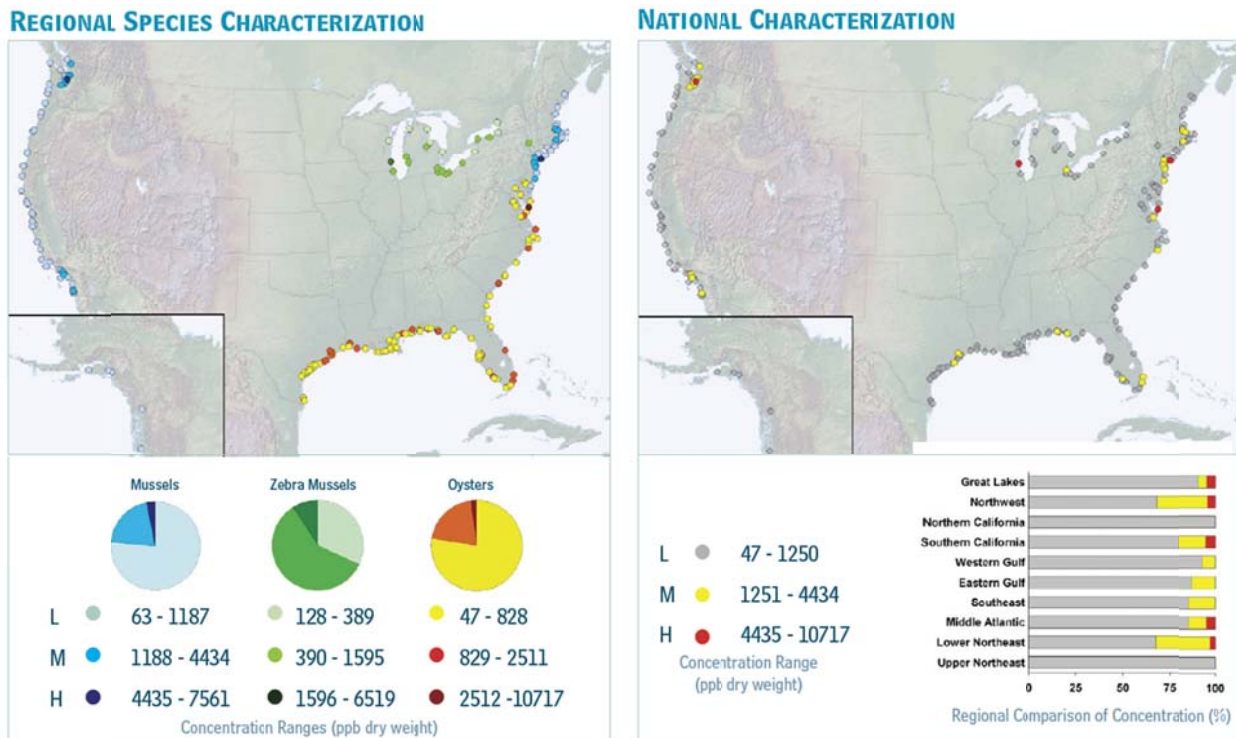
320 Another question is how low will this generally declining trend in TPAH values go? Obviously, at some point, the  
321 trend must level out. Follow-up sampling in 2004 for oil residues from the 1997 *M/V Kuroshima* grounding in  
322 Summer Bay, Unalaska, found TPAH levels between 25 and 85 ng/g DW, with an average of 57 ng/g DW (Table 4,  
323 Helton et al., 2004). This compares favorably and is actually higher than LTEMP's April 2009 range of 4-23 ng/g DW  
324 inside the Sound and 6-18 ng/g DW at Gulf of Alaska sites (overall average 12 ng/g DW). These data also suggest a  
325 natural dissolved-phase background TPAH somewhere below 20-30 ng/g DW--a range in which analytical  
326 sensitivity can be highly susceptible to procedural artifacts. It could easily be the case that the LTEMP data are  
327 currently tracking subtle variations in the background PAH, and that we are near or at the minimum.

328 For relative comparisons, data from the 2004-2005 National Status and Trends, Mussel Watch Program (Table 4,  
329 Figure 21) and 2008-2010 Alaskan sites (Figure 22) (now summing 38 parent and alkylated PAH homologues versus  
330 44 LTEMP PAH analytes) show that PAH concentration in mussels for other West Coast sites is nearly 66 times  
331 higher at 825 ng/g DW. The highest level reported on the West Coast was 6,962 ng/g DW in Seattle, Washington.  
332 The lowest, 63 ng/g DW, was from mussels collected on Santa Catalina Island, 26 miles offshore of Orange County  
333 in Southern California. In 2004-5, the average TPAH concentration in mussels from the five Alaskan Mussel Watch  
334 sites (Ketchikan, Nahku Bay, Port Valdez, Unakwik Inlet, and Cook Inlet) was 267 ng/g DW with levels ranging from  
335 105-441 ng/g DW (Kimbrough et al., 2008). Considering these and even more recent 2008-2010 values from the  
336 MW data portal (Table 4), the LTEMP results for the PWS and GOA sites demonstrate that these remote locations  
337 are exceptionally clean.

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338 Table 4 Current TPAH concentrations in regional mussel tissues (ppb, ng/g DW) relative to 2004-2010 NOAA Mussel Watch  
 339 monitoring data and a recovered Alaskan oil-spill event.

LTEMP	2012	Port Valdez	15-73	
	2009	PWS	4-23	
	2009	GOA	6-18	
West Coast Mussel Watch	average (Kimbrough et al., 2008)		825	
	So. Calif.	Santa Catalina Island	63	
	Seattle	Elliot Bay, WA	6,962	
Alaska Mussel Watch	2008	Cook Inlet	Bear Cove	119
		Cook Inlet	Homer Spit	208
		Port Valdez	Alyeska Marine Terminal	52
	2009	Port Valdez	Gold Creek	31
		Ketchikan	Mountain Point	231
		Nahku Bay	East Side	229
		Port Valdez	Mineral Creek Flats	332
		Resurrection Bay	Nash Road	602
		Resurrection Bay	Nash Road	765
		Resurrection Bay	Nash Road	929
		Resurrection Bay	Nash Road	713
		Unakwit Inlet	Siwash Bay	257
	2010	Resurrection Bay	Nash Road	1,570
		Nanwalek	Nanwalek	194
		Port Graham	Port Graham	376
Port Graham		Murphy's Slough	428	
Seldovia Bay		Powder Island	652	
M/V Kuroshima (1997)	Unalaska	2004 followup	25-85	



340

341 Figure 21 Status and trends result from national Mussel Watch data (Kimbrough et al., 2008). All Alaskan sites characterized as  
 342 low concentrations.

## APPENDIX 2: RESULTS BY STATE



### ALASKA (AK)

Regional (r) Status (s) Trend (t)  
 Mussels (M) National Status National Trend  
 ● Medium ● Medium ▼ Decreasing  
 ● High ● High ▲ Increasing

Zebra Mussels (ZM)  
 ● Medium  
 ● High

Oysters (O)  
 ● Medium  
 ● High

Concentrations derived from 2004-2005 data.

Markers represent the Regional Species Characterization (r), National Characterization (s) and National Trends maps (t).

Site	Latitude	Longitude	General Location	Location
KTMP	55.2938	-131.5480	Ketchikan	Mountain Point
NBES	59.4533	-135.3365	Nahku Bay	East Side
PVMC	61.1328	-146.4610	Port Valdez	Mineral Creek Flats
UISB	60.9608	-147.6460	Unakwit Inlet	Siwash Bay
CIHS	59.6145	-151.4442	Cook Inlet	Homer Spit

#### METALS (ppm)

Site	Spec	AS	r	s	t	CD	r	s	t	CU	r	s	t	HG	r	s	t	NI	r	s	t	PB	r	s	t	SN	r	s	t	ZN	r	s	t
KTMP	M	11		●		7.1	●	●		7				0.06				1.2				0.59				0				97			
NBES	M	9.2				5.4	●	●	▲	6				0.1				2				2.1	●			0				72			
PVMC	M	12	●	●		3.5		●		27	●			0.09				8.9	●	●		3	●	●		0.18				89			▼
UISB	M	12	●	●		2.6				33	●			0.11				7.4	●			2				1.4	●	●		108			
CIHS	M	12	●	●		1.7				10				0.12				3.4				1.3				0				105			

343

#### ORGANICS (ppb)

Sites	Spec	Butyltins	r	s	t	Chloroanes	r	s	t	DDTs	r	s	t	Dieldrins	r	s	t	PAHs	r	s	t	PCBs	r	s	t
KTMP	M	2.1				0.47				1.4				0.58				152				3.5			
NBES	M	3.7				2.7				2.2				0.98				316				7.7			
PVMC	M	7.3				2.6				1.7				0.31				441				6.4			
UISB	M	1.7				0.87				0.38				0.56				176				3.7			
CIHS	M	4.4				1.1				0.3				0.42				250				11			

344

345 Figure 22 Summary page of Alaska regional Mussel Watch results and trends based on 2004-05 data from Kimbrough et al.,  
 346 2008.

347 Finally, a 2005 EVOS Trustees Program, Long-term Monitoring of Anthropogenic Hydrocarbons in the Exxon Valdez  
 348 Oil Spill Region, examined ten intertidal sites within the Naked-Knight-Southwest Island complex to measure the  
 349 extent of buried oil still present 16 years after the spill. At previously heavily-oiled EVOS sites, 10 to 50 random  
 350 pits (depending on the beach width) were excavated to a depth of ~0.5 m to look for residual oil. If oil was  
 351 detected, available nearby mussels were also collected. The results have been published elsewhere (Short et al.,  
 352 2007) but, as co-authors, PAH and SHC sample profiles were included in Appendix E of our 2005/2006 Report  
 353 (Payne et al., 2008a). Briefly, TPAH levels in the oiled pits ranged from a low of 42 ng/g (on Knight Island) to a high  
 354 of 567,000 ng/g (on Latouche Island) with the oil showing states of weathering varying from very fresh to



355 extensively degraded. On the other hand, nearby mussel samples only showed low dissolved-phase TPAH (11-42  
356 ng/g DW, derived primarily from naphthalenes and phenanthrenes/anthracenes) that were in the same  
357 concentration range but compositionally different from the signals observed at LTEMP PWS and GOA sites sampled  
358 in April 2009. From these studies, it was concluded that although in 2005, there were still persistent buried EVOS  
359 residues at a number of beaches, they were highly sequestered and did not appear to be bioavailable unless  
360 disturbed. Rates of disappearance had diminished to an estimated 4% yr<sup>-1</sup>. If left undisturbed, Short et al. (2007)  
361 predicted they would be there for decades. To address these residual deposits, for the last three years EVOS  
362 Trustees have sponsored beach remediation projects.

## 363 CONCLUSIONS

364 Based on the sediment and mussel data, petrogenic hydrocarbon inputs from the Alyeska Marine Terminal and  
365 tanker operations continue to decline. Within Port Valdez, and as concluded in our last LTEMP report (Payne et al.,  
366 2010a), the decrease likely reflects a combination of reduced BTWF discharge volumes from decreased North  
367 Slope oil production, the transition to double-hulled tankers with segregated ballast tanks, and improved BWTF  
368 efficiency at removing particulate/oil-phase PAH. In the 2012 tissue collections at the Terminal, the mussels  
369 showed a predominant pyrogenic, not petrogenic, profile, and even the sediments at this site are beginning to look  
370 more like Gold Creek's background-reference, pyrogenic profiles. Biomarker data show evidence of low-level ANS  
371 oil in the Port sediments. At GOC, concentrations are seven times lower than those at AMT and the profiles  
372 indicate mixing with background sources.

373 With the April 2009 mussel collections at the central PWS and GOA sites, the overall TPAH levels at these stations  
374 continued to decline. All of the PAH levels measured in the April 2009 samples are extremely low (mostly near or  
375 just below even the new MDLs), yet remarkable site fidelity exists within the PAH and SHC patterns for each  
376 stations' replicates. The SHC patterns at all of these sites are mostly below the sample-specific MDLs, but there  
377 appears to be sporadic indications of petrogenic SHC (n-C<sub>23</sub>-n-C<sub>31</sub>), again with remarkable site fidelity at several of  
378 the stations. However, in corresponding PAH fractions when trace-level petrogenic SHC signals were observed,  
379 there was little or no evidence of oil contamination. Compared to 2004-05 West Coast Mussel Watch data and the  
380 more recent Alaskan Mussel Watch sites, the LTEMP results for the PWS and GOA sites demonstrate that the  
381 region is exceptionally clean.

## 382 ACKNOWLEDGEMENTS

383 This success of this project reflects the efforts of many more individuals than currently appear in the author's list.  
384 From our stalwart vessel operator, tireless supporter and colleague, Dave Janka (and Annette Janka), our primary  
385 pilots, Terry Kennedy and Jose DeCreft, our former and current colleagues at Auke Bay Lab (including the project  
386 newbie, Dr. L. Schaufler), project managers, Lisa Ka'aihue and Joe Banta, our many prior and current SAC advisors,  
387 and the every-ready folks on RCAC staff...and our families...for the years of Alaskan adventures and productive  
388 science--BIG THANKS!

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431 **APPENDICES**

432 Appendix 1 – PAH and SHC Plots

433 Appendix 2 – Biomarker Plots

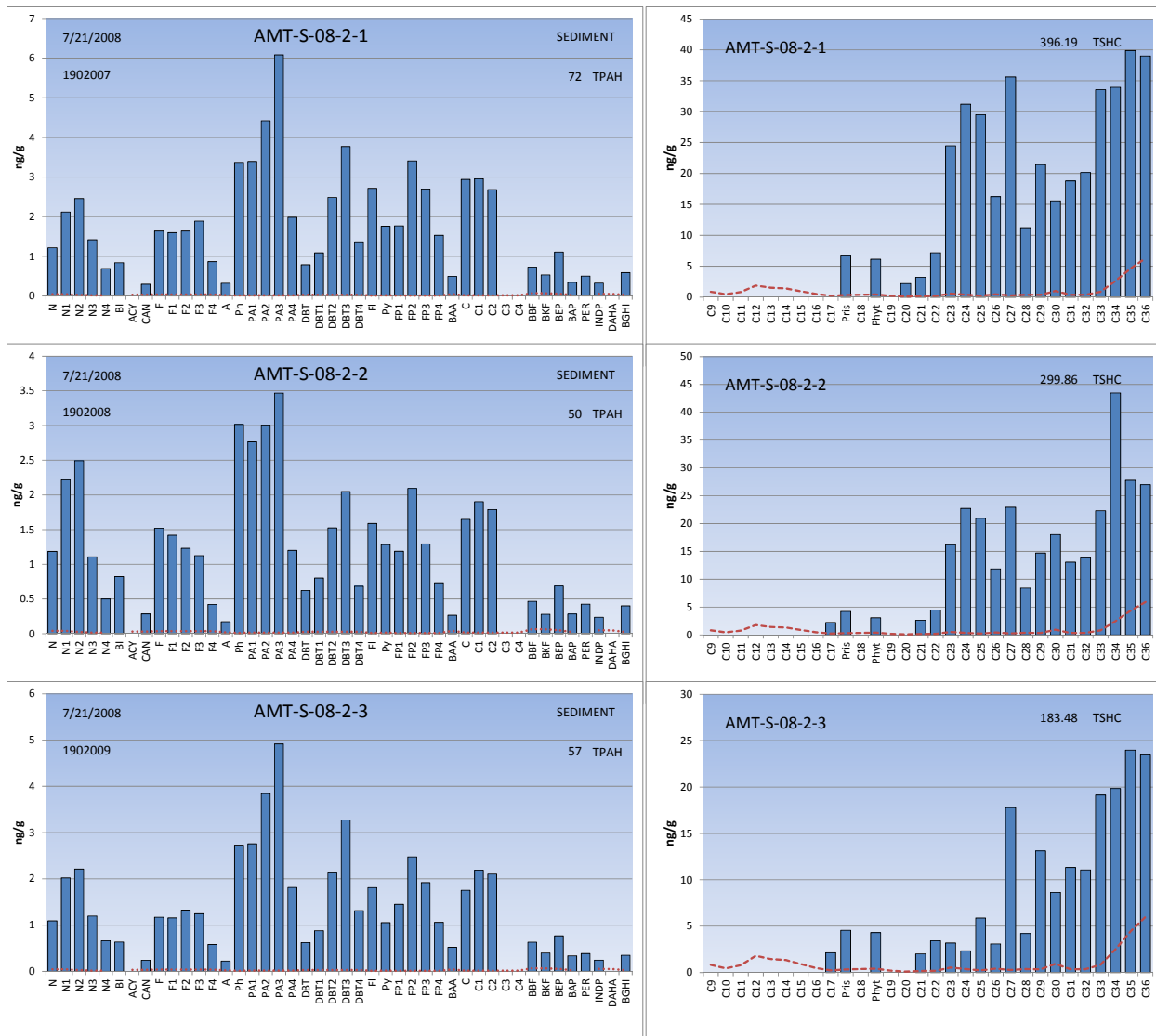
434 Appendix 3 – LTEMP PAH interpretation using multivariate scoring methods

435 APPENDIX 1 – PAH AND SHC PLOTS

436 This appendix contains plots for all 2008-2012 LTEMP samples. Sample ID code comprises station-matrix-year-season-rep, e.g.,  
 437 AMT-S-12-2-1 means Alyeska Marine Terminal, Sediment, 2012, summer (July), rep 1.

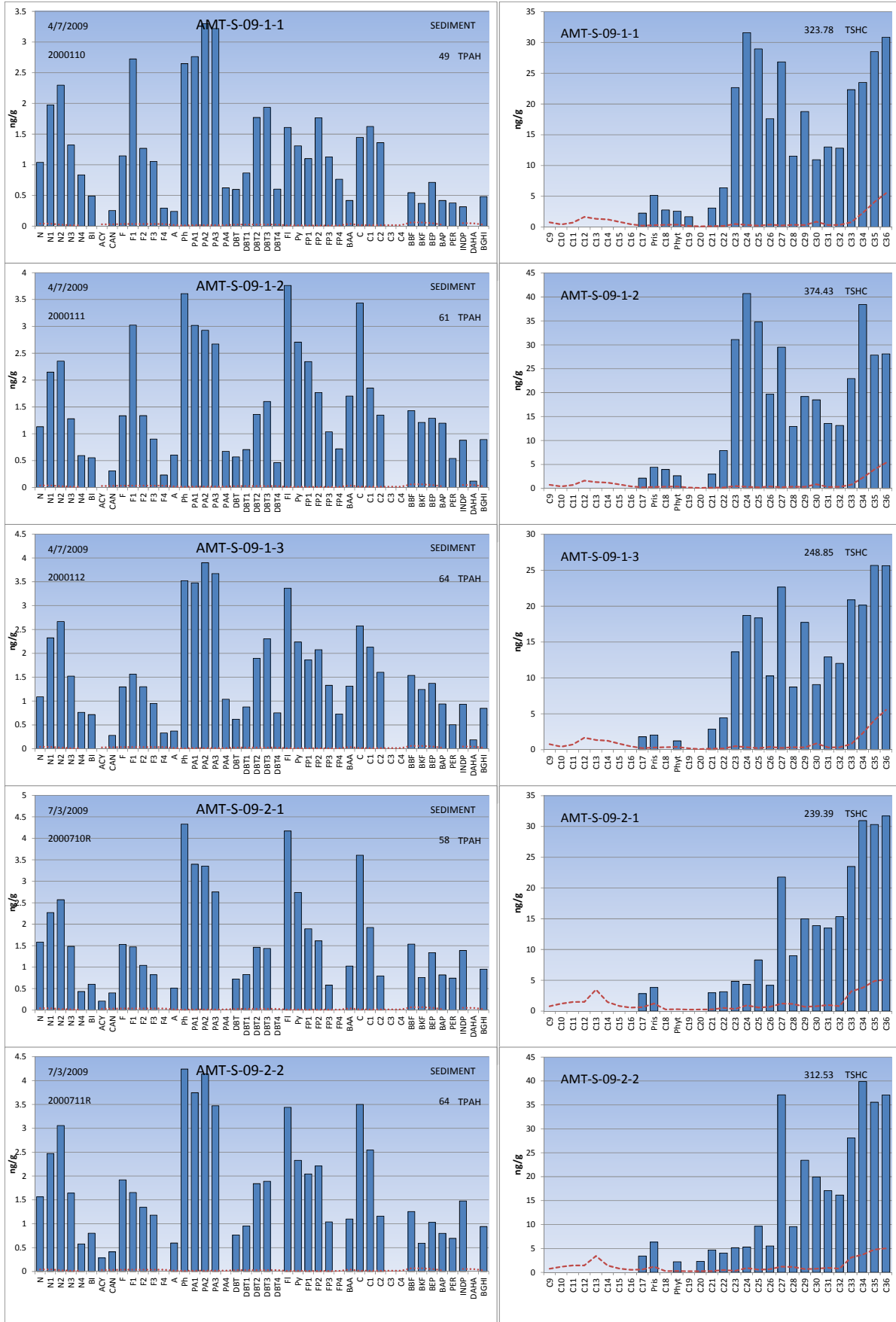
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439 ALYESKA MARINE TERMINAL-- AMT

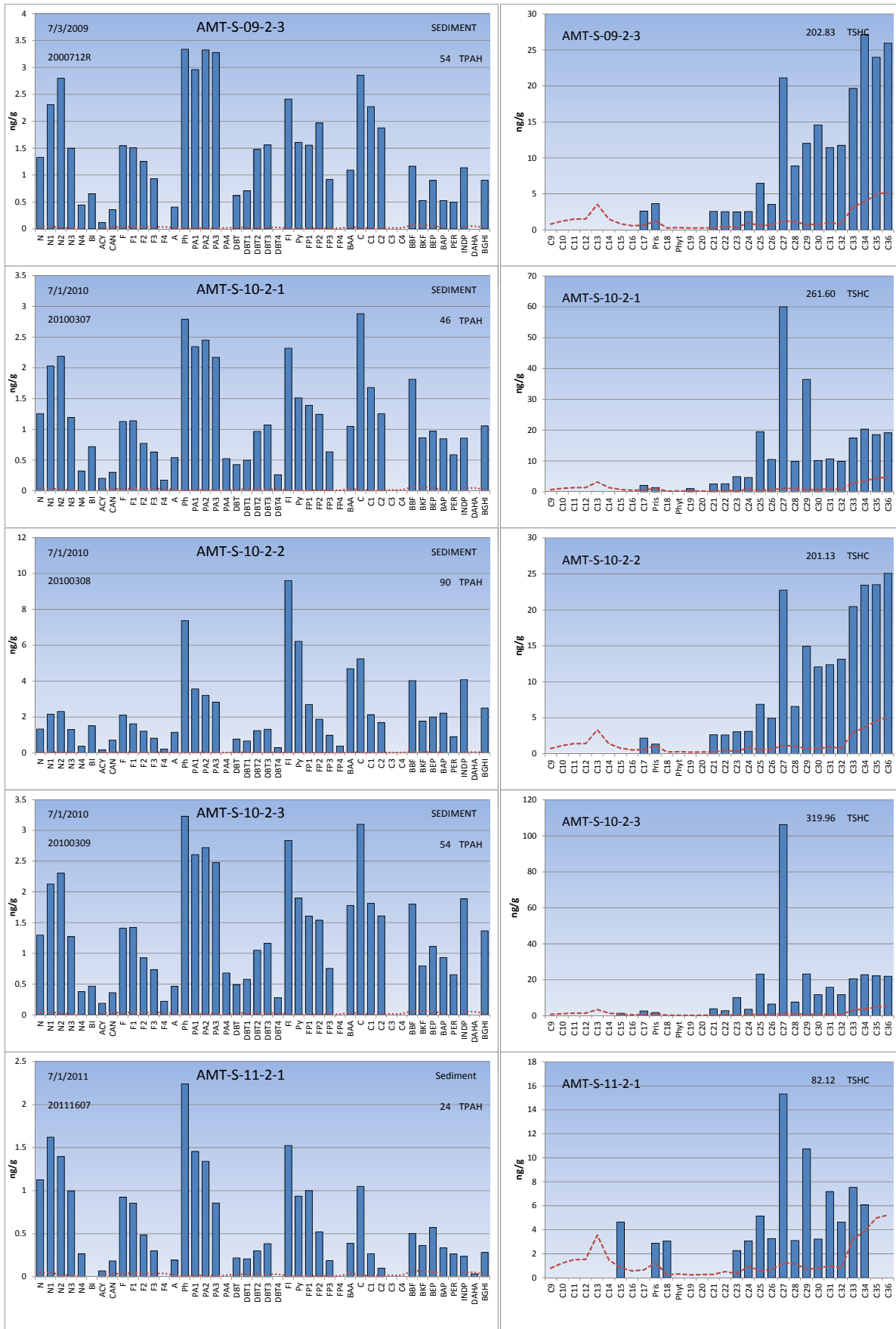


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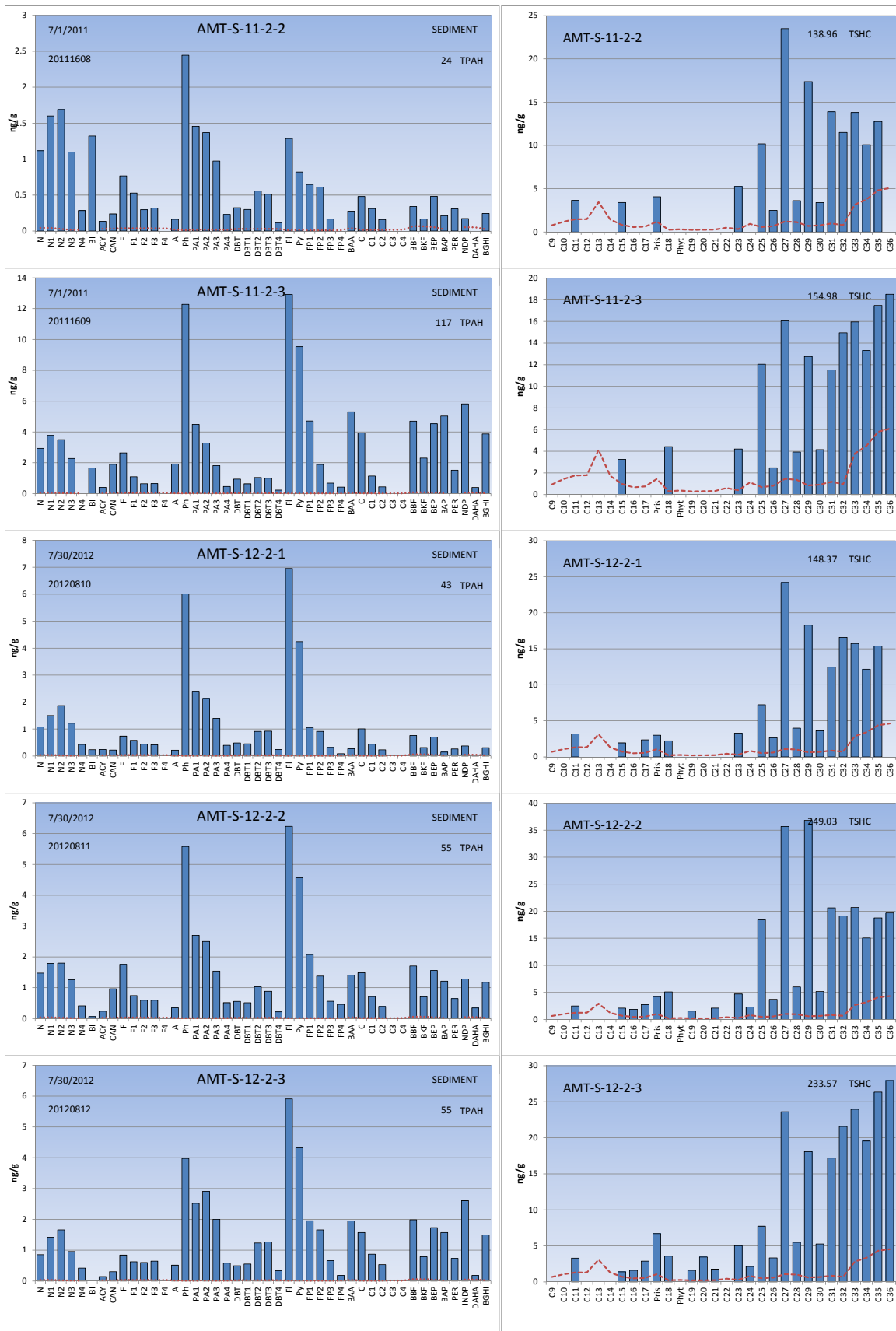
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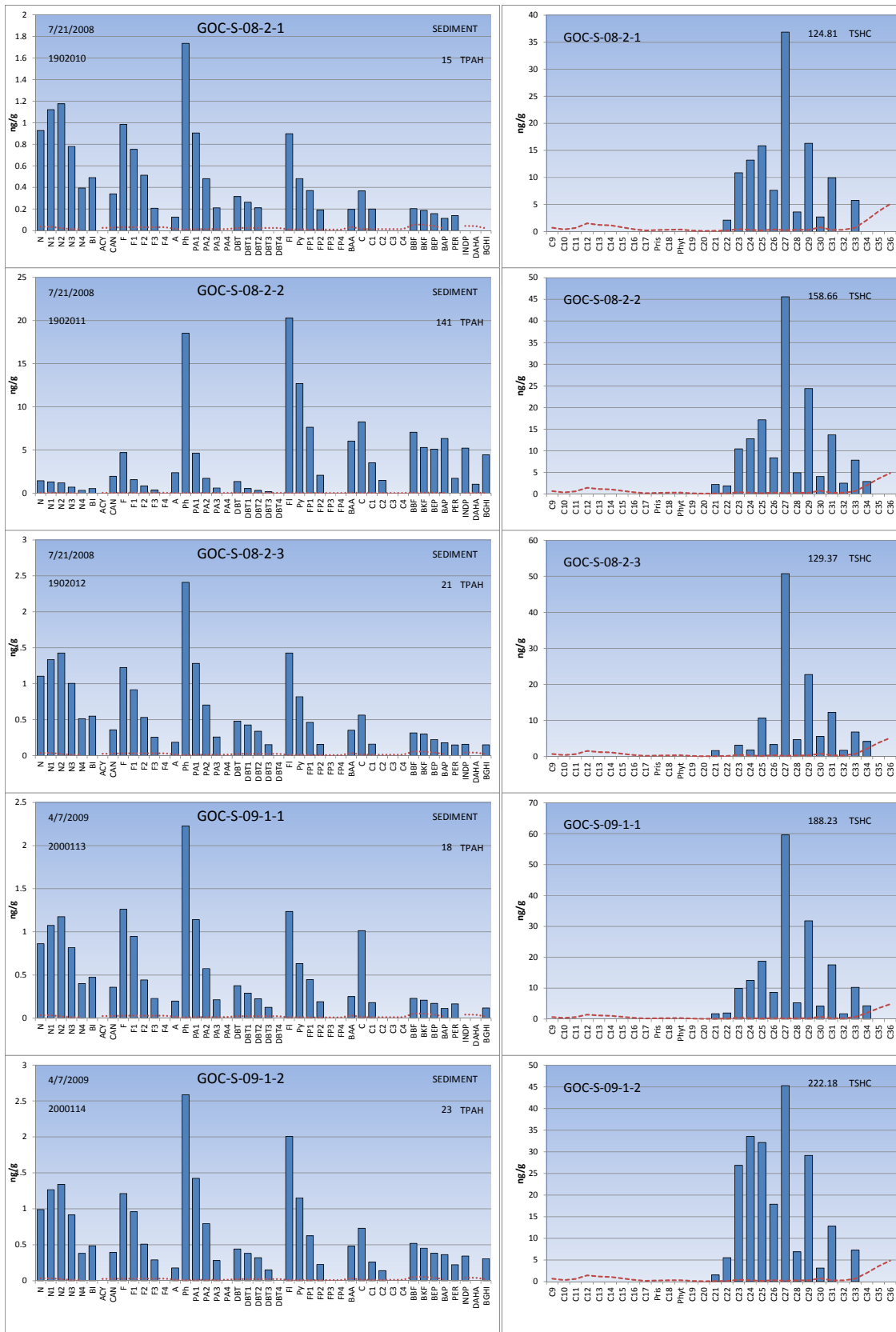


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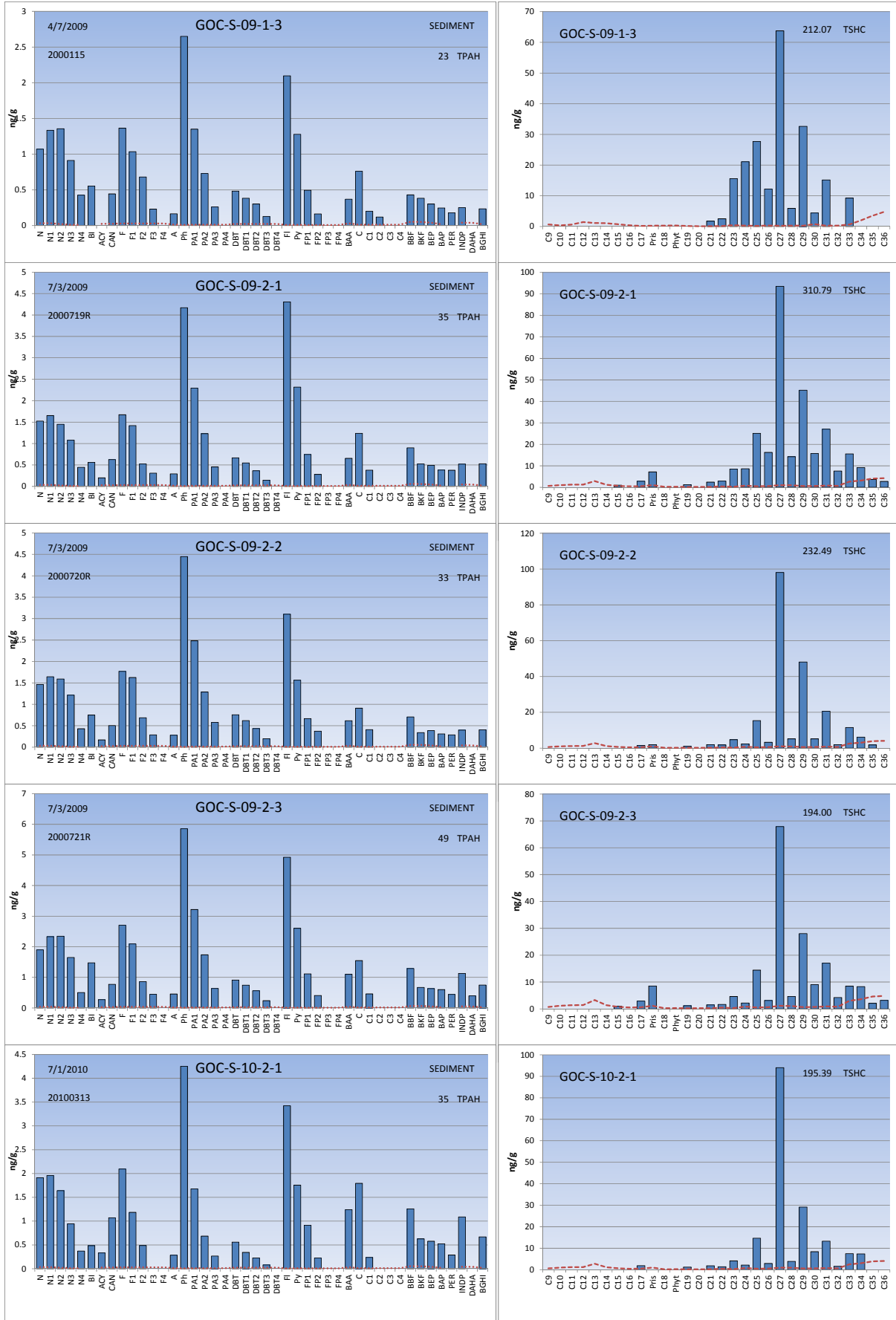




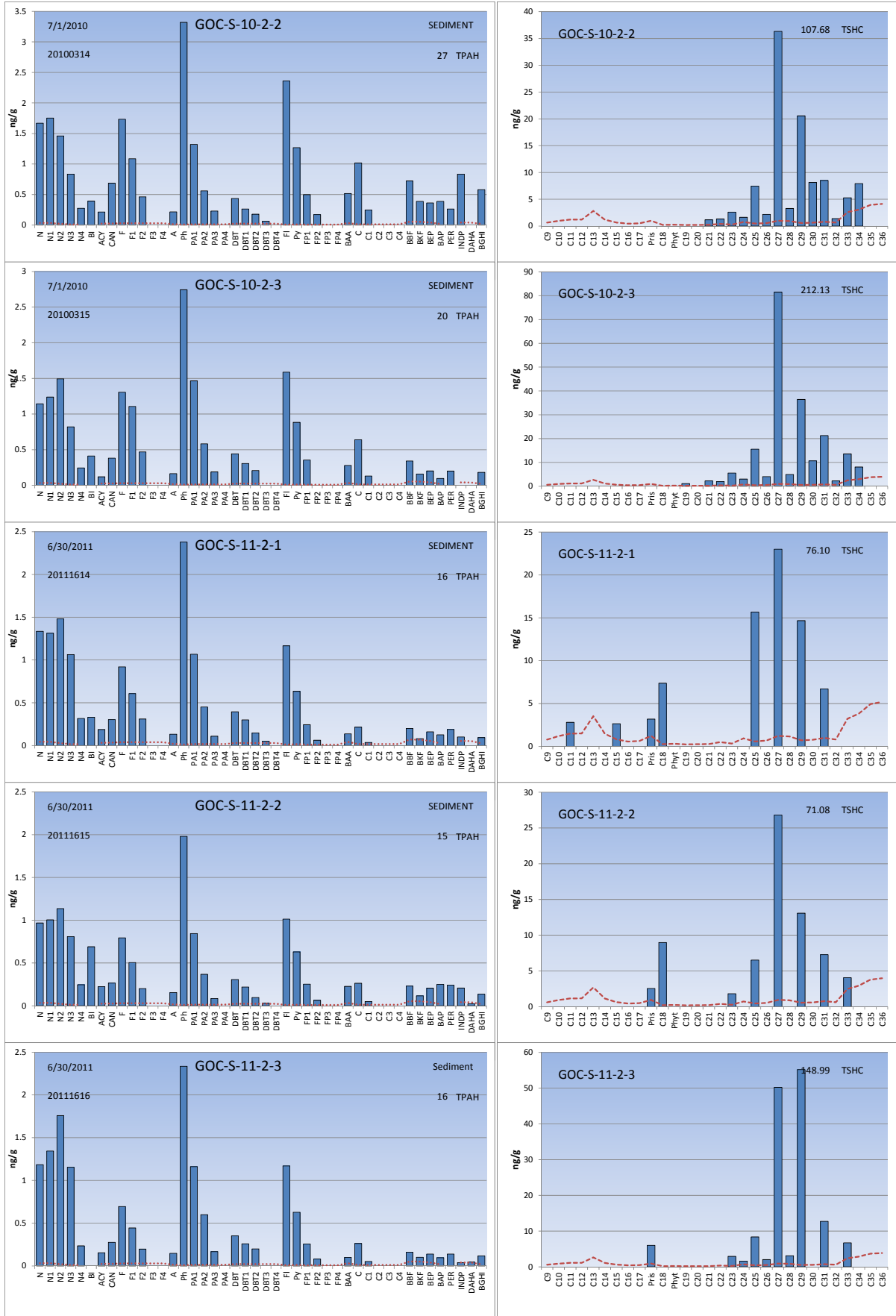
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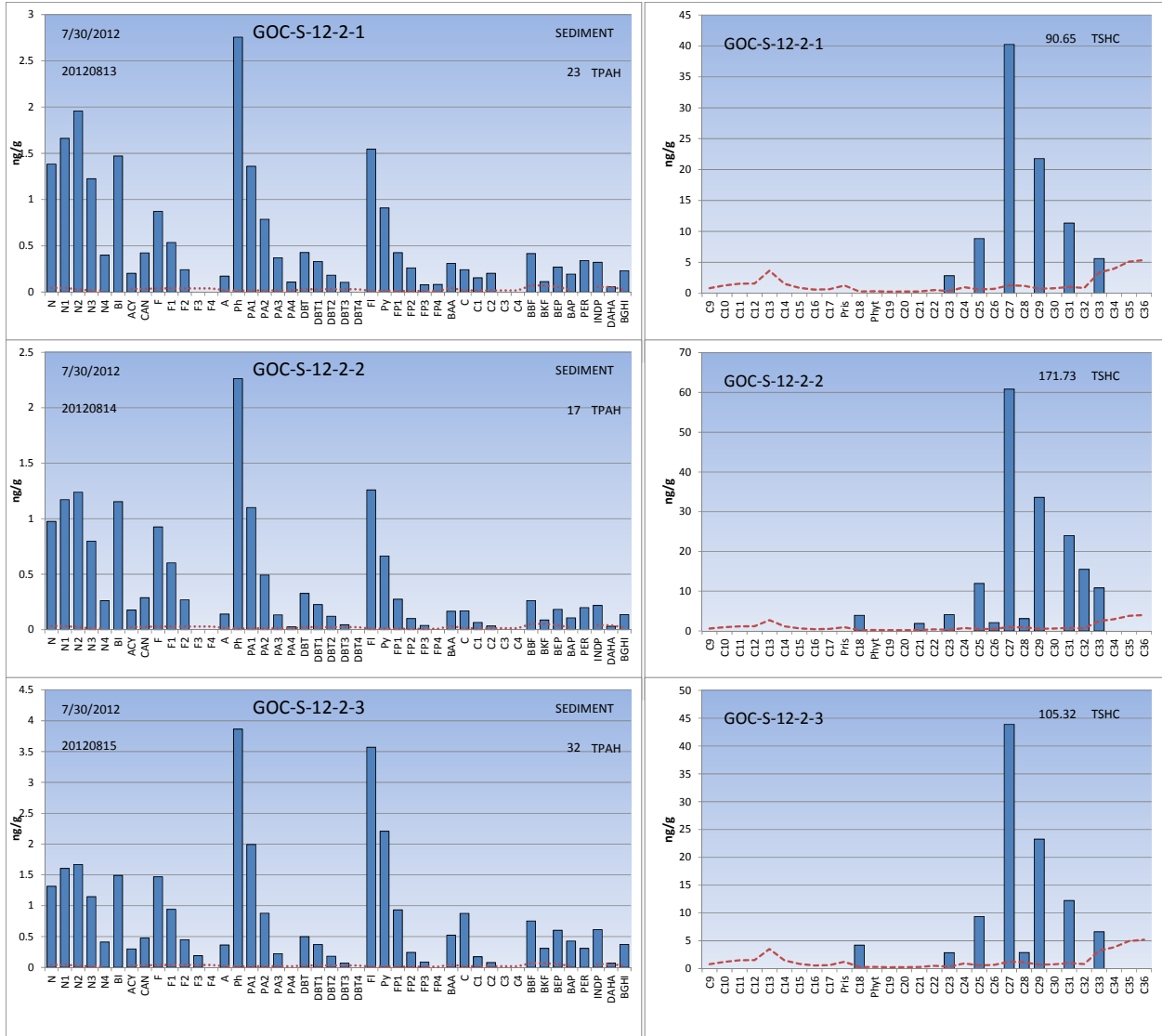
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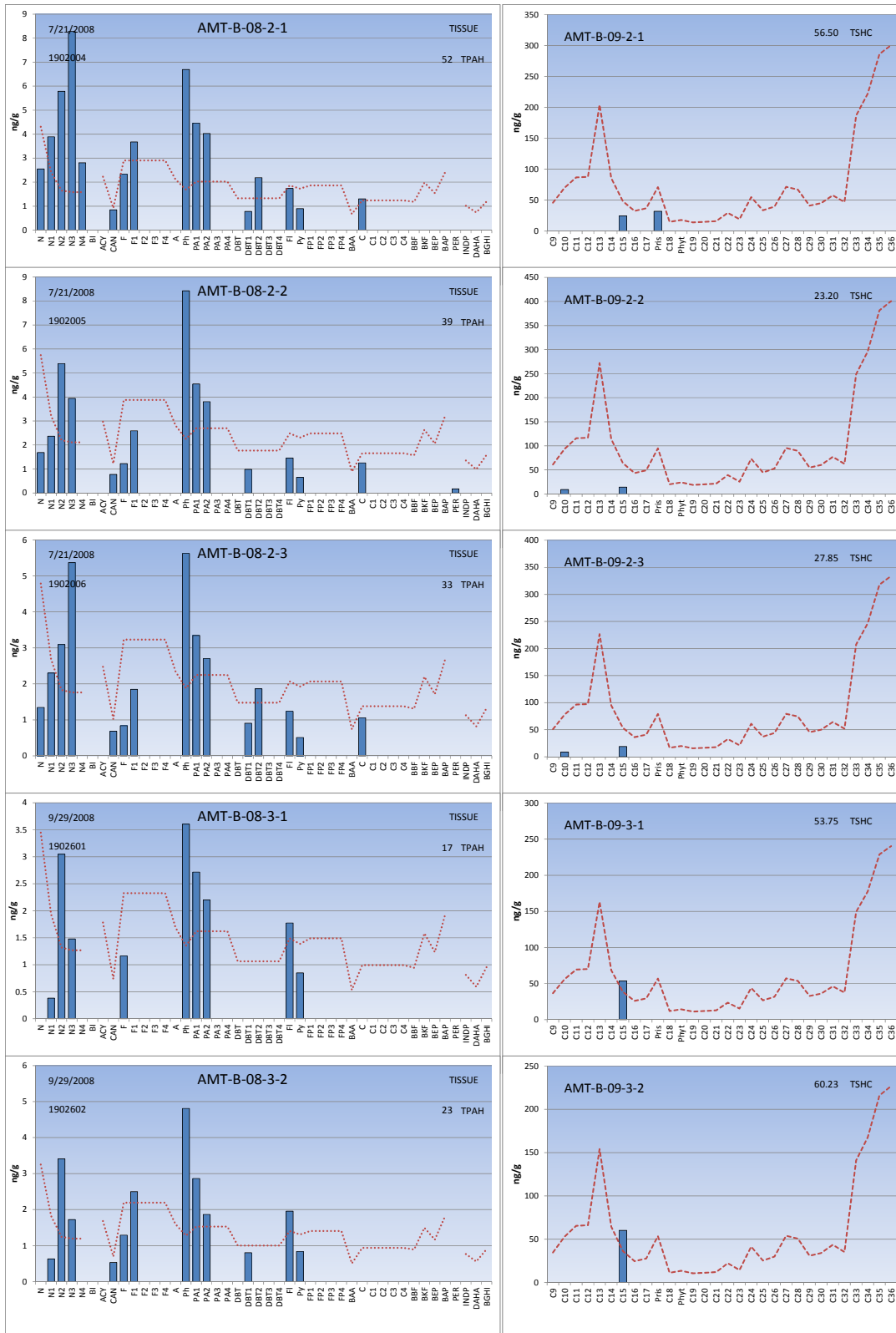
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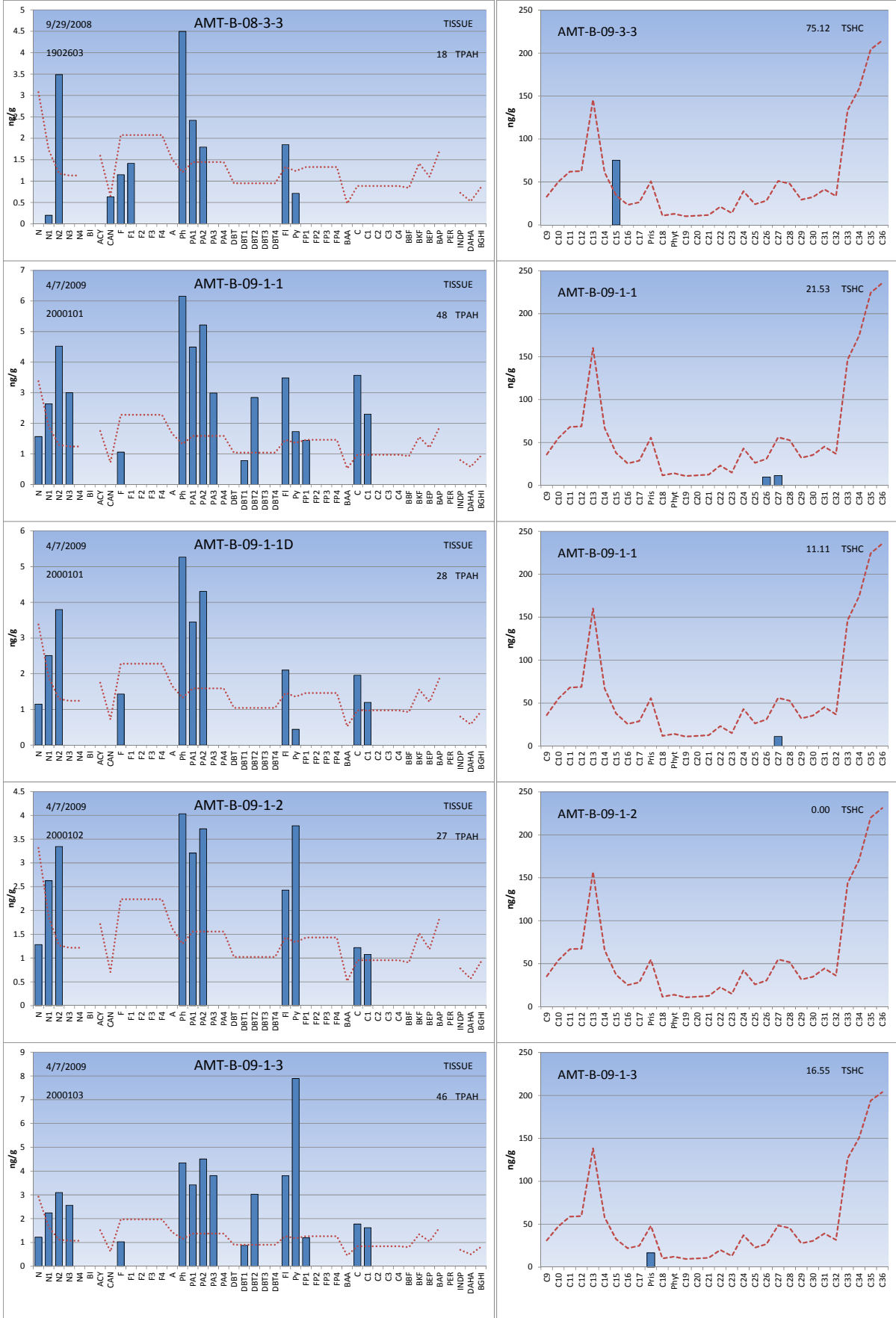
449 TISSUES

450 ALYESKA MARINE TERMINAL--AMT

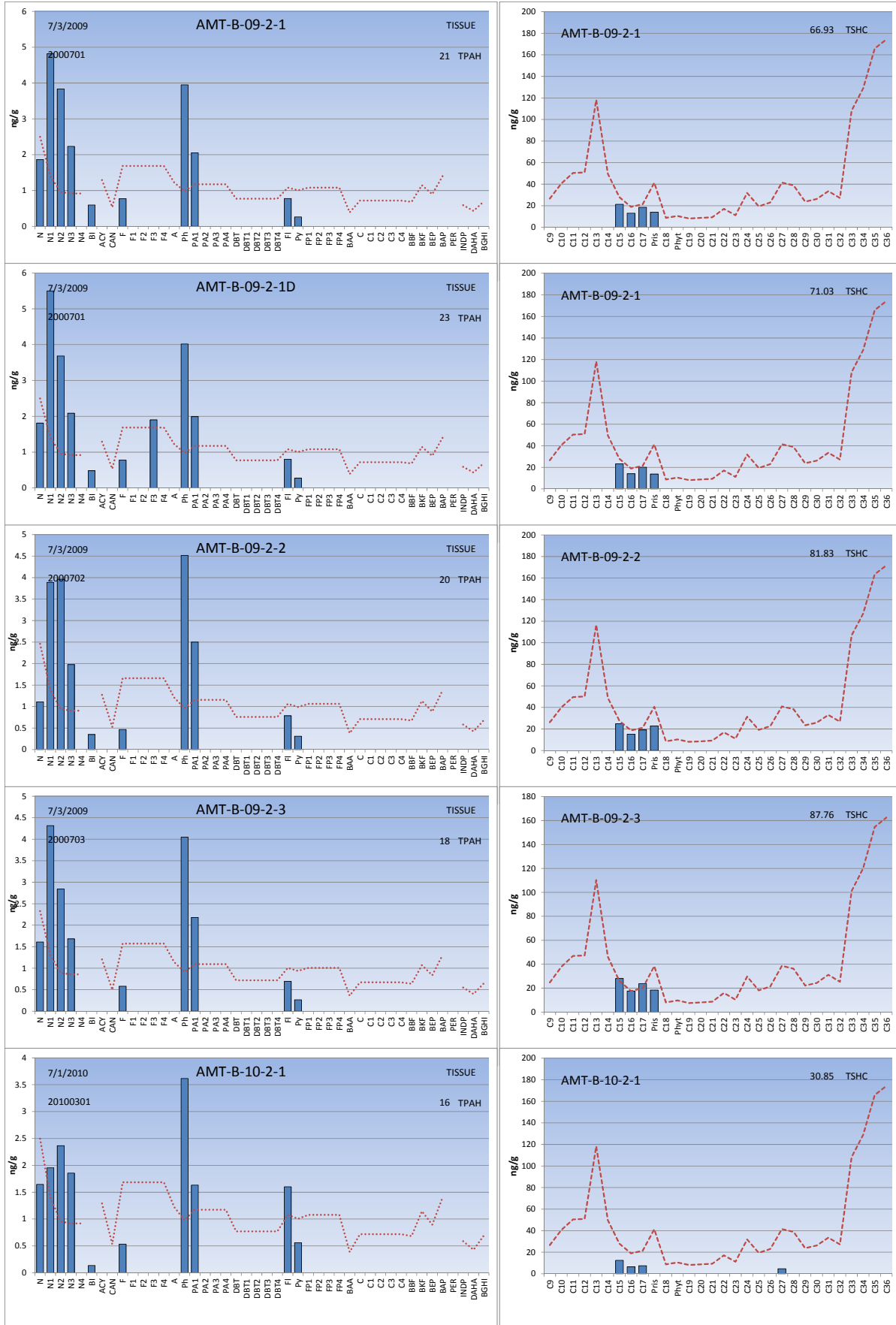


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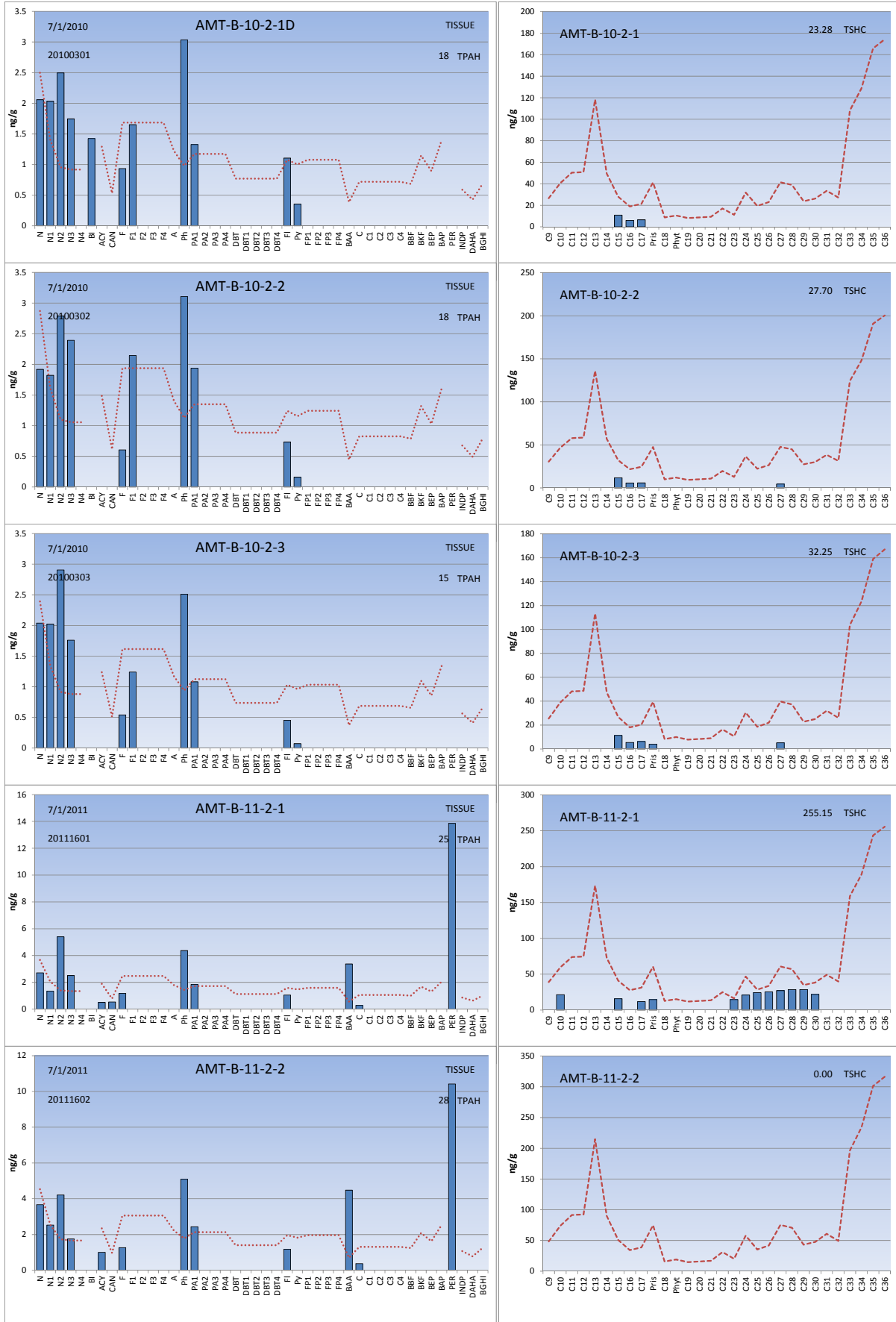
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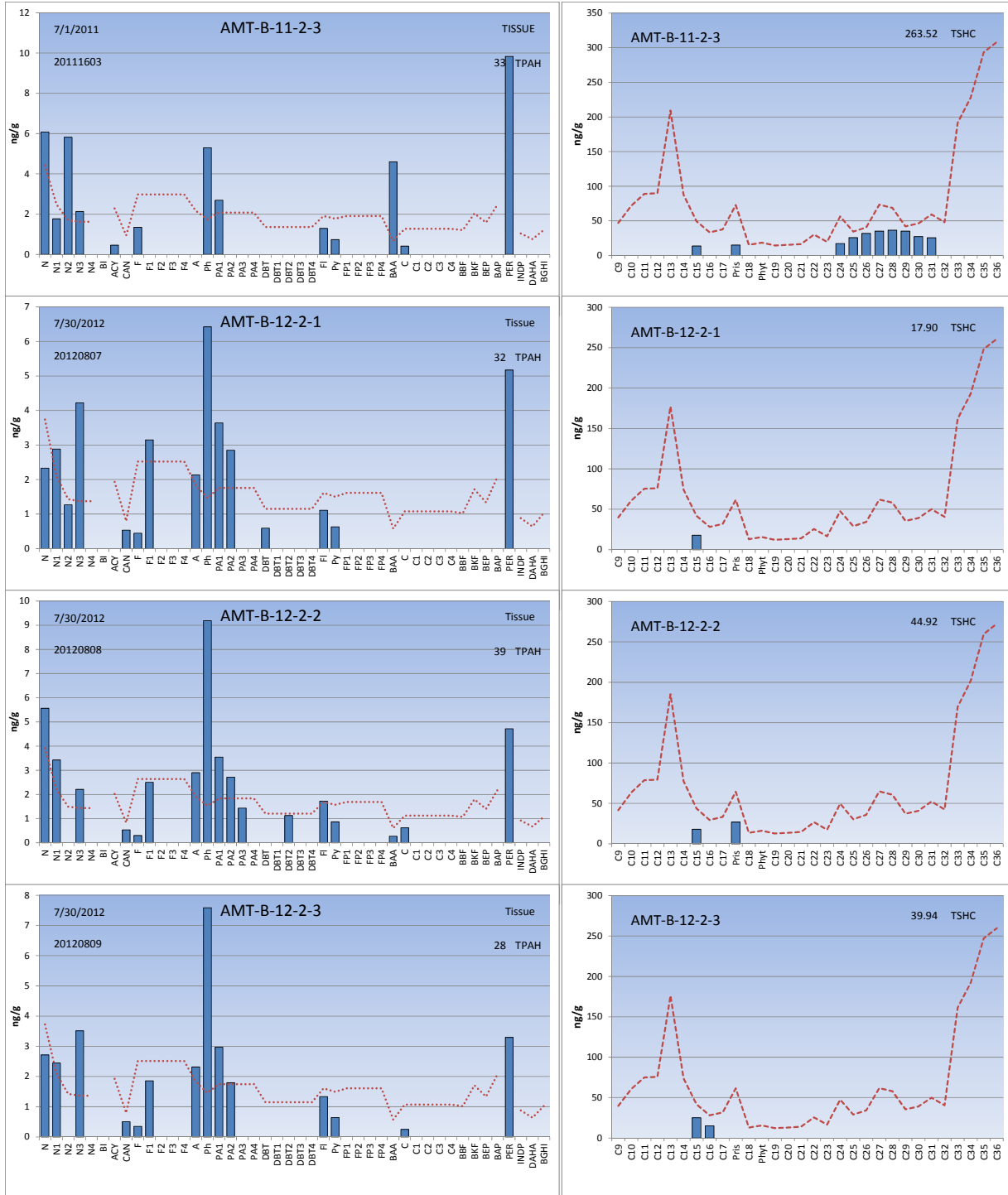


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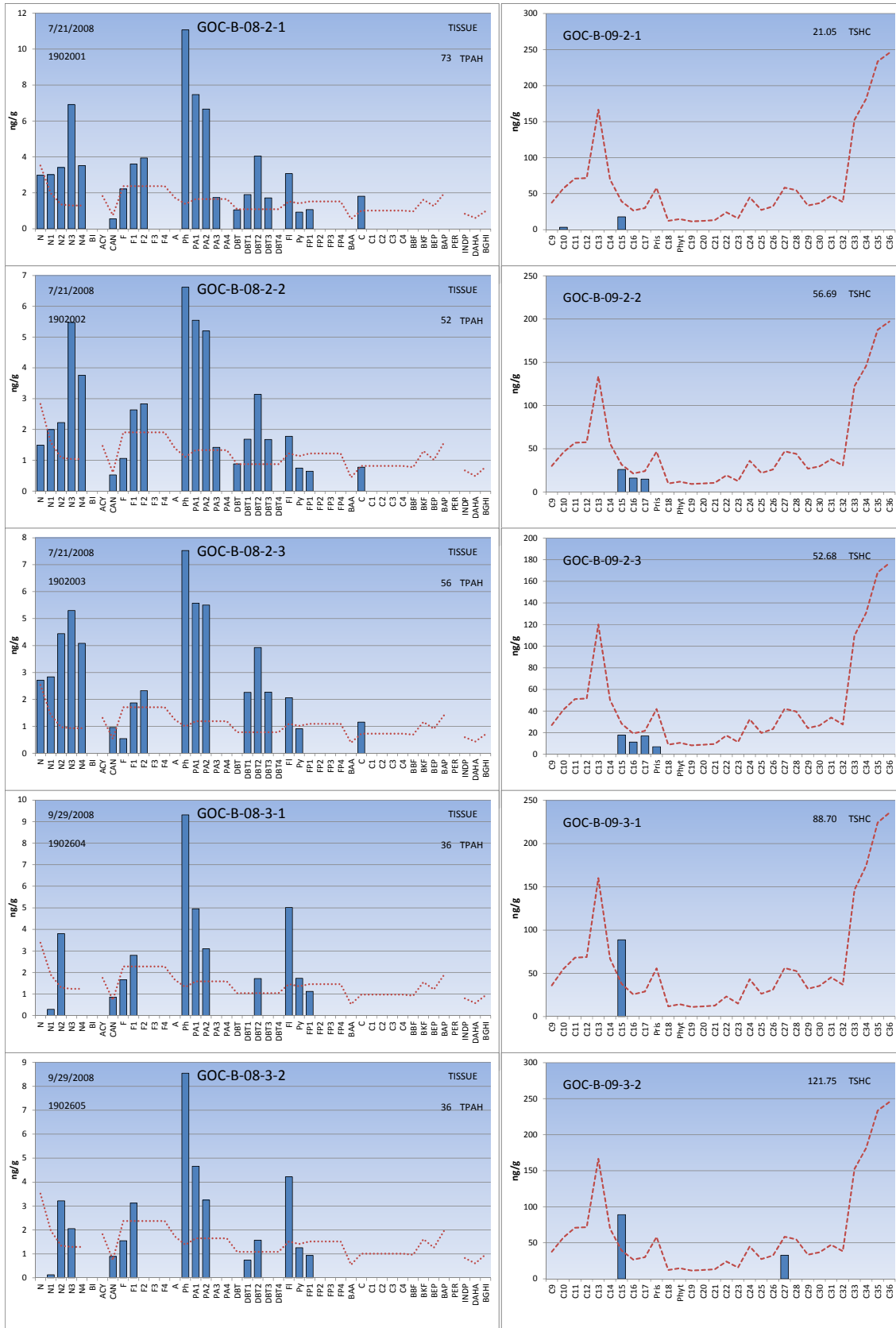




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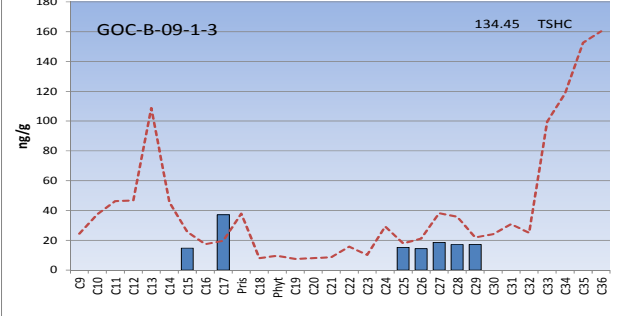
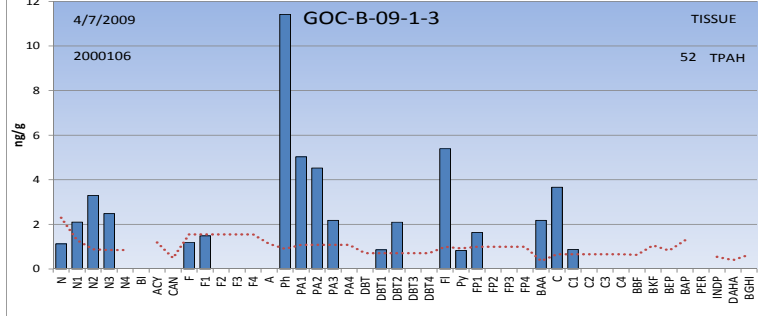
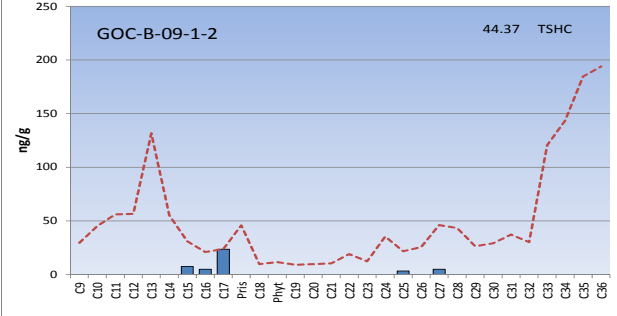
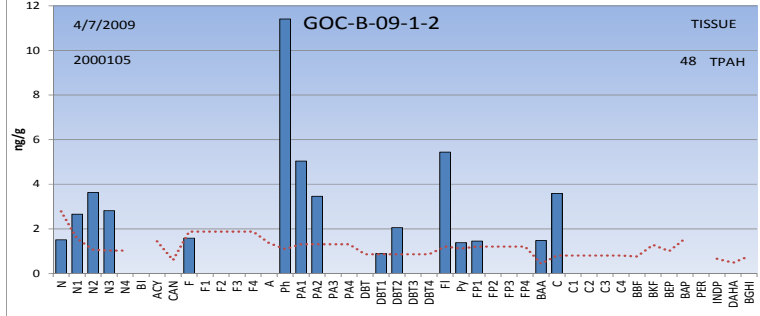
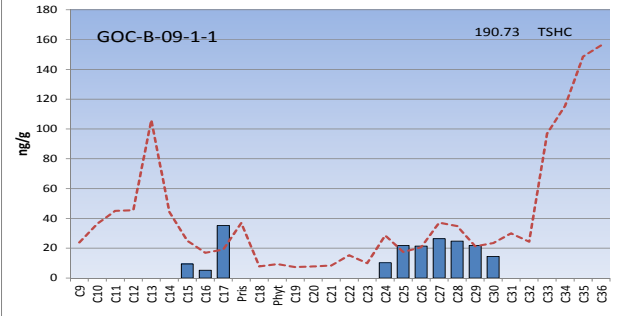
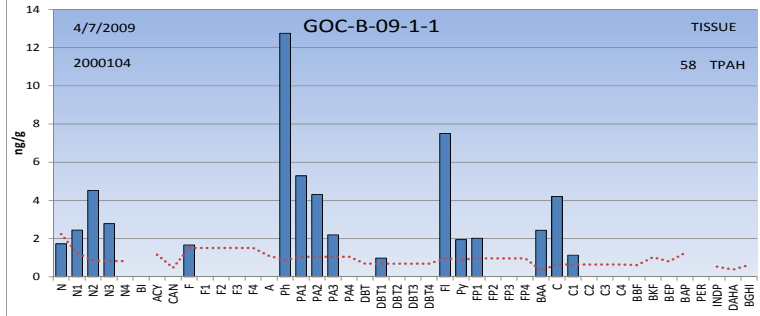
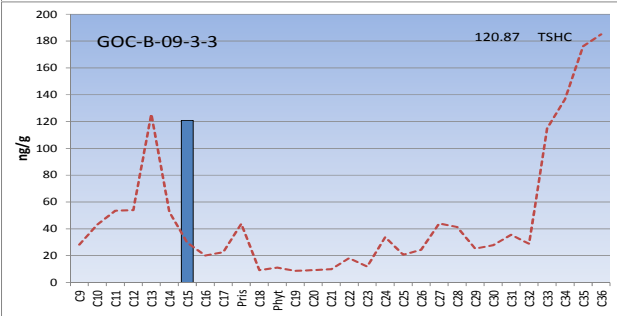
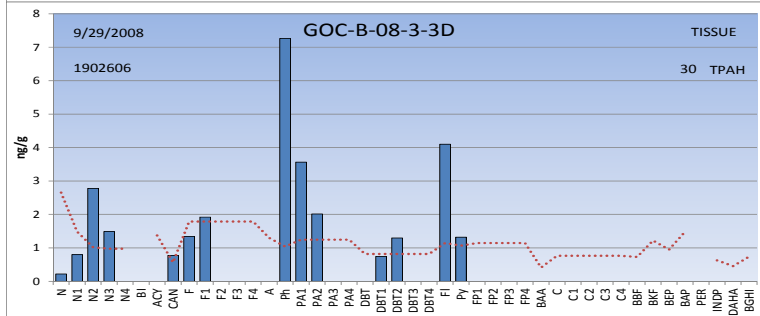
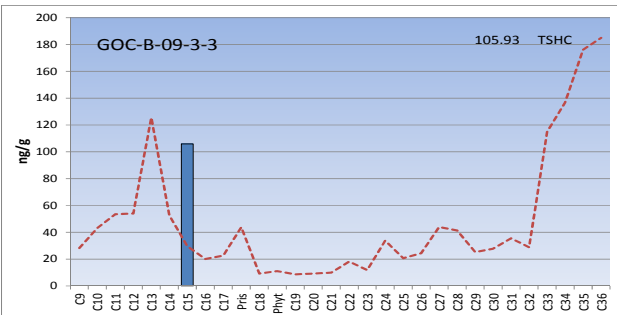
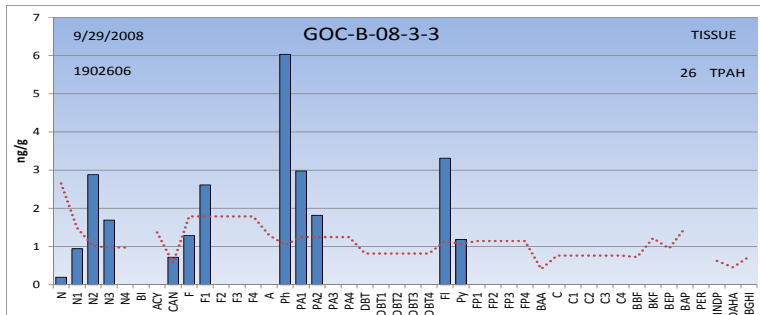


456 GOLD CREEK--GOC

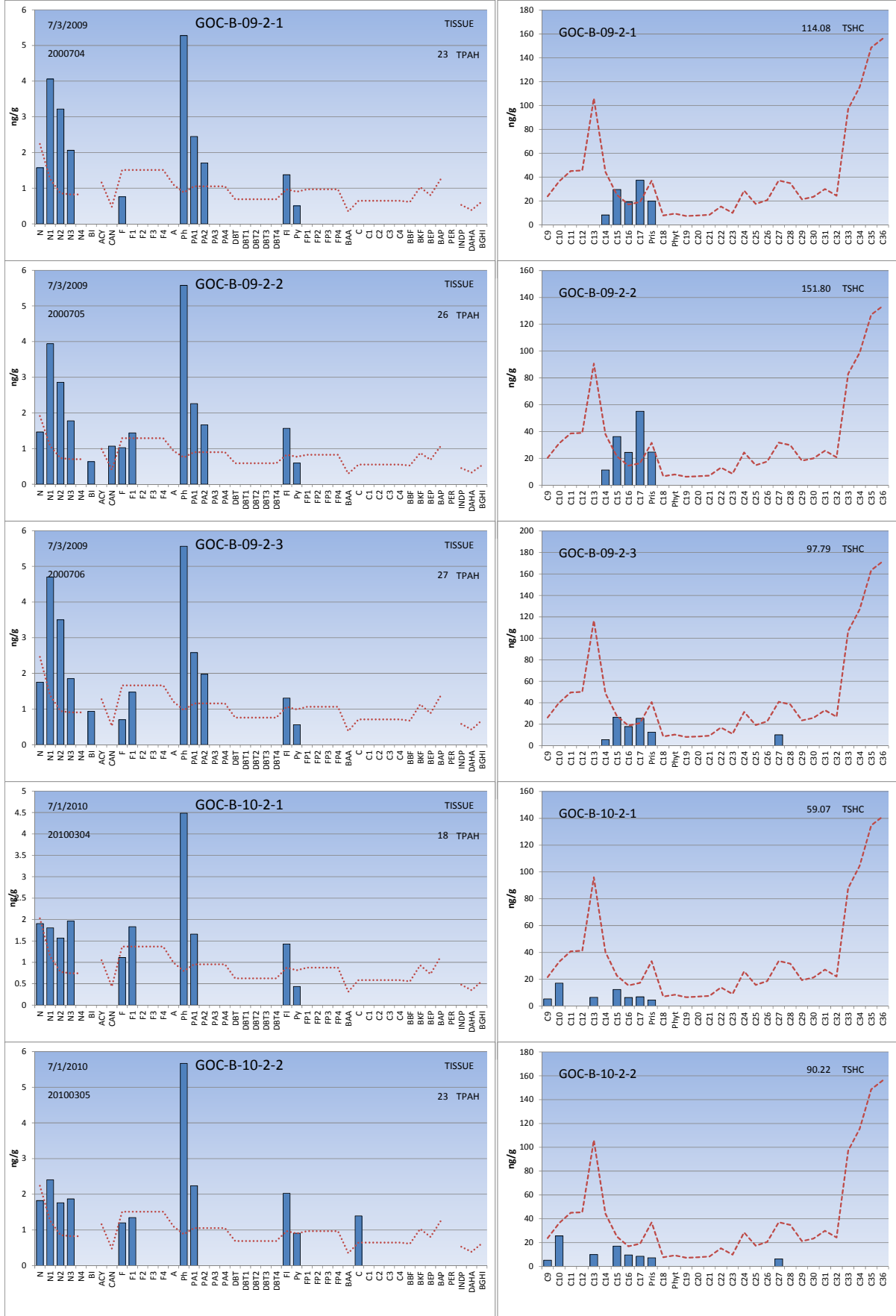


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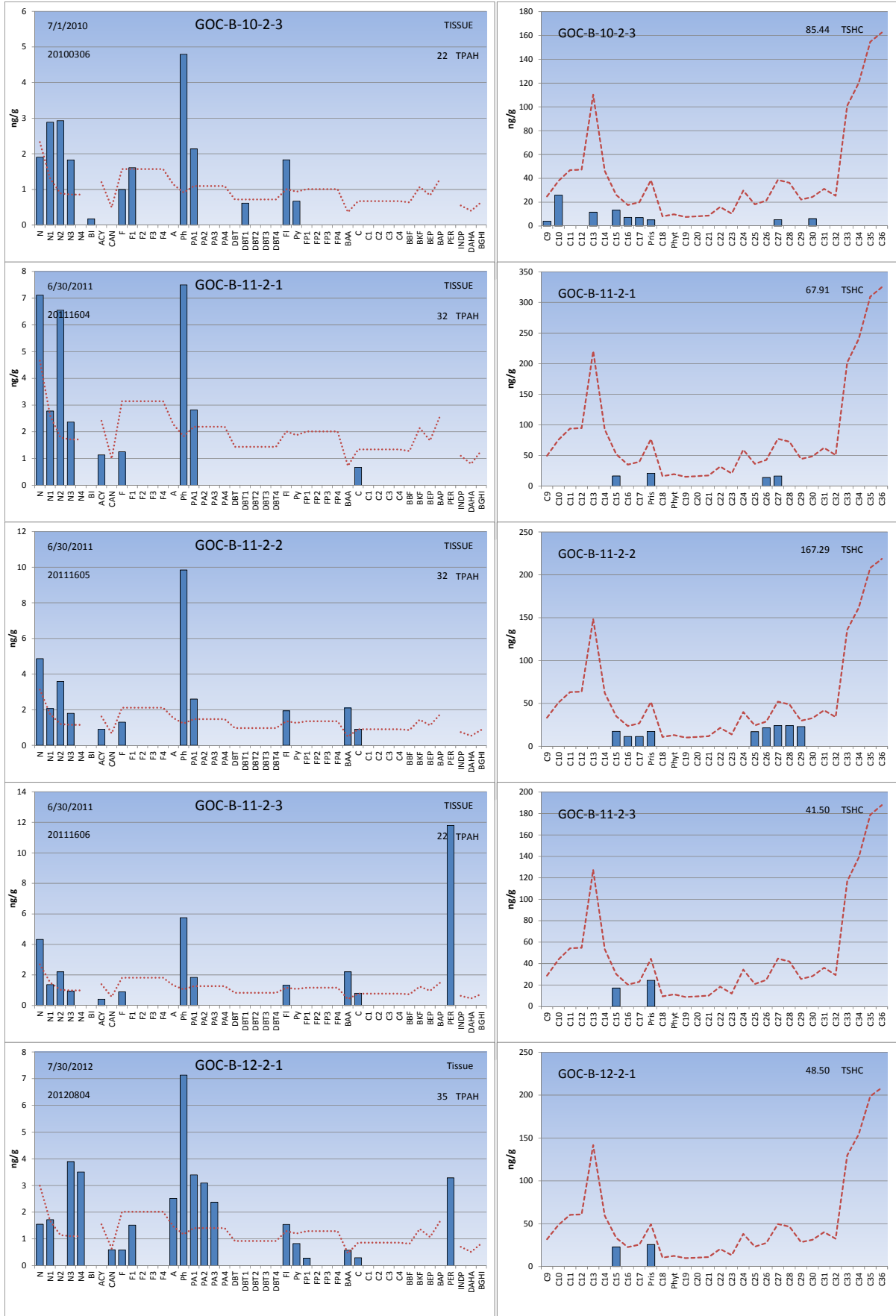
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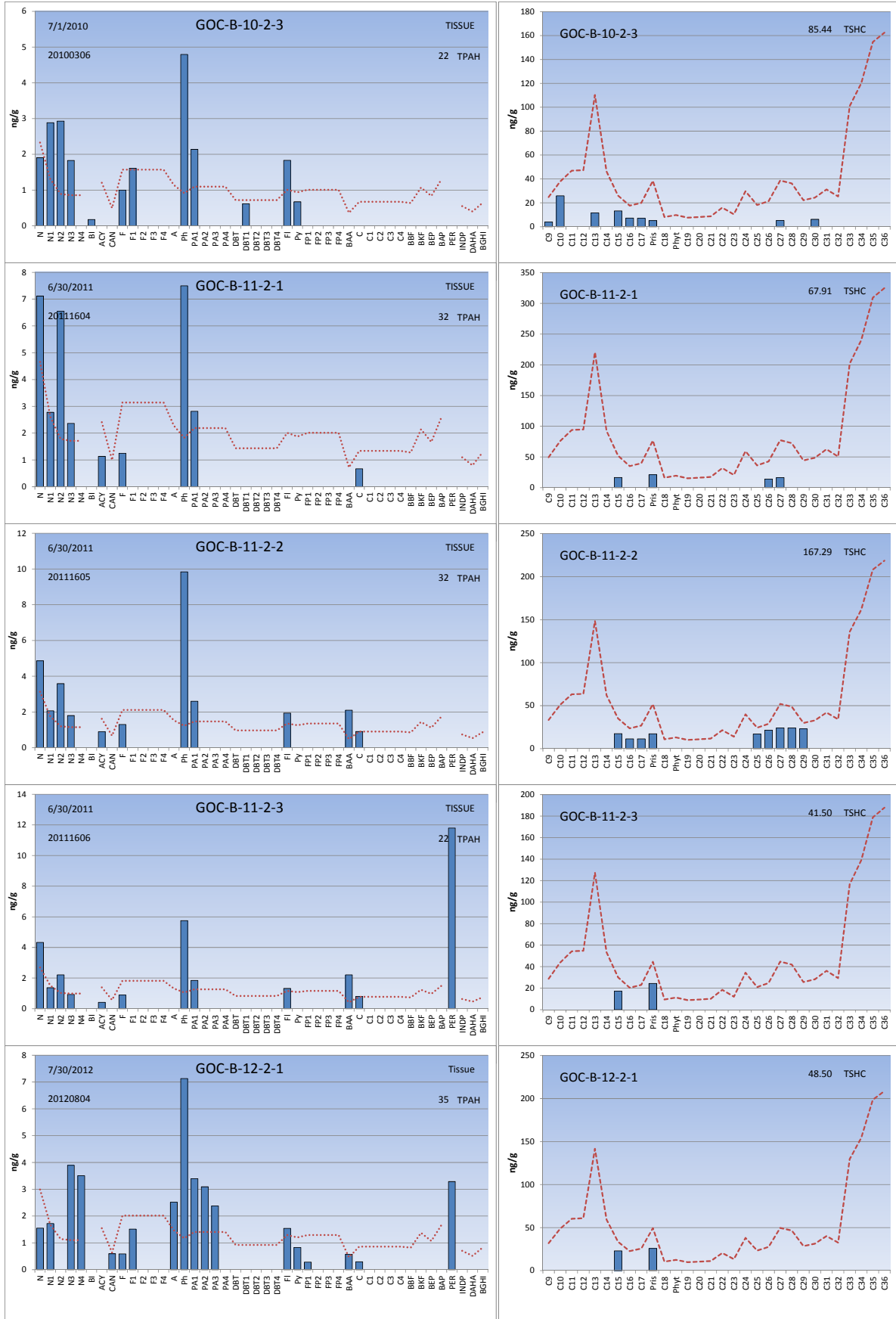
# LTEMP Report 2008-2012



# LTEMP Report 2008-2012

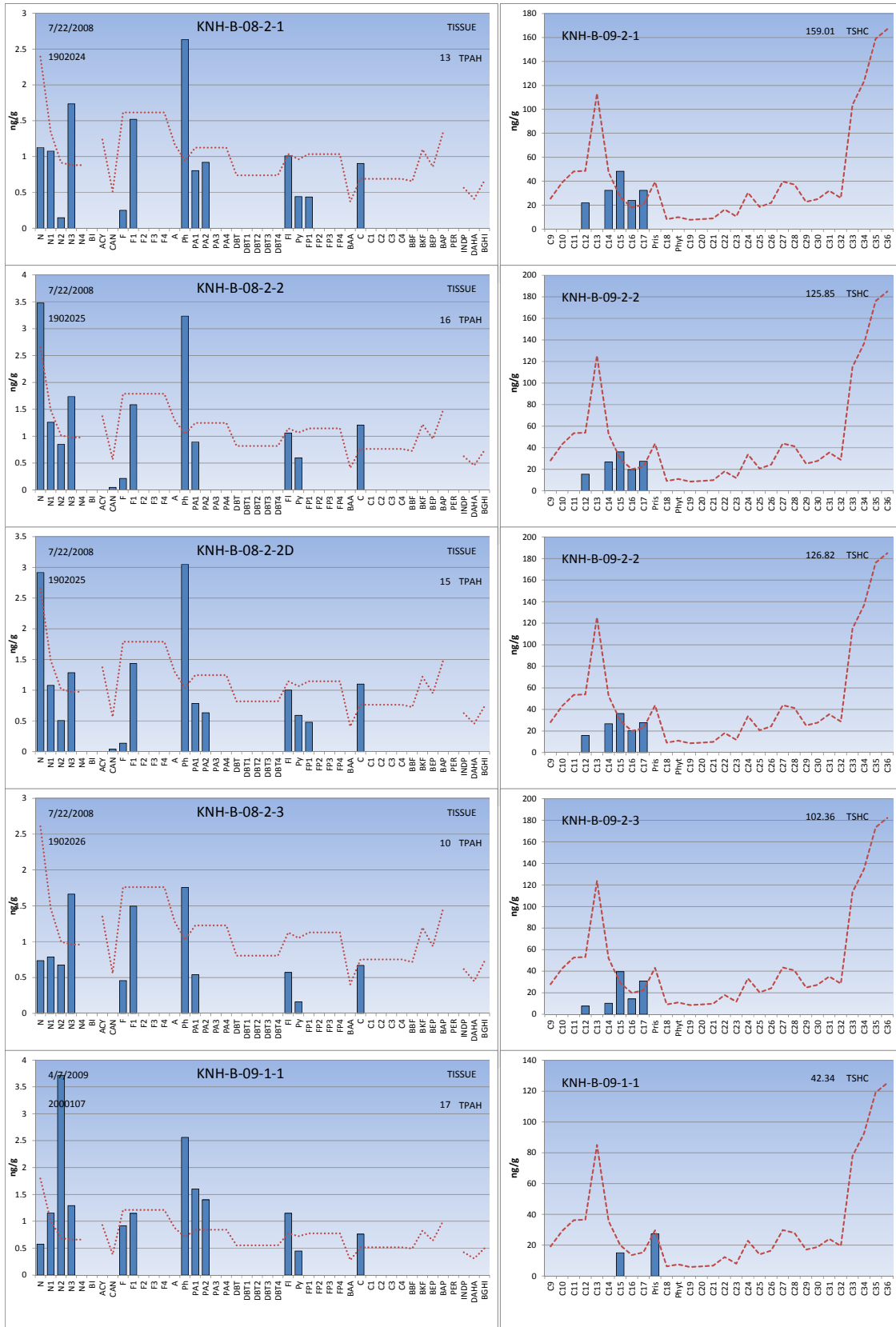


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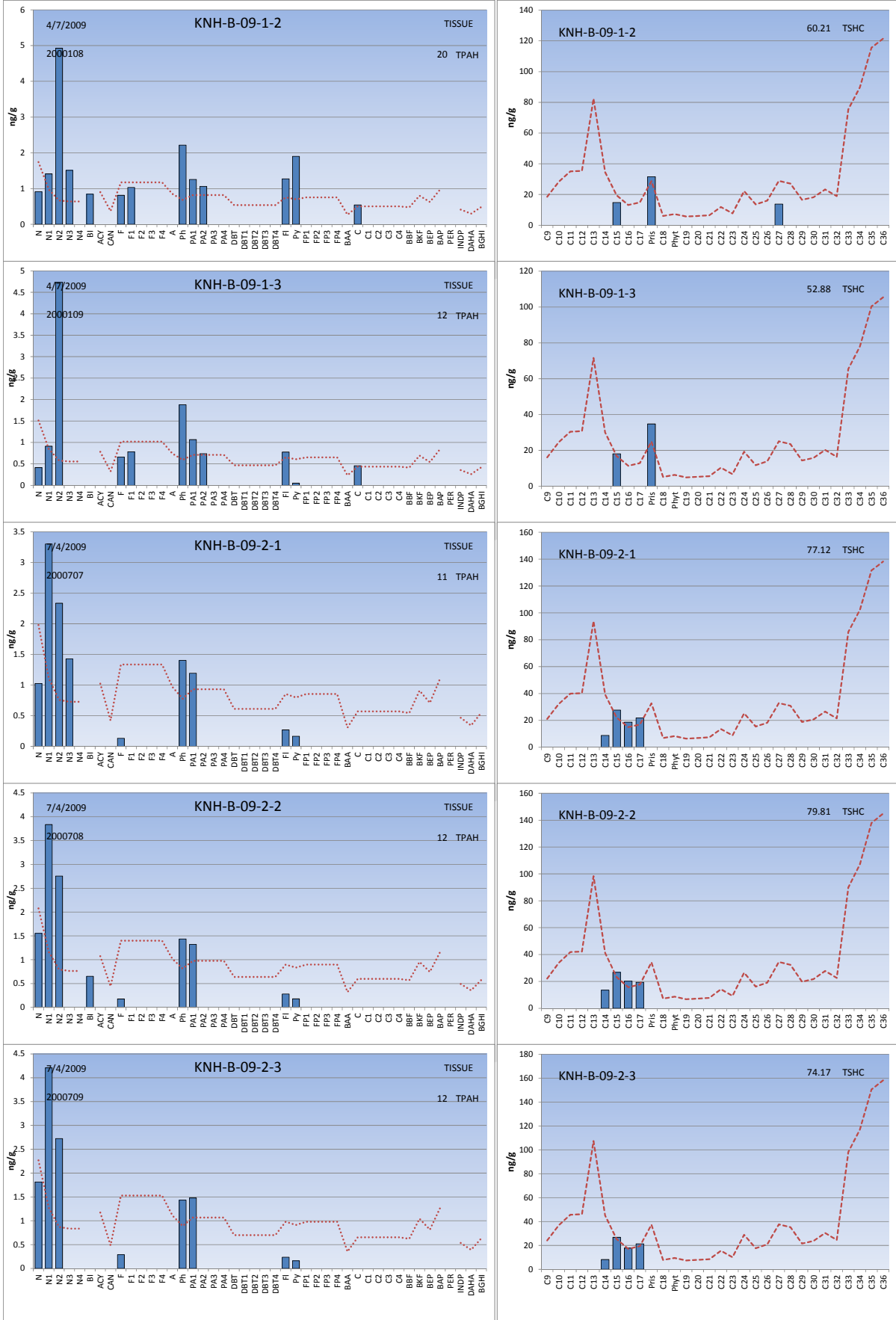
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462 KNOWLES HEAD--KNH



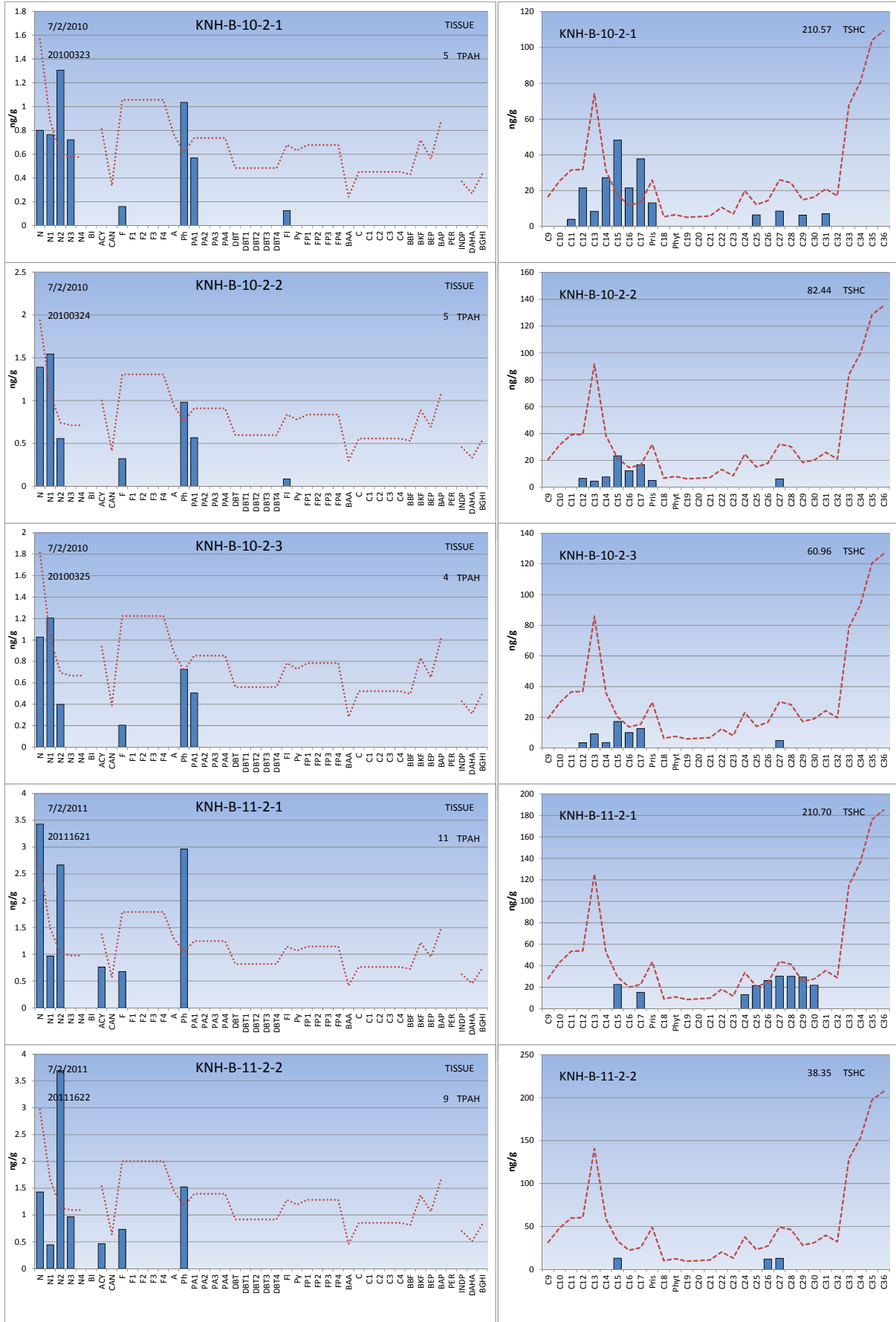
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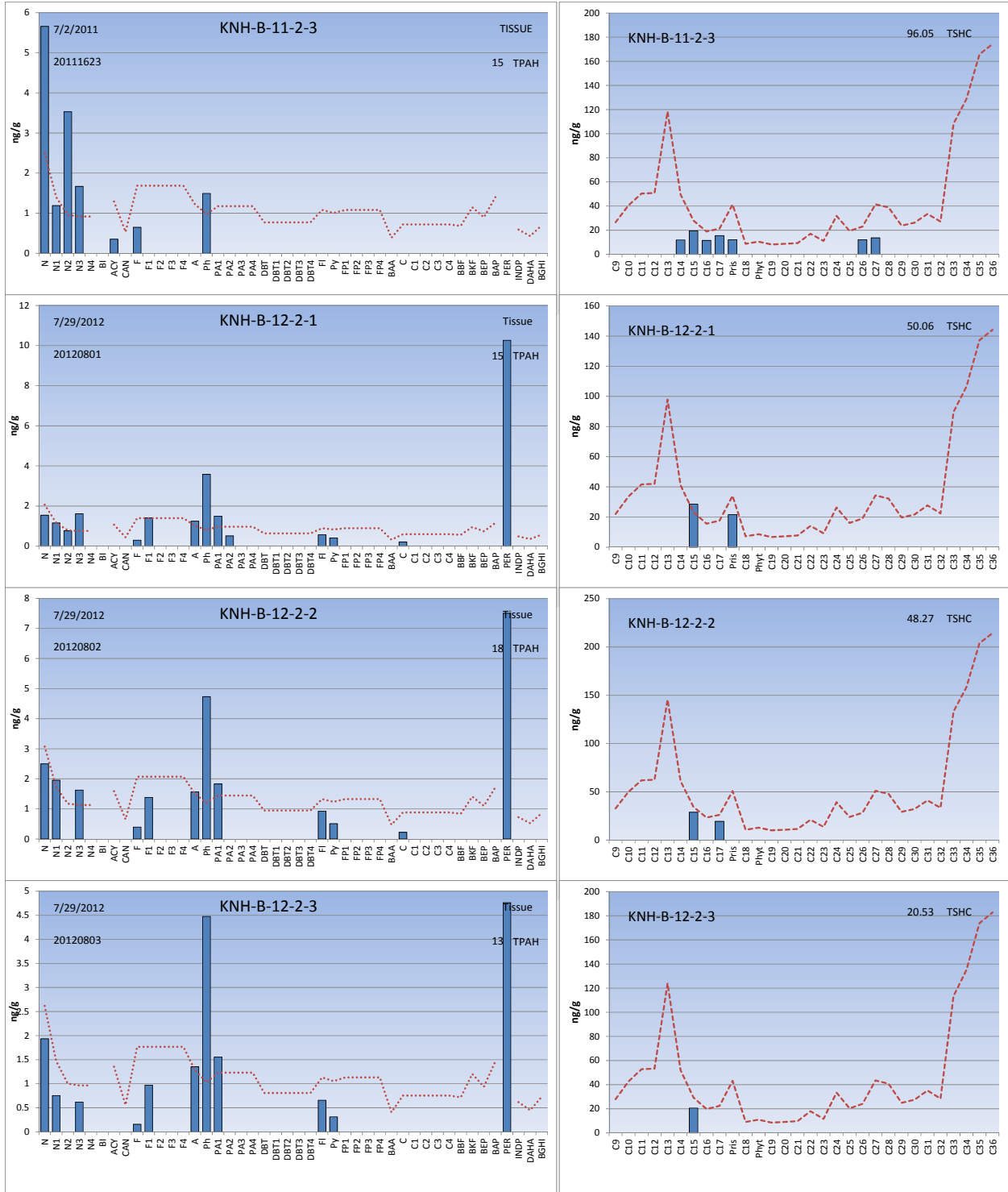




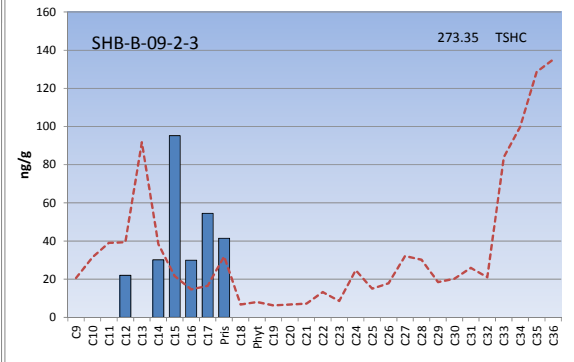
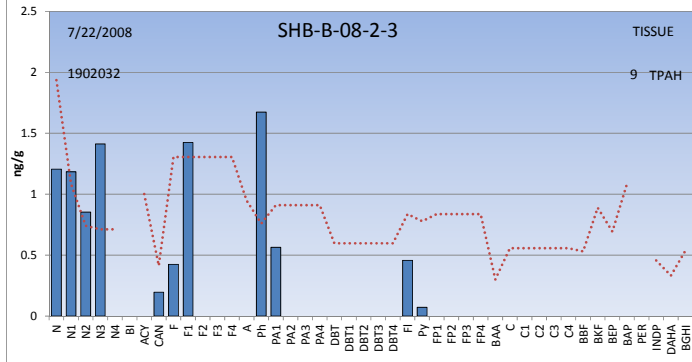
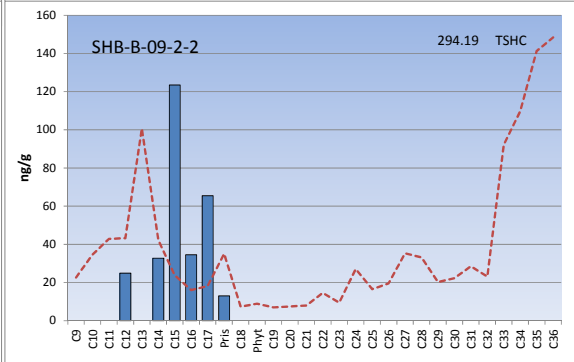
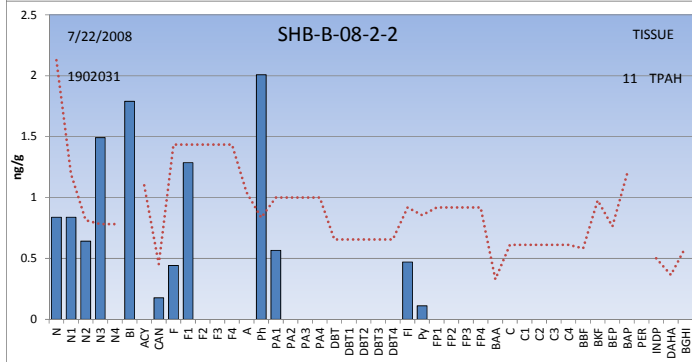
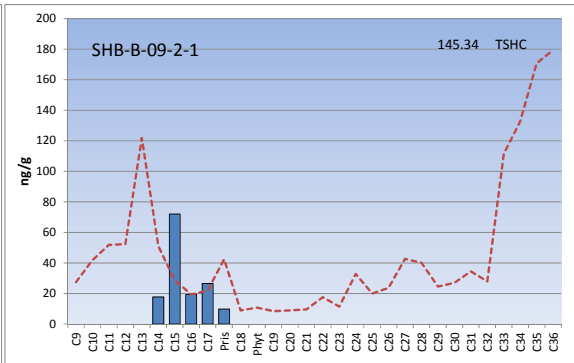
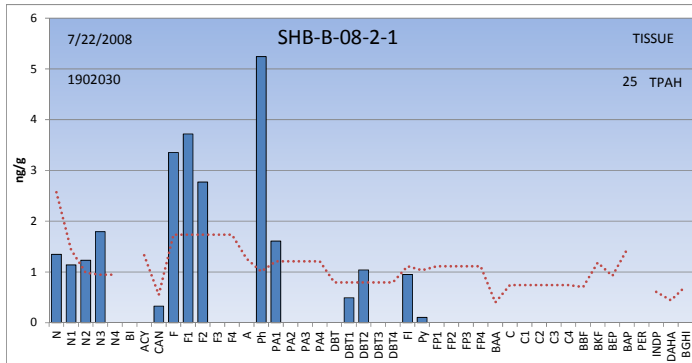
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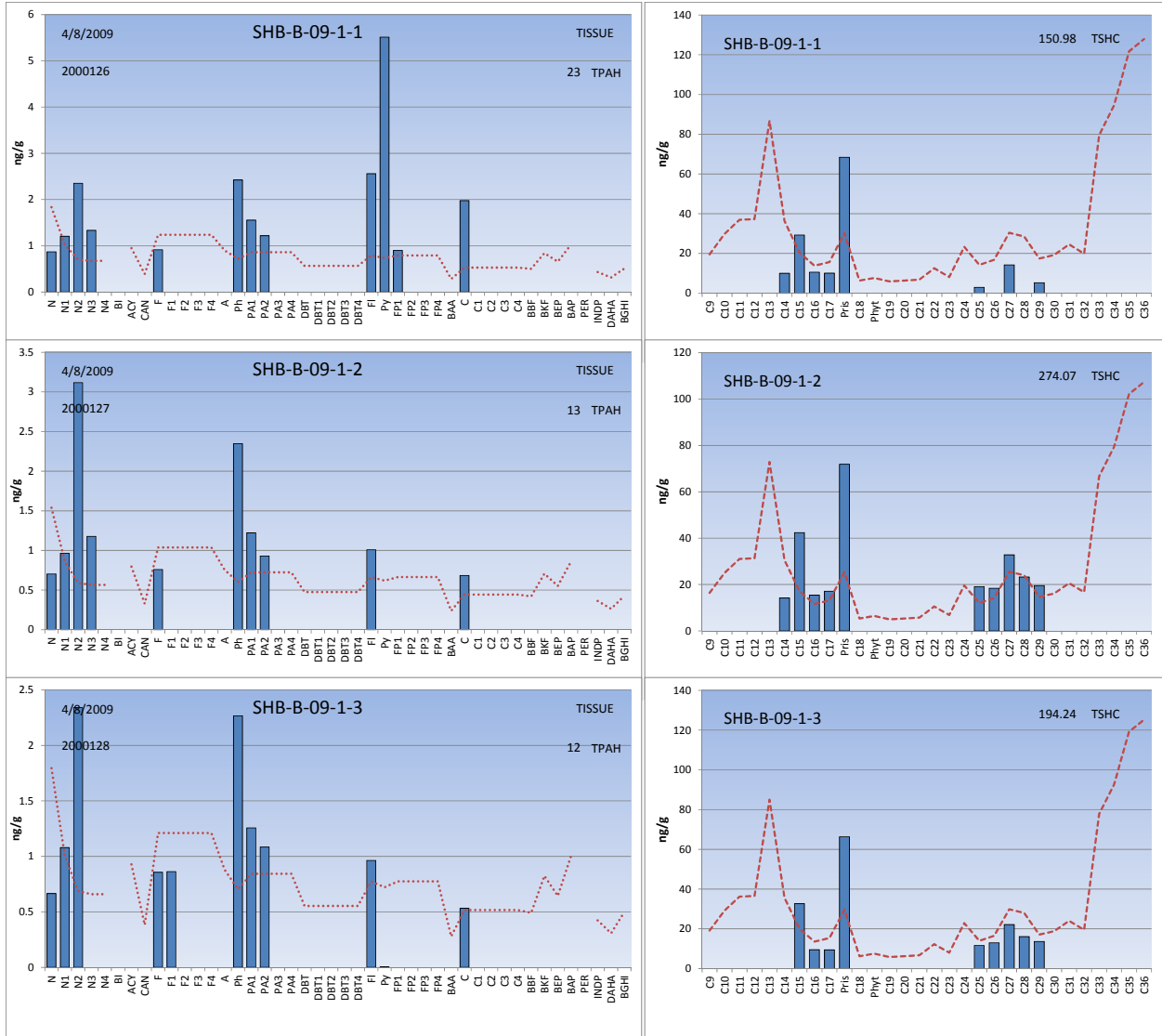


467 SHEEP BAY--SHB

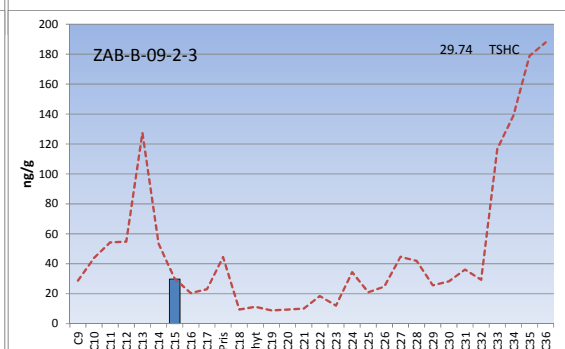
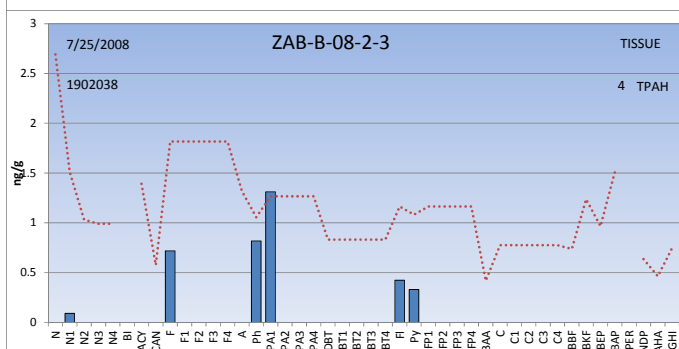
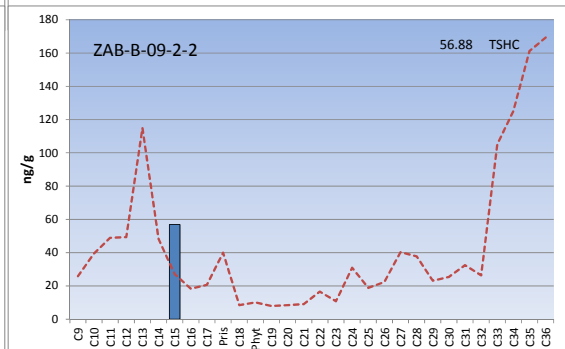
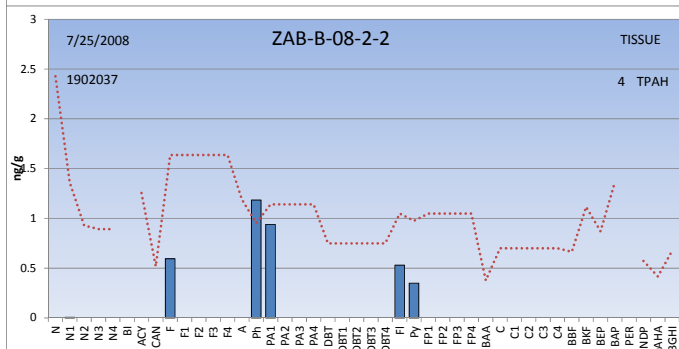
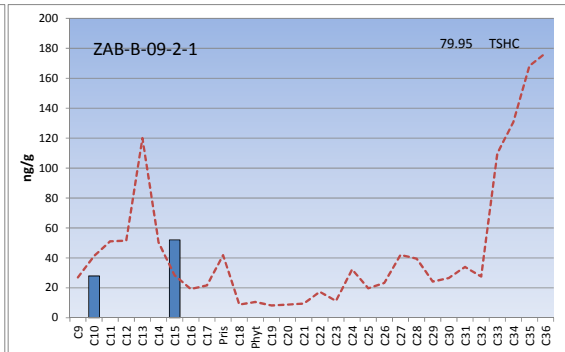
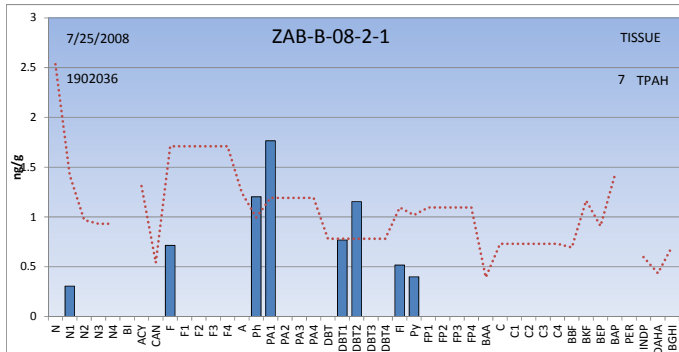


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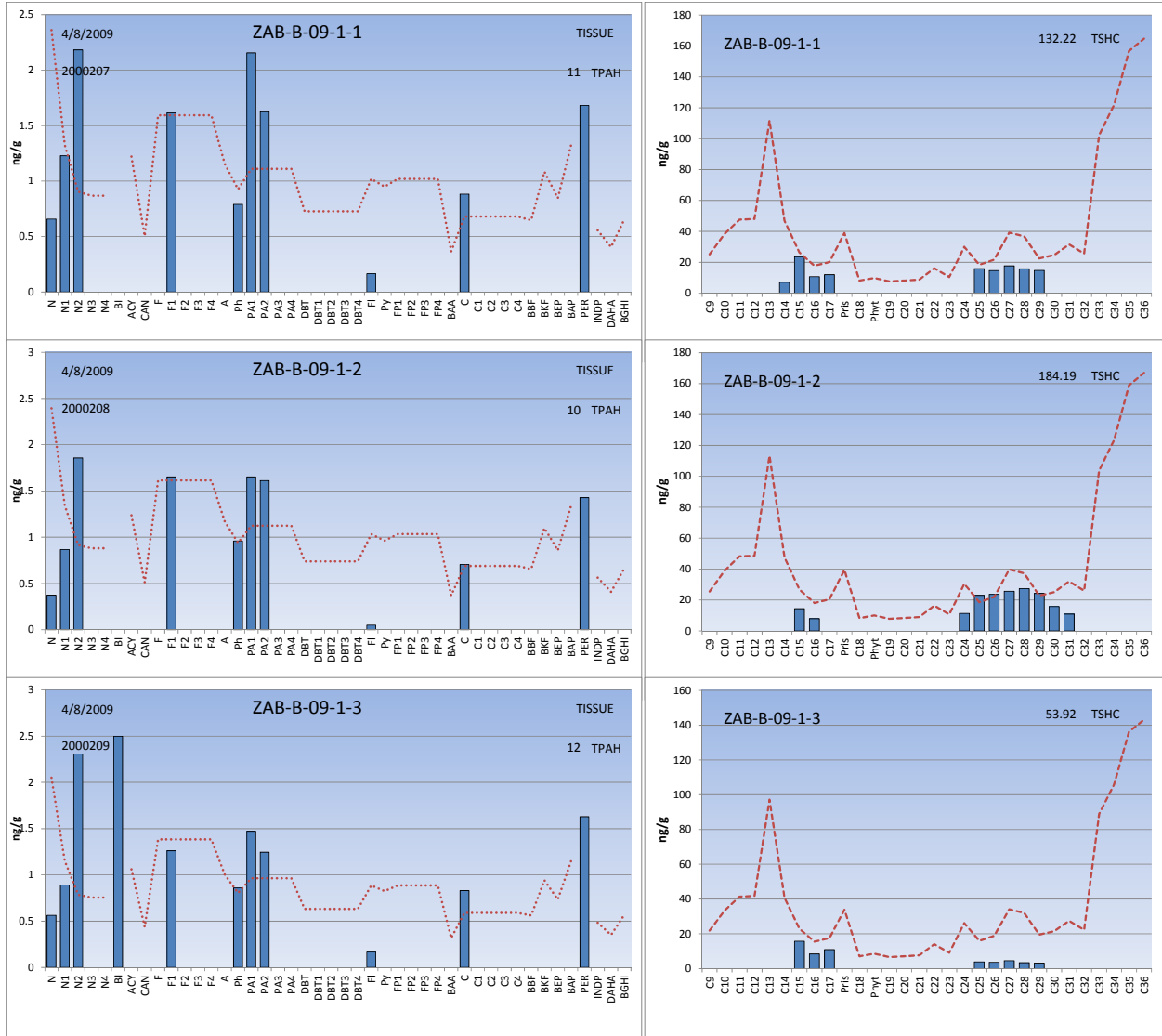


470 ZAIKOF BAY--ZAB



471

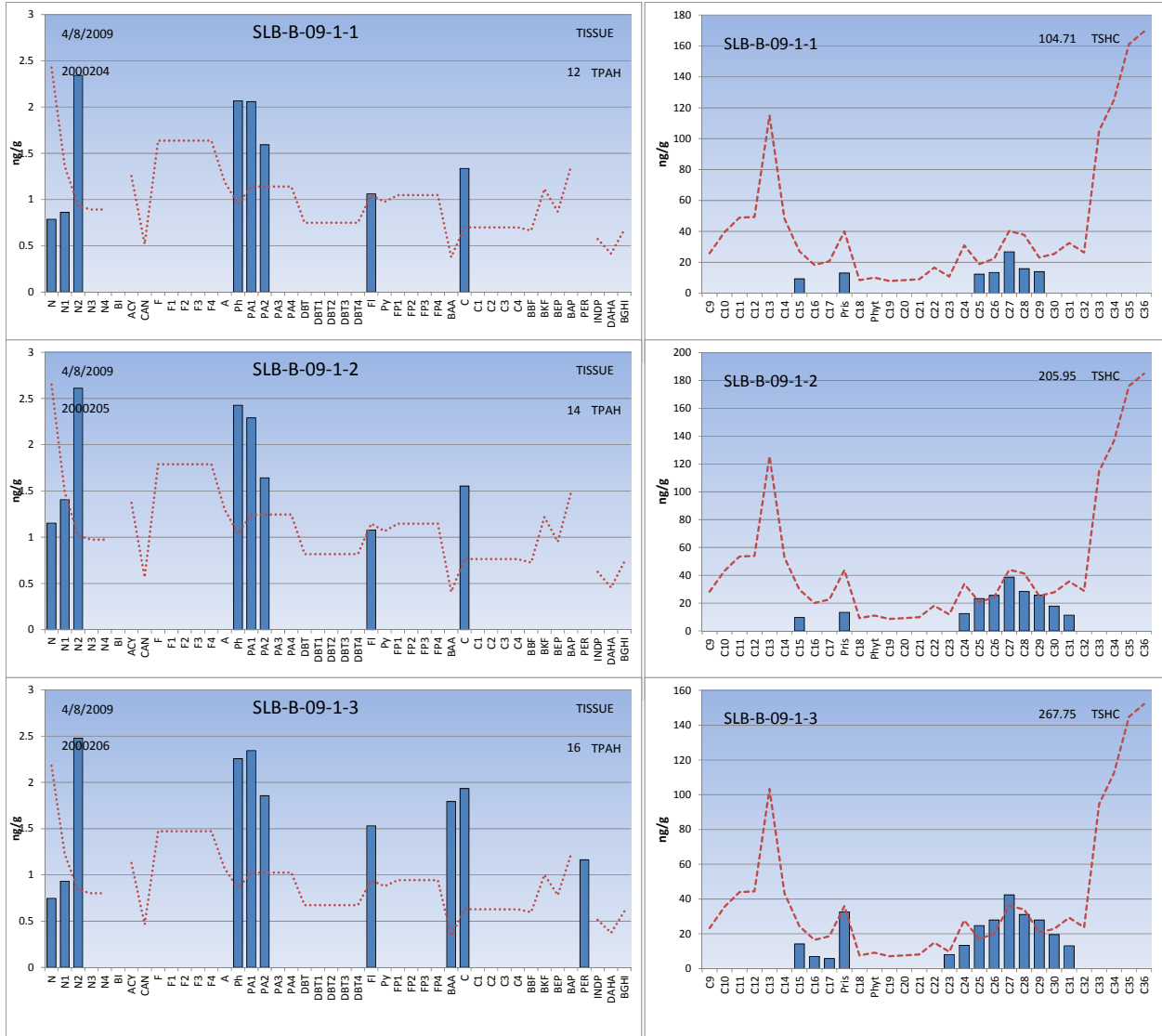
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472



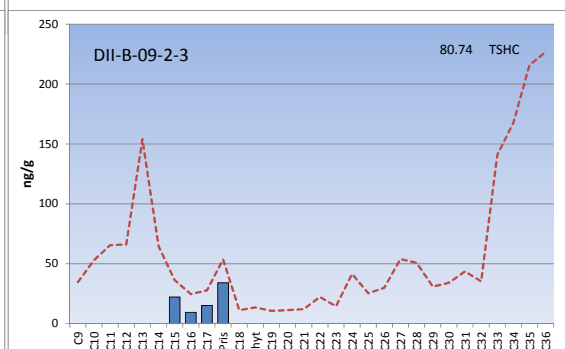
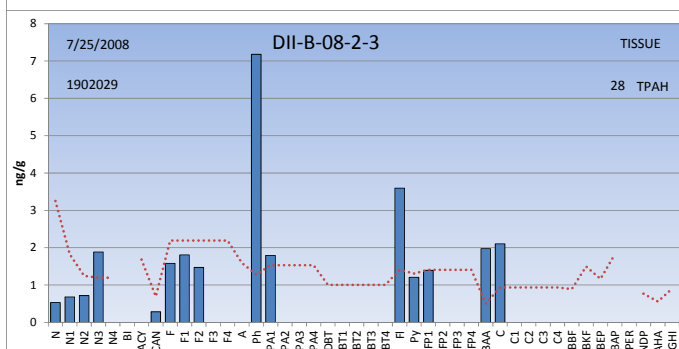
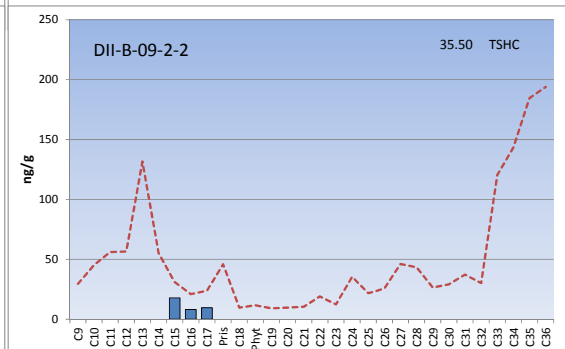
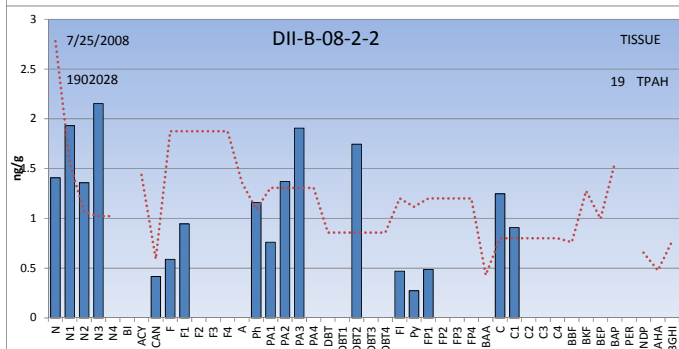
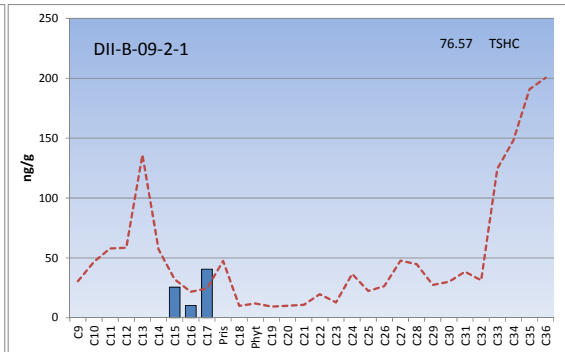
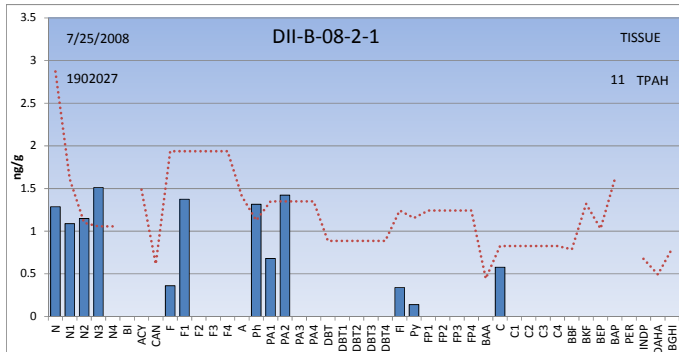
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475



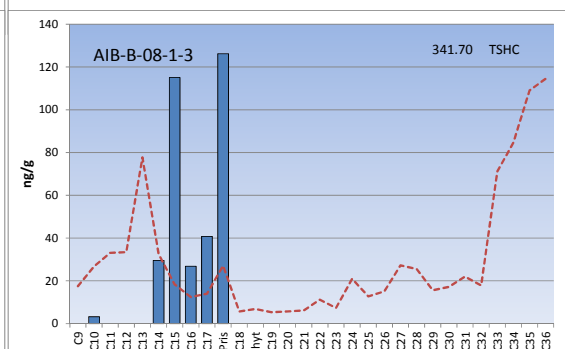
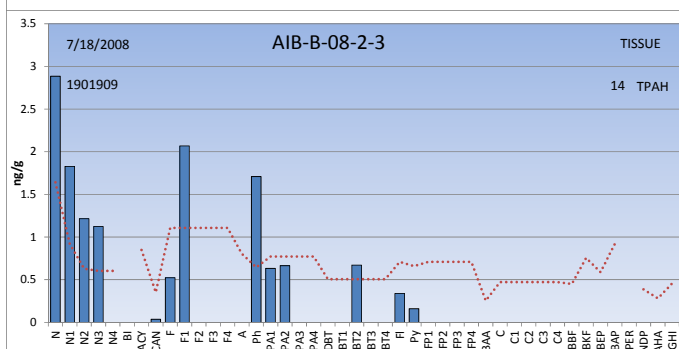
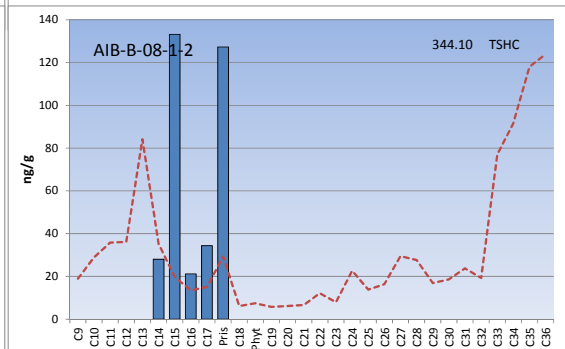
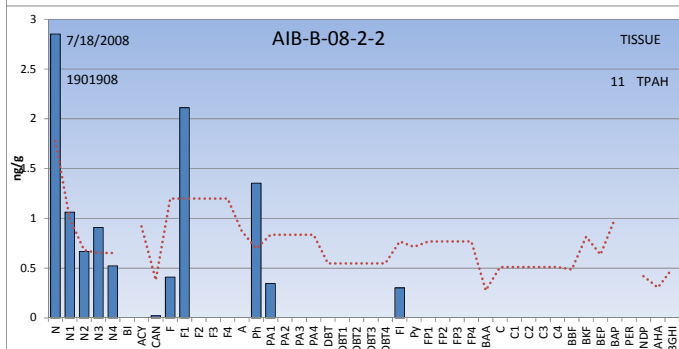
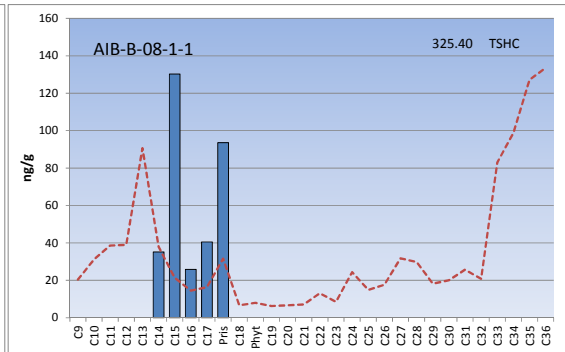
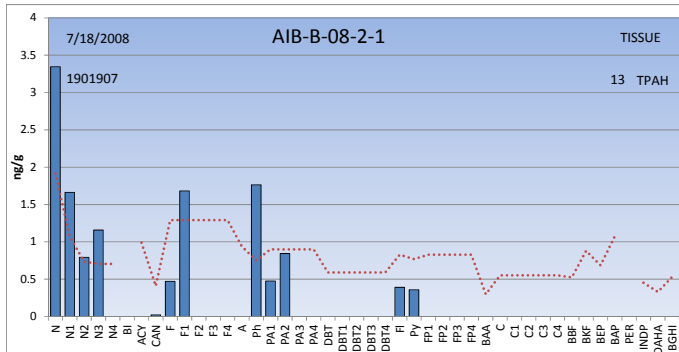
476 DISK ISLAND--DII



477

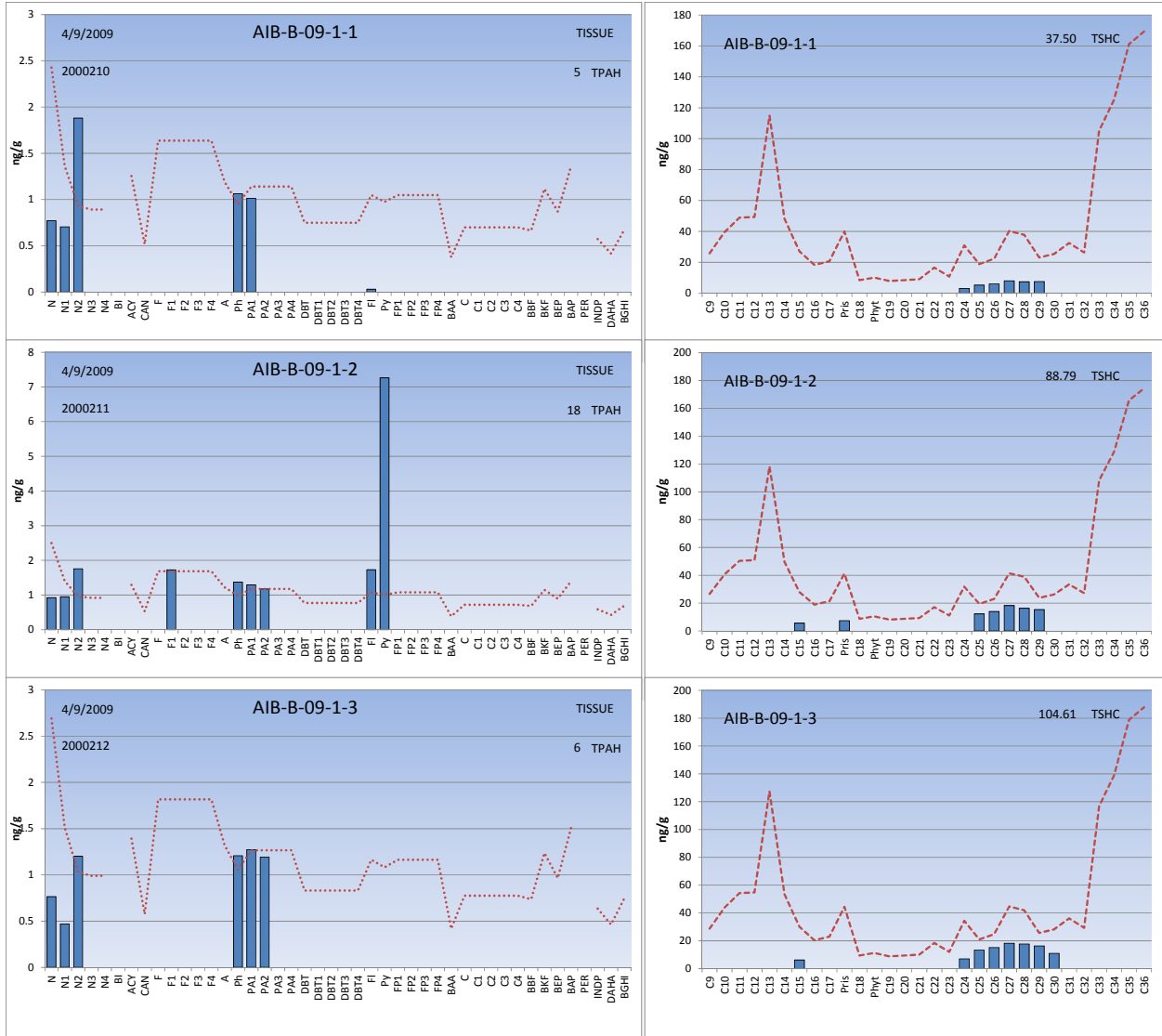


479 AIALIK BAY--AIB



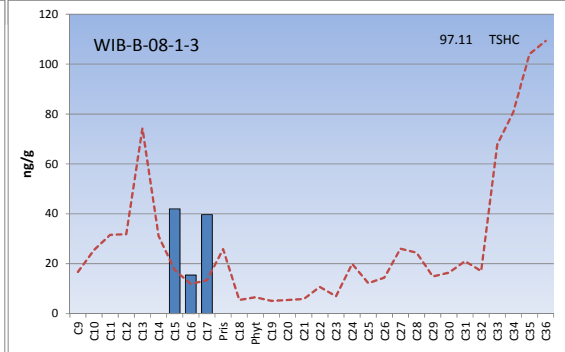
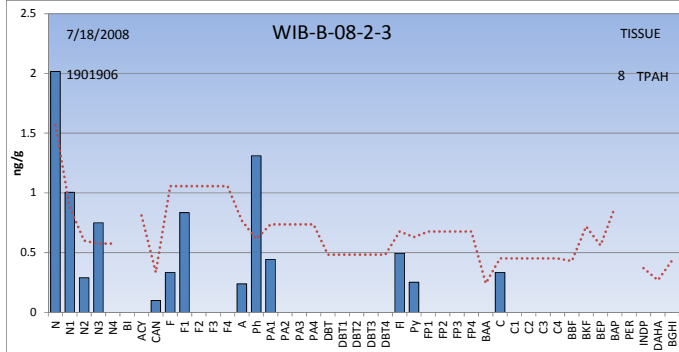
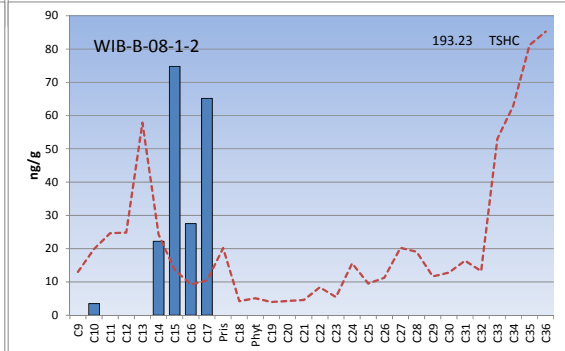
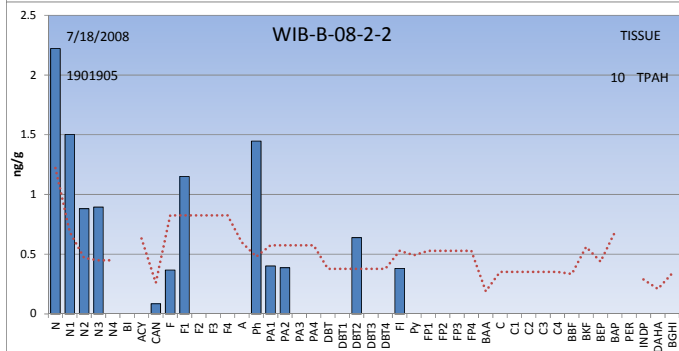
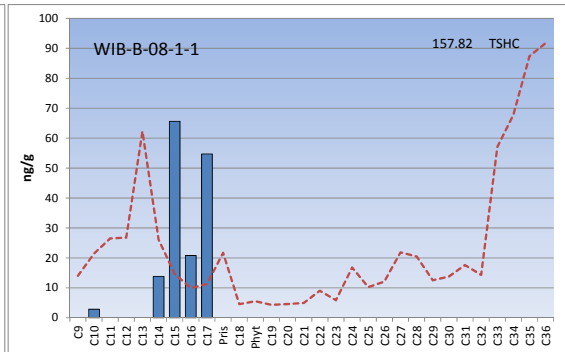
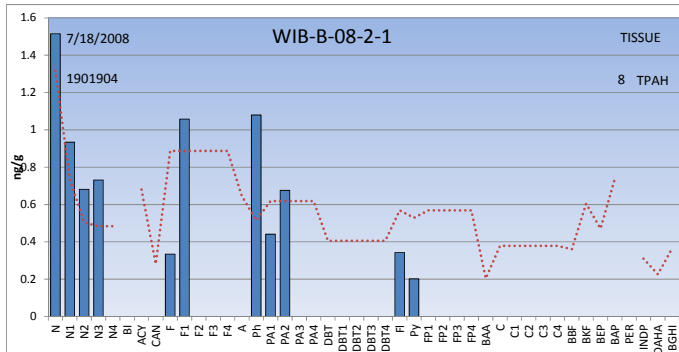
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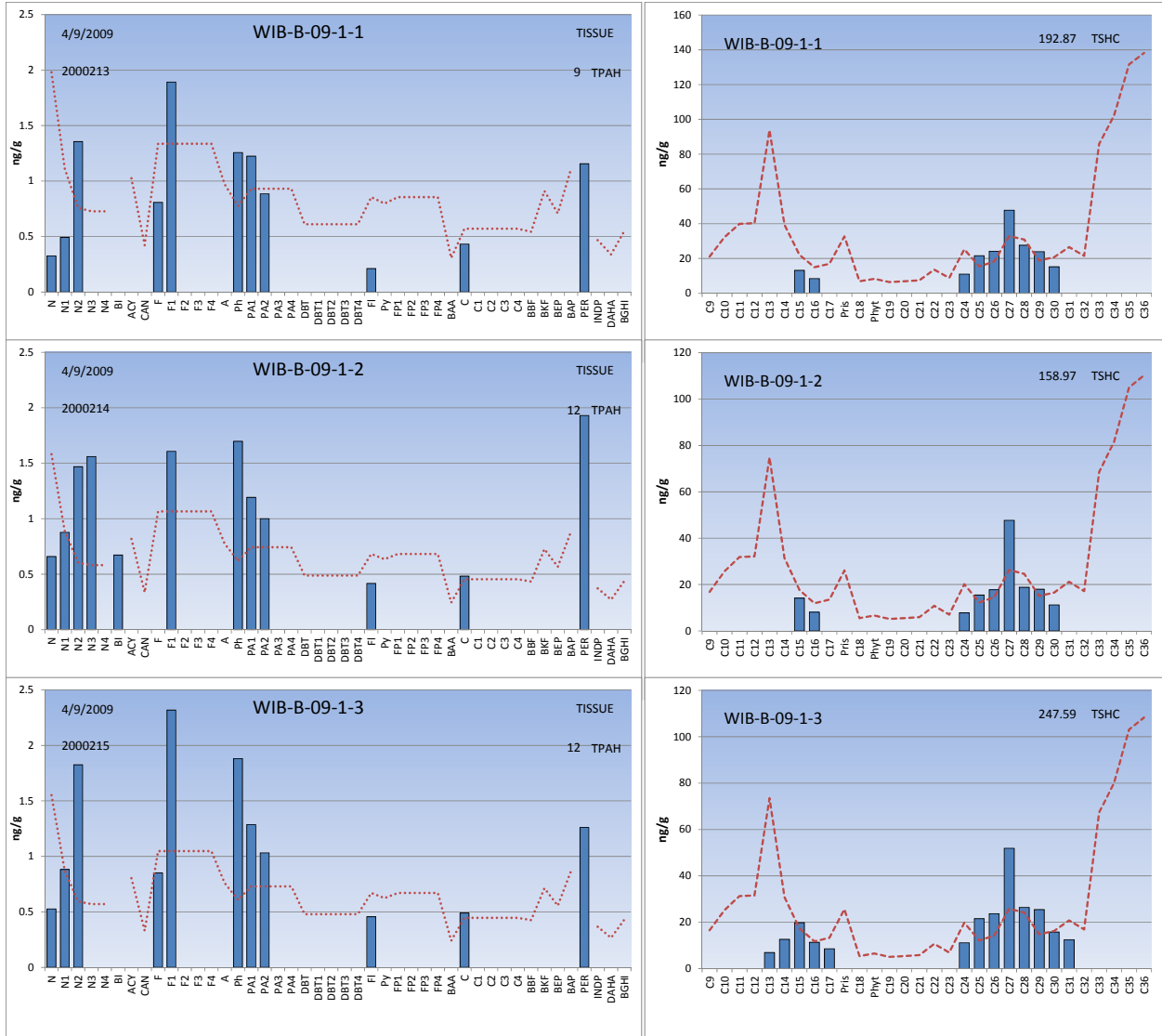
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482 WINDY BAY--WIB



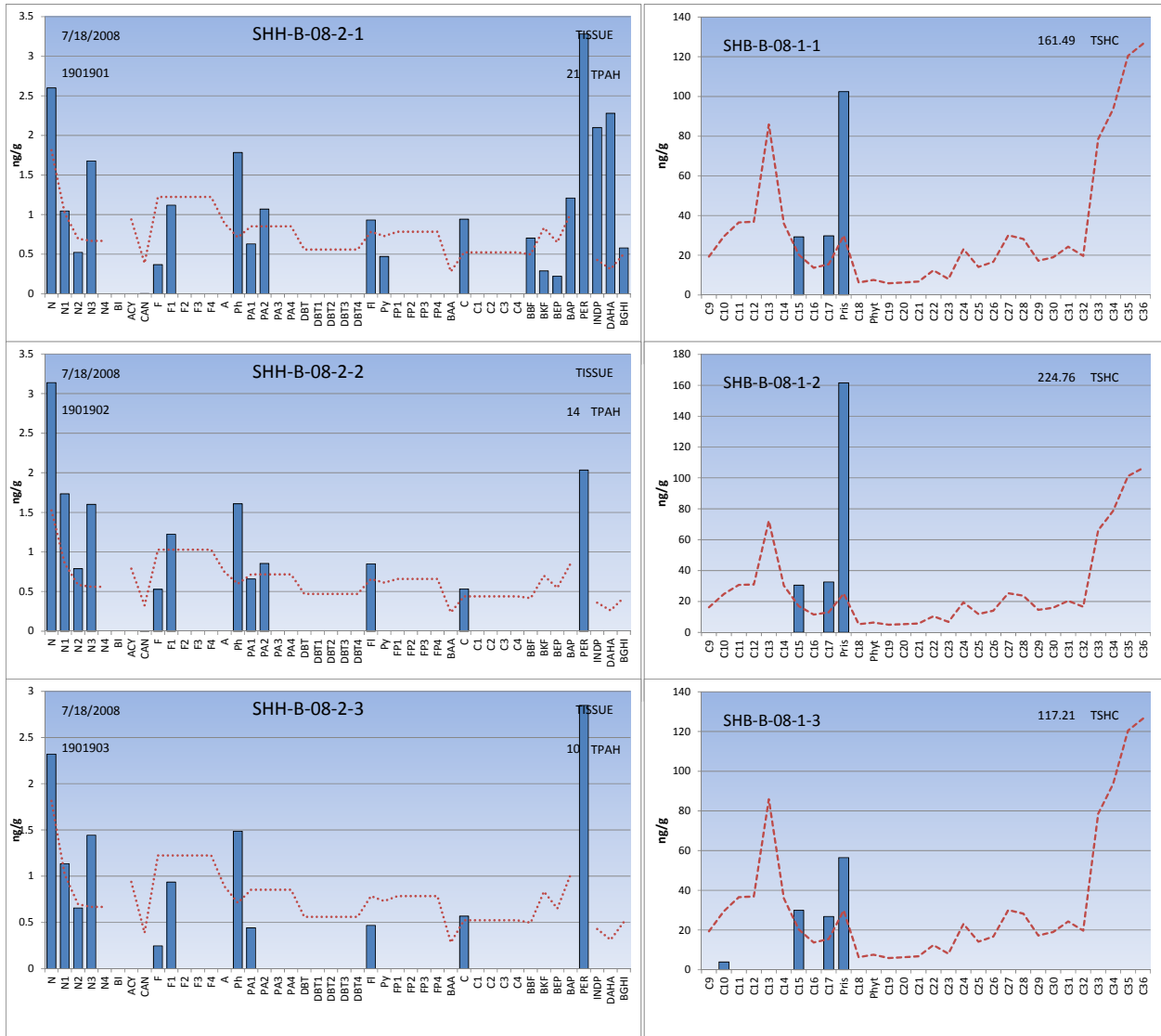
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# LTEMP Report 2008-2012



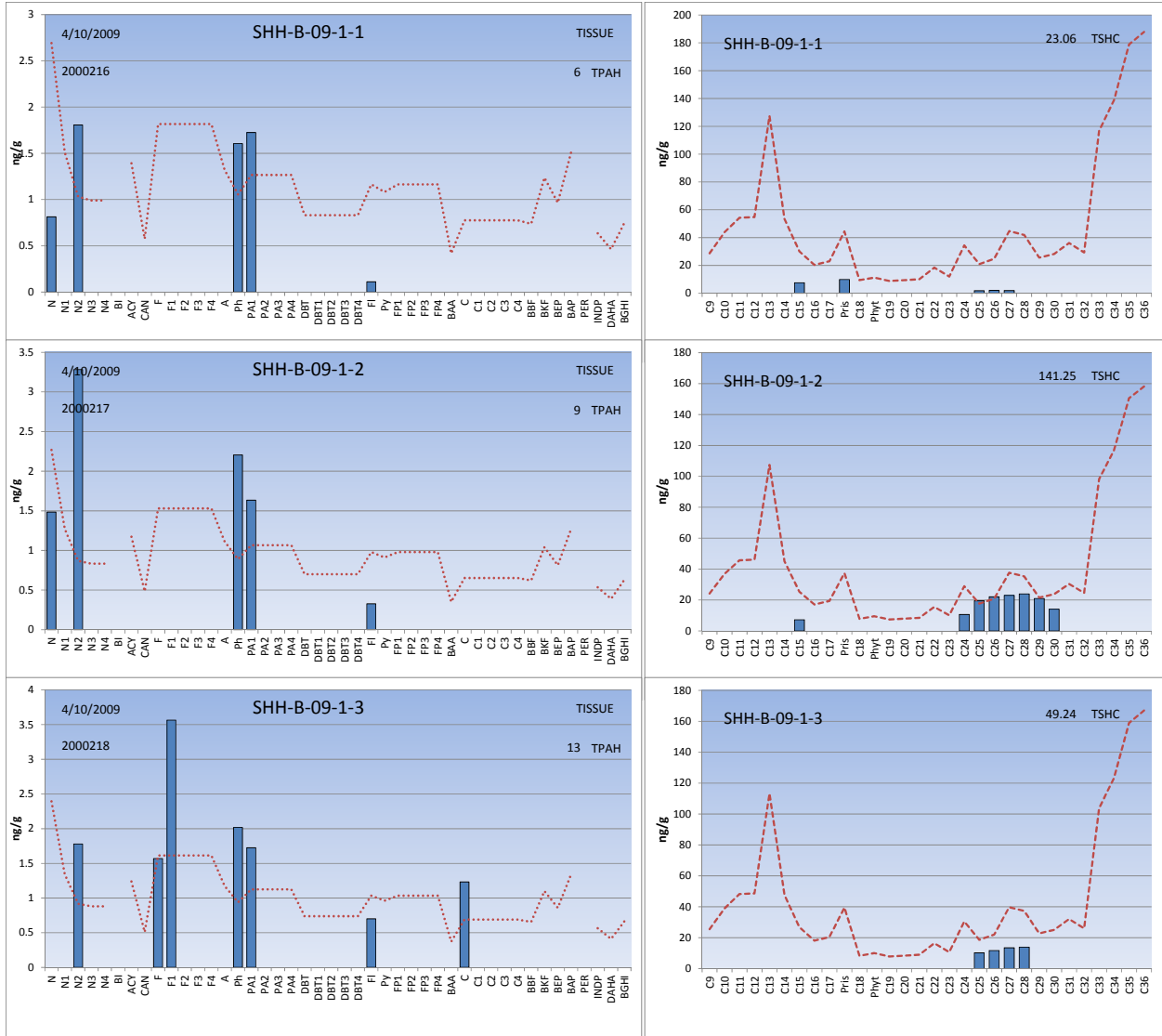
484

485 SHUYAK HARBOR--SHH



486

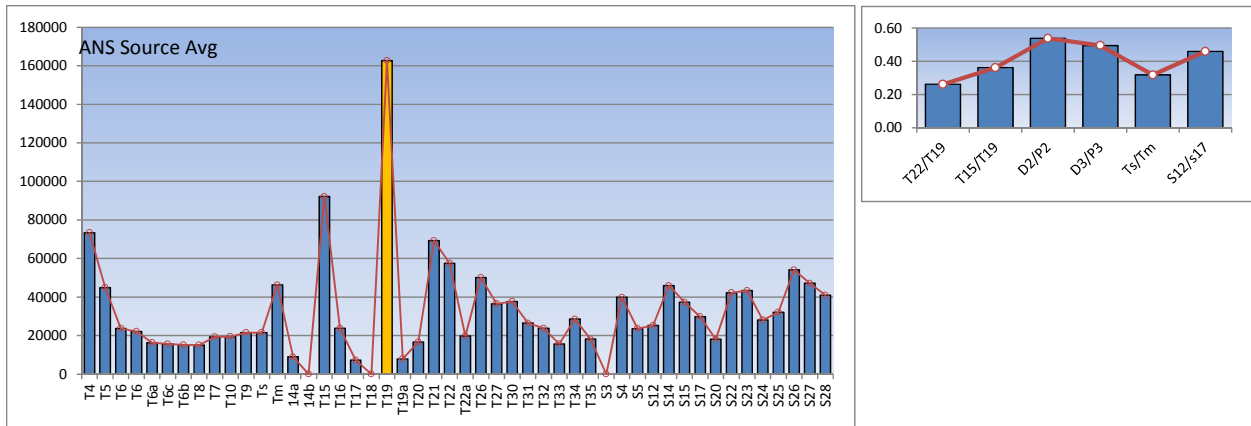
# LTEMP Report 2008-2012



487



488 APPENDIX 2 – BIOMARKER PLOTS



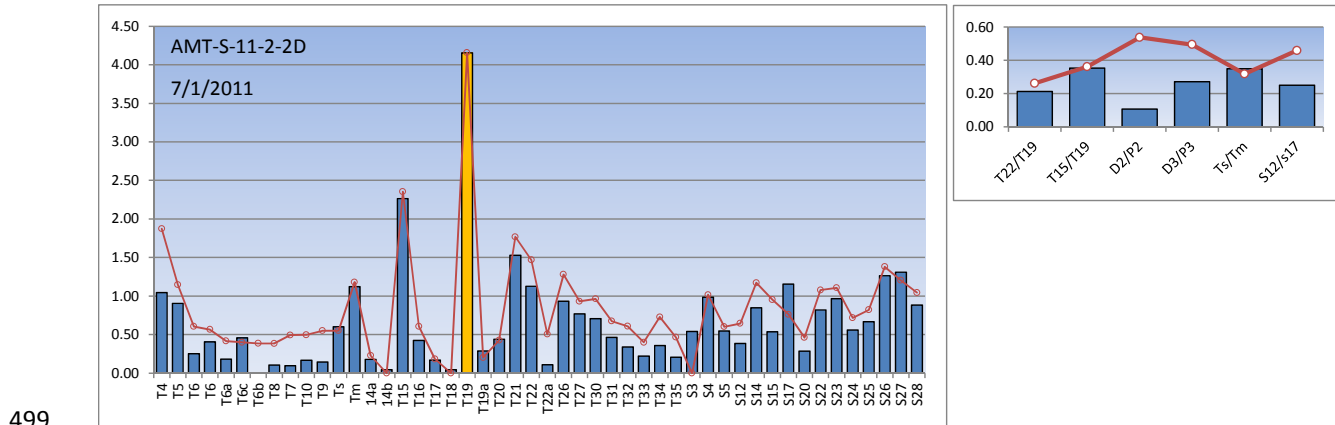
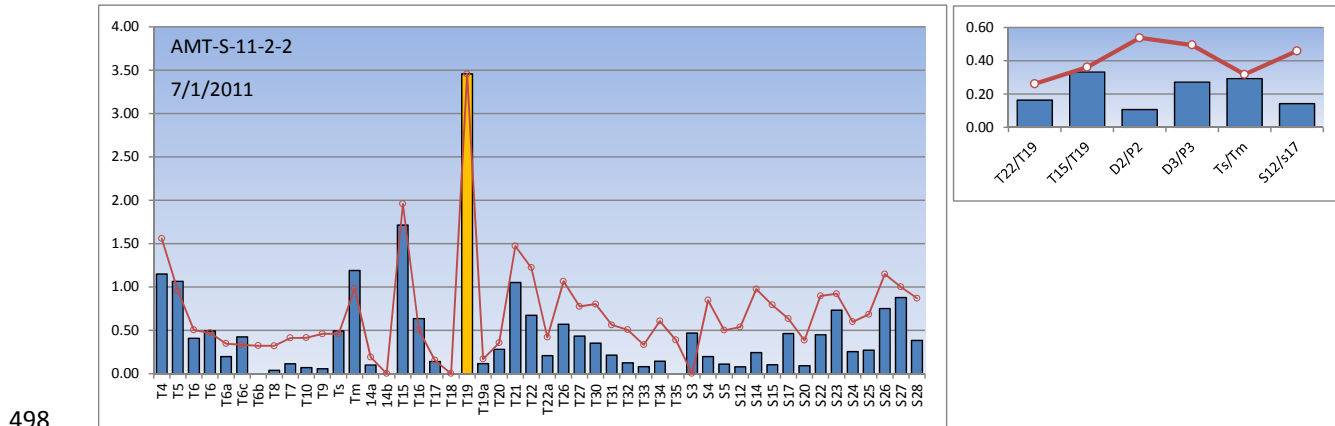
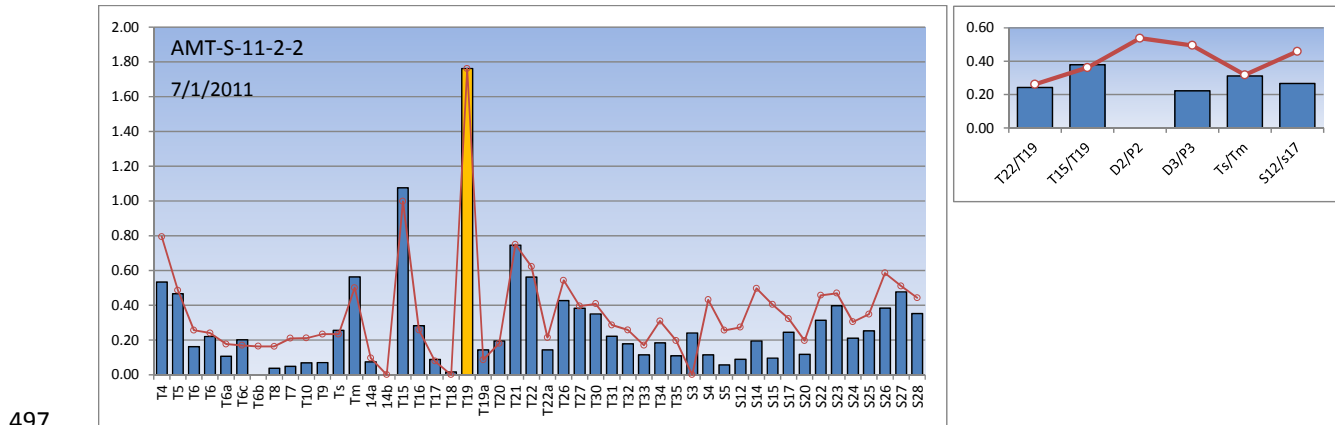
489

490 Quantified biomarkers for ABL’s EVOS-era ANS reference oil (left). Hopane highlighted with gold fill. Selected  
 491 diagnostic ratios (right). Red line in both plots represents ANS reference values, here matching the source sample  
 492 values.

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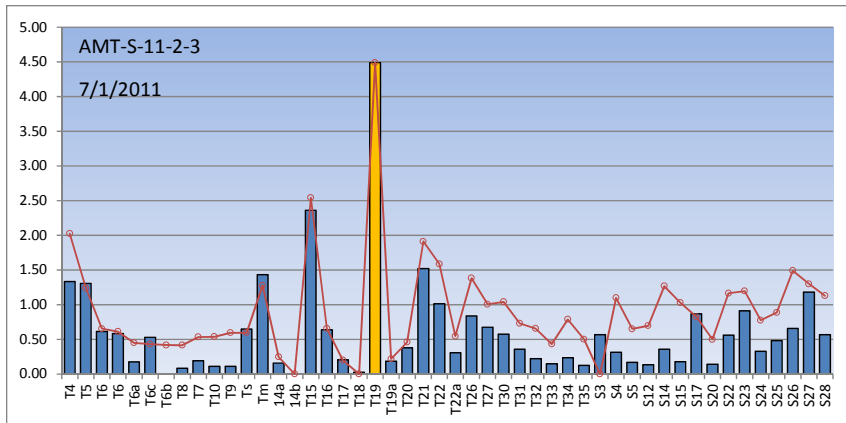
493 This section of the appendix contains the quantified biomarker plots and diagnostic ratios for all sediment samples  
 494 from AMT and GOC 2011-2012. In each plot, the red line denotes ANS reference oil. In the left plots, the ANS  
 495 profile is normalized to each sample's C30 hopane (T19).

## 496 ALYESKA MARINE TERMINAL--AMT JULY 2011



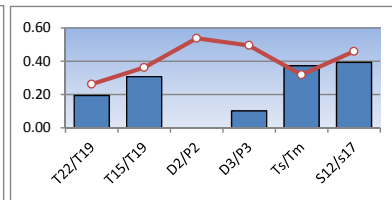
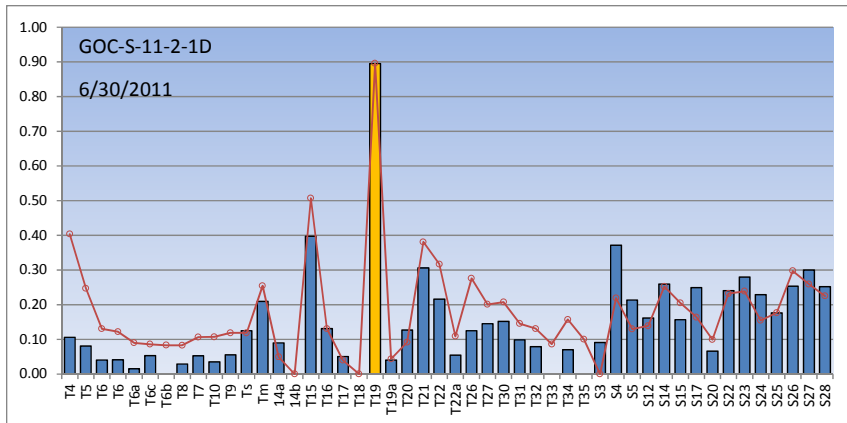
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500

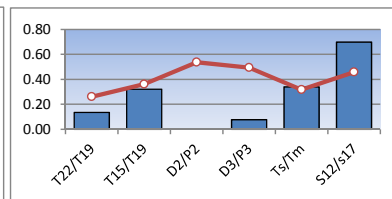
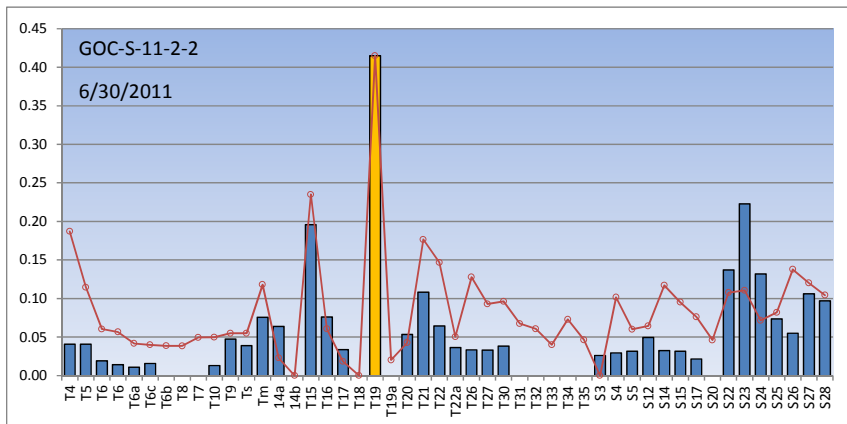


LTEMP Report 2008-2012

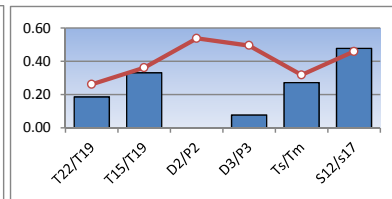
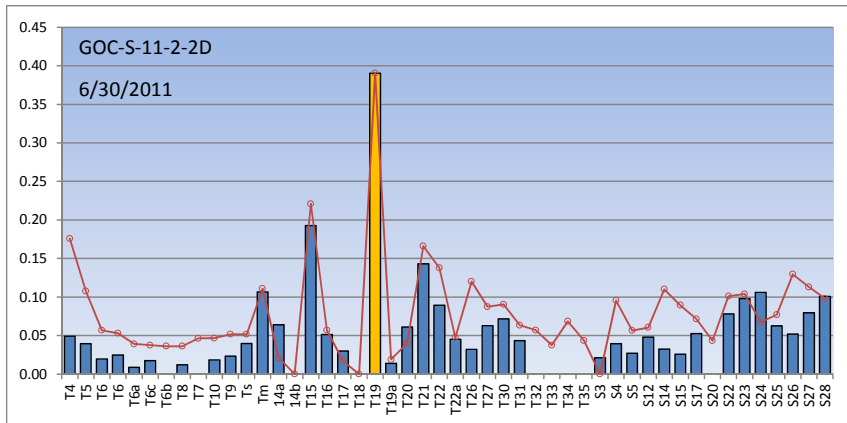
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505



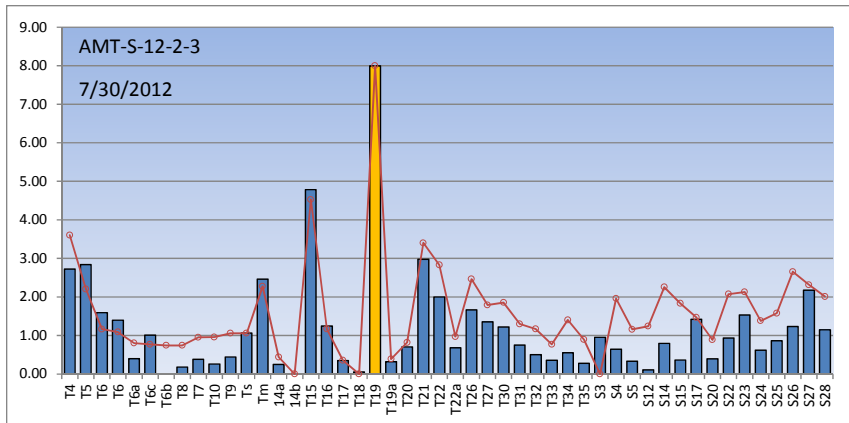
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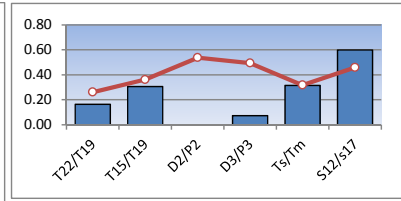
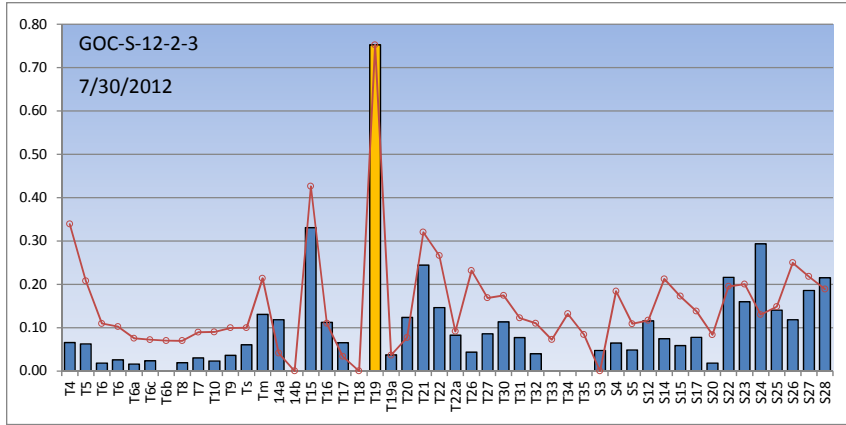


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512



# LTEMP Report 2008-2012



516

517 APPENDIX 3 – LTEMP PAH INTERPRETATION USING MULTIVARIATE SCORING METHODS

518

519 **Preliminary report of LTEMP data, 2008 – 2012**

520 **Mark Carls, Auke Bay Laboratory, NMFS**

521 **Summary**

522 Using nonparametric and multivariate methods described in Carls, 2006, oil is evident in Alyeska sediment based  
523 on PAH and alkanes. Hydrocarbons from the ballast treatment facility were not detected in tissues at any site.  
524 However, hydrocarbons from a probable diesel spill in 2008 were detected in tissue at Gold Creek (PAH only).  
525 There was also a hint of the diesel spill in sediment from Gold Creek – but only if it represents the dissolved  
526 fraction. Quality control samples were generally clean.

527

528 **Water (rinsate) samples**

---

529 **PAH**

530 Controls generally looked fine; TPAH range 0.3 to 21.4 ng.

531 Many PAHs in rinsate were below MDL

532 **Alkanes**

533 Calibrated alkanes were generally not detected in control samples

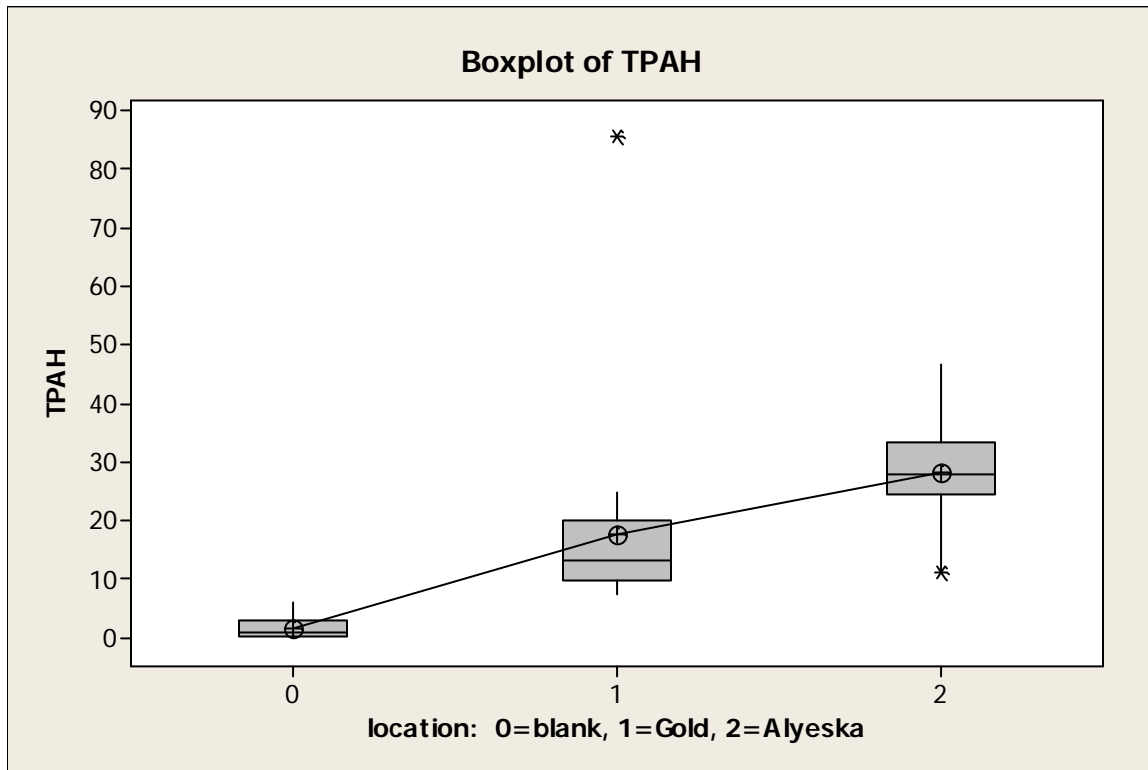
534 Alkane patterns are unusual in the last five samples collected (20111613, 20111620, 20120822-  
535 24); there was no discernible pattern, thus this is likely an artifact. Alkanes were generally not  
536 detected in these samples but C16, C18, and C25 alkanes were present.

537 Two samples (20100319-20) had distinct alkane patterns, C25-C32 and C22-C34 alkanes were  
538 present with no obvious odd-even preference



539 **Sediment: PAH**

540 TPAH concentration was least in blanks, intermediate at Gold Creek and highest at Alyeska ( $P_{ANOVA} =$   
 541 0.000). There is a high outlier at Gold Creek and a low outlier at Alyeska. Nearly all detected analytes  
 542 were above MDL in sediment.



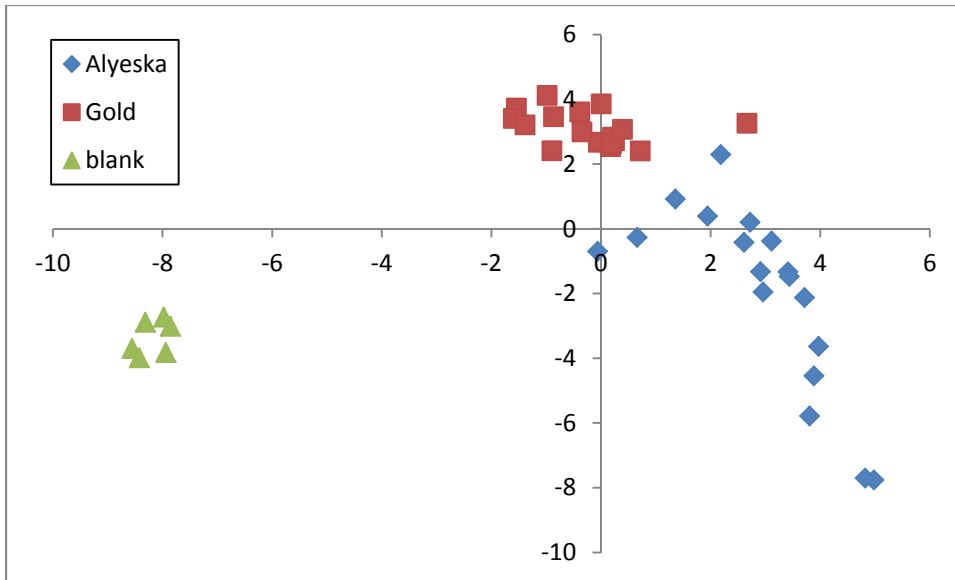
543  
 544 Alyeska Marine Terminal

545 Complex signal; weathering often estimable. The PAH source model finds oil sometimes – but  
 546 signal is mixed and balance may be shifting towards more pyrogenic.

547 Gold Creek

548 Signal is consistently pyrogenic. The outlier concentration may have been caused by a diesel spill  
 549 at Gold Creek about the time of sample collection (July 2008) but only if it represents the  
 550 dissolved fraction because composition is pyrogenic.

551 All three locations were distinguishable by normalized PAH composition with principal components  
 552 analysis. The blanks were completely separated (green triangles). The signal at Alyeska approached that  
 553 at Gold Creek in some samples and was different in others, forming a gradient. This gradient (at Alyeska)  
 554 is explained at least in part by weathering, a result consistent with weathering oil at other locations, such  
 555 as Unalaska Island in the Aleutians (C1 is correlated with percent chrysenes, for example; percent  
 556 chrysenes is a surrogate for weathering:  $r = 0.74$ ).



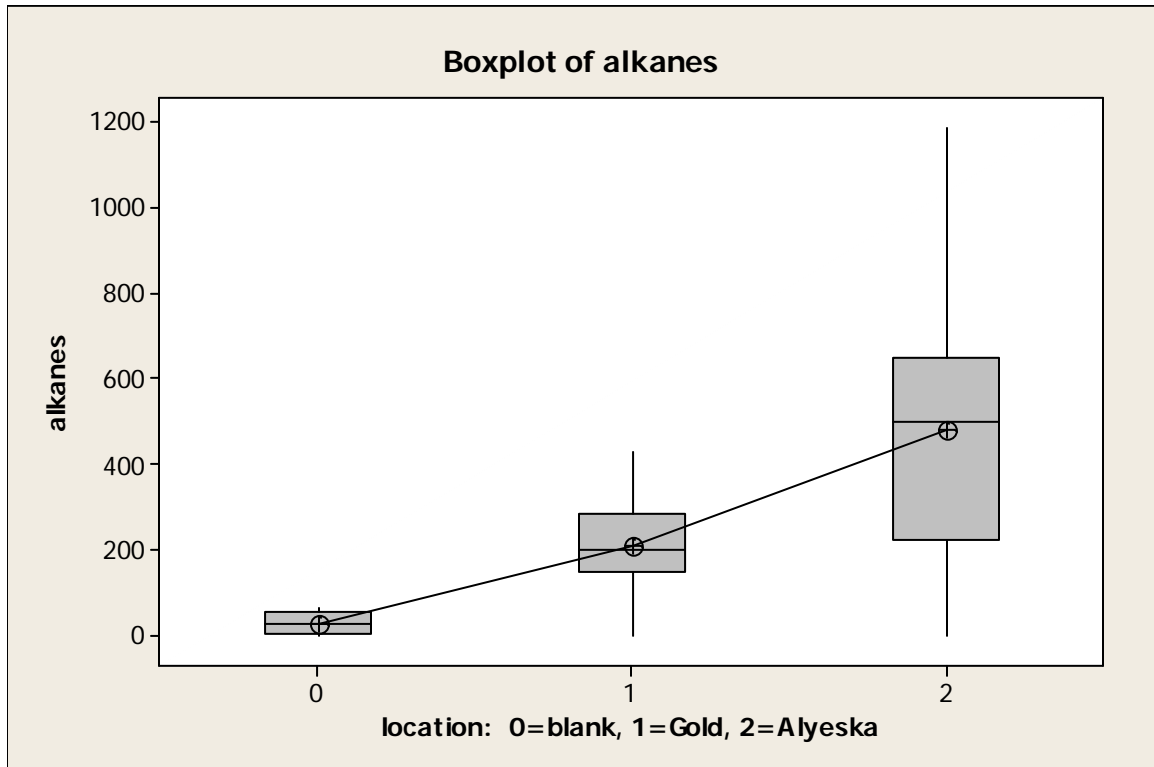
557

558 *PCA results; x-axis is the first component, y-axis is the second component; these explain 61% of the variance*

559 One sample from Gold Creek is separated from all other Gold Creek sediment samples in PCA space; it is  
 560 the sample with an unusually high TPAH concentration at this site and may be contaminated with fuel oil.  
 561 However, its PAH source score was -6.0, i.e., 100% pyrogenic. The only way this sample could represent  
 562 oil is if it were the water-washed fraction. This merits more discussion.

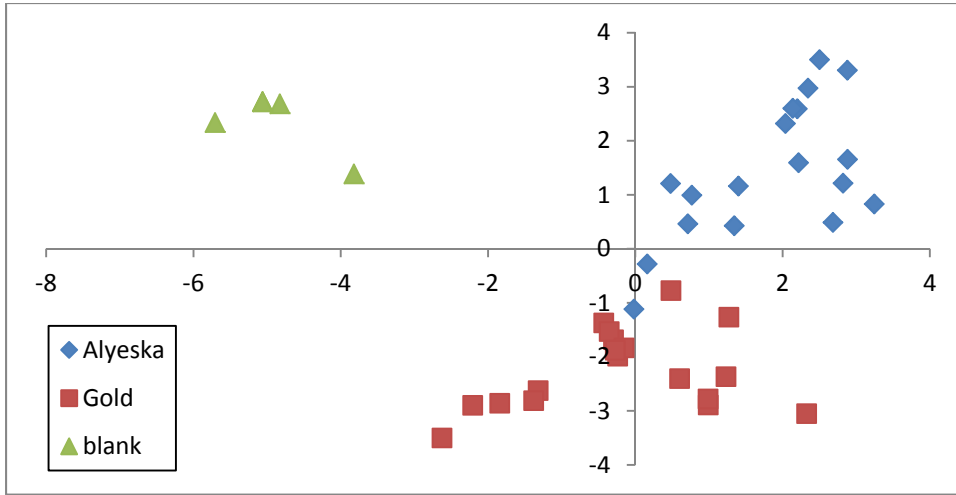
563 **Sediment, alkanes**

564 Alkane and calibrated alkane concentrations were least in blanks, intermediate at Gold Creek and highest  
 565 at Alyeska ( $P_{ANOVA} = 0.000$ ). This result is consistent with the PAH pattern. Variance is greatest at the site  
 566 most likely to be oiled (Alyeska); this is consistent with a range of samples, some clean, some  
 567 contaminated. Most detected alkanes in sediment were above MDL.



568  
 569 All three locations were distinguishable by normalized alkane composition with principal components  
 570 analysis, consistent with the PAH PCA. The blanks were completely separated (green triangles). The  
 571 signal at Alyeska approached that at Gold Creek in some samples and was different in others, forming a  
 572 gradient.

573



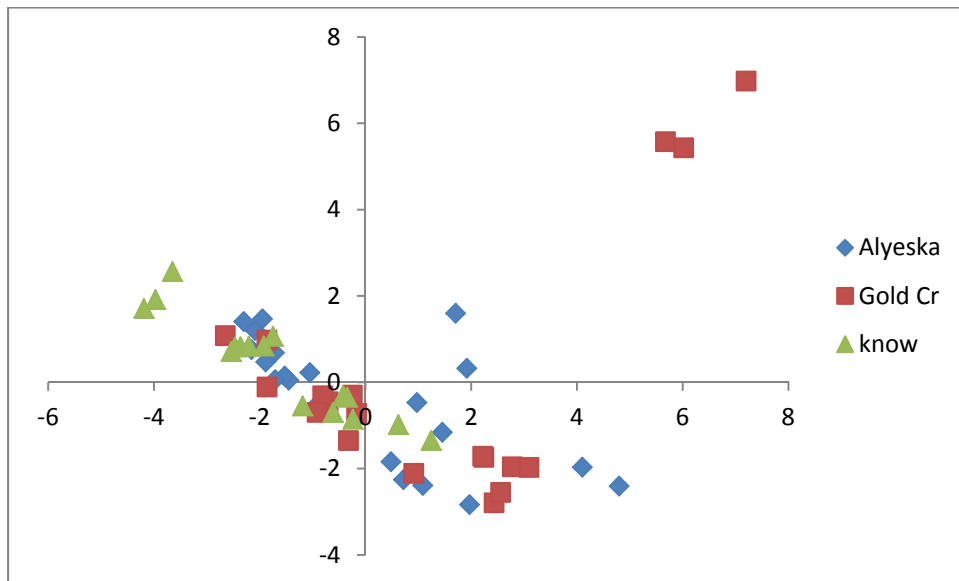
574

575 *PCA results; x-axis is the first component, y-axis is the second component; these explain 43% of the variance*

576

577 **Tissue, PAH**

578 PAH source modeling generally yielded mixed results (with both petrogenic and pyrogenic characteristics).  
 579 Composition of PAH in tissue at the three sites monitored through 2012 closely overlapped except for three  
 580 unusual Gold Creek samples and perhaps two Alyeska samples (see figure below). The general absence of  
 581 separation suggests that most mussels did not accumulate petroleum hydrocarbons, or at least not in detectable  
 582 quantities, but there were three notable exceptions. PAH concentrations in tissue were often below MDL.



583  
 584 *PCA results; x-axis is the first component, y-axis is the second component; these explain 38% of the variance*

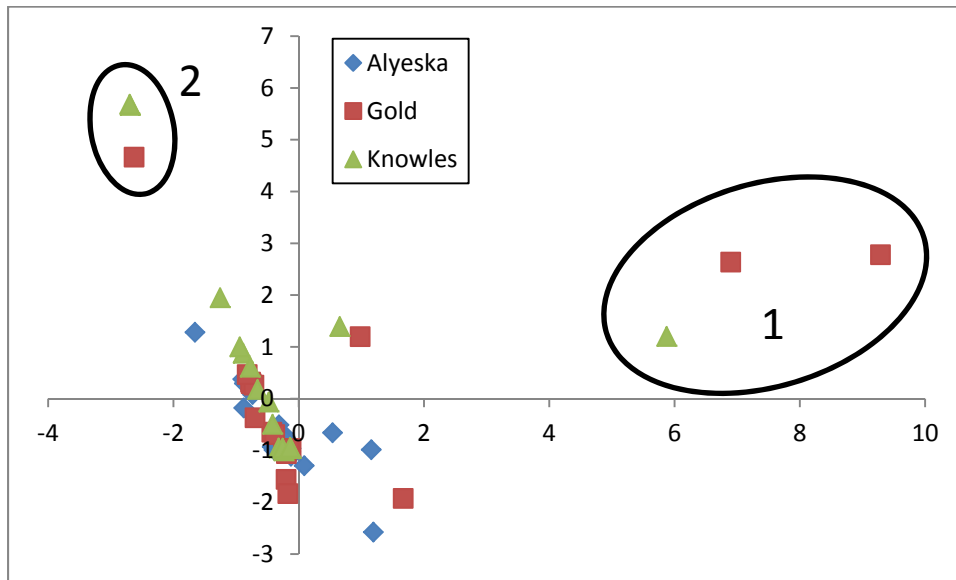
585 The three unusual Gold Creek tissues, collected in July 2008, had markedly different PAH source model scores (1.9)  
 586 than all other Gold Creek samples (-0.5 to 1.0). They had significant dibenzothiophene concentrations, unlike the  
 587 other Gold Creek tissues. Total PAH concentrations in the unusual samples were among the highest at this site.  
 588 This is consistent with a diesel oil spill reported about this time (Payne et al. 2006; 2008).

589 The three unusual Gold Creek samples suggest where oiled tissue might fall in PCA space. Only two other tissue  
 590 samples, both from Alyeska, had a proclivity in this direction. Their PAH source model scores, TPAH  
 591 concentrations, and proportions of naphthalenes, fluorenes, dibenzothiophenes, phenanthrenes,  
 592 fluoranthene/pyrenes, and chrysenes were within the range of other scores at Alyeska hence there were no  
 593 obvious reasons to identify them as oiled.

594 Bottom line; most tissues do not have evidence of petroleum hydrocarbons. The exceptions were associated with  
 595 a diesel spill.

596 **Tissue, alkanes**

597 Composition of alkanes in tissue at the three sites monitored through 2012 closely overlapped except for two  
 598 unusual clusters (circled groupings in figure below). Most detected alkane concentrations in tissue were above  
 599 MDL.



600  
 601 *PCA results; x-axis is the first component, y-axis is the second component; these explain 37% of the variance*

602 The exceptional alkane tissues in cluster 1 (lower right-hand quadrant) did not correspond to exceptional PAH  
 603 composition. The two unusual Gold Creek tissues and single Knowles head tissue may be due to unusual  
 604 approximately C<sub>23</sub>-C<sub>29</sub> alkanes generally not present in tissue at these sites. However, these mid-range alkanes  
 605 were present in some tissues collected elsewhere and the same pattern occurred in a couple rinsate samples.

606 Exceptional tissue samples in cluster 2 (upper left-hand quadrant) are not consistent with oil. It contains a Gold  
 607 Creek sample collected in July 2008 (1902002; post-diesel spill) but does not have an alkane pattern consistent  
 608 with oil (only C<sub>14</sub>-C<sub>16</sub> biogenic alkanes were detected). Likewise, alkane composition in the Knowles Head tissue in  
 609 cluster 2 has few detected alkanes and does not look like oil.

610 There is no obvious evidence of odd-even preference indicative of plant waxes.

611 We conclude that alkane composition does not provide evidence for oil contamination in tissue.